

**HYDROGEOLOGY AND WATER MANAGEMENT OF THE CLEAR  
LAKE AQUIFER, WITH EMPHASIS ON THE SOUTH MEDICINE  
LAKE MANAGEMENT AREA**



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Ground Water Investigation Program**

*Cover: Sun setting over Medicine Lake. Photo by Kevin Chandler, MBMG.*

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## PREFACE

The Ground Water Investigation Program (GWIP) at the Montana Bureau of Mines and Geology (MBMG) investigates areas prioritized by the Ground Water Assessment Steering Committee (2-15-1523 MCA) based on current and anticipated growth of industry, housing, commercial activity, or agriculture. Additional program information and project ranking details are available at <http://www.mbm.mtech.edu/WaterEnvironment/GWIP/main.asp>. GWIP also collects and compiles groundwater and surface-water data for each study area. These data are used to interpret how groundwater has responded to past stresses and to project future responses.

The final products of the Clear Lake aquifer study include:

**An Interpretive Report** (this report) that presents data interpretations and summarizes the project results. This report's main focus is potential effects to surface water and groundwater from increased use of the Clear Lake aquifer, and whether water development or other land uses have impacted water quality in the Clear Lake aquifer.

**A Groundwater Modeling Report** (Chandler and Reiten, 2019) documents development of groundwater flow models, including a detailed description of the procedures, assumptions, and results of the models. Groundwater modelers and other qualified individuals can evaluate and use the models as a starting point to test additional water-use scenarios and for site-specific analyses. The MBMG website includes the model files (<http://www.mbm.mtech.edu/gwip/gwip.html>).

MBMG's Groundwater Information Center (GWIC) online database (<http://mbm.gwic.mtech.edu/>) provides a permanent archive for the data from this study.

## ABSTRACT

The Clear Lake aquifer of northeast Montana is contained in a buried valley underlying broad topographic swales associated with the ancestral Missouri River and glacial meltwater. The aquifer consists of the alluvial gravel of the ancestral Missouri and overlying glacial outwash gravel. These deposits form distinctive zones in the aquifer with varying connectivity.

Local precipitation and snowmelt recharges the aquifer, which replenishes lakes and wetlands. Evapotranspiration removes from 3 to 5 ft of water annually from the lakes and wetlands, creating gradients in the groundwater flow system.

This project focused on the South Medicine Lake management area, where the alluvial gravel of the Clear Lake aquifer is not continuous with the lower alluvial zone of the aquifer east of Medicine Lake. A groundwater divide underlying an area of sand hills southeast of Medicine Lake separates the flow system. Precipitation falling on the sand hills recharges the aquifer. As a result, good-quality irrigation water in the alluvial aquifer is restricted to an area within about 5 to 6 mi of the sand hills. Water level, water quality, and aquifer test results all indicate a hydraulic barrier to groundwater flow between this segment of the Clear Lake aquifer and the surface water of Medicine Lake.

Extensive monitoring and pumping volume estimates from irrigation wells provided the background for conceptual and numeric model development. Steady-state and transient finite difference numerical models simulate conditions in the aquifer. The transient model was calibrated to water-level fluctuations resulting from pumping at four existing irrigation wells in 2015. Predictive simulations evaluate changes in the flow system from increased pumping, including simulations doubling the number of wells and increasing pumping withdrawals by a factor of five. A simulation with twice the number of irrigation wells and pumping rates similar to 2015 indicated no change in water levels near Big Muddy Creek. The simulation with eight wells pumping nearly continuously indicated groundwater-level declines near Big Muddy Creek. Predictive model simulations can be used to assess additional pumping from proposed wells as new permits are requested.

Water quality in the upper outwash gravel zone is generally acceptable for irrigation, while much of the alluvial gravel zone is not. High concentrations of sodium and elevated dissolved minerals in the Fort Union aquifer locally degrade water quality in the Clear Lake aquifer, predominately in the lower alluvial zone.

## INTRODUCTION

### Background

The Clear Lake aquifer is an important water resource in northeastern Montana. The aquifer contains (1) sand and gravel (Wiota Gravel) deposited as alluvium by the ancestral Missouri River and (2) glacial outwash sand and gravel deposited by meltwater streams. The aquifer, in eastern Sheridan and Roosevelt Counties, Montana, extends from Westby along the state line southwest to Big Muddy Creek (fig. 1). These deposits form a complex aquifer system, with some areas capable of supporting high-yield irrigation wells. Groundwater flow is to the southwest from the North Dakota border to the Big Muddy Creek Valley. In some locations, the aquifer units are hydraulically connected, but elsewhere, confining beds separate them. Precipitation recharges the aquifer, but estimates of the amount and distribution of recharge vary widely. The aquifer discharges to wells, lakes, wetlands, and the alluvium and outwash underlying the Big Muddy Creek Valley. The volume of water that discharges from the aquifer is not well defined. For example, groundwater reaches surface sloughs and lakes where it evaporates, and phreatophytes, plants with roots tapping the water table, consume groundwater through transpiration.

Area lakes and wetlands are important habitat for migratory birds and other wildlife. The U.S. Fish and Wildlife Service (USFWS) manages many of these lakes and wetlands as part of the Medicine Lake Wildlife Refuge (fig. 2). The USFWS is concerned that irrigation withdrawals will deplete water from the wetlands, thereby adversely affecting the habitat (S. Lofgren, oral commun., 2017). The USFWS is also concerned with impacts of hydrocarbon development to groundwater and surface water in these areas.

The Sheridan County Conservation District (SCCD) has a reserved groundwater right for irrigation from the Clear Lake aquifer under the Sheridan County Water Reservation (SCWR). The Montana Bureau of Mines and Geology (MBMG) has monitored

and evaluated the effects of irrigation development in the Clear Lake aquifer since the 1980s, providing hydrogeologic data that support aquifer management. Currently, the combined irrigation authorizations from the pre-reservation State permits and the SCWR permits total about 13,737 acre-ft of water per year. Developing additional irrigation wells and appropriating more groundwater could potentially affect existing water-rights holders and could affect water levels in area wetlands.

### Purpose and Scope

The project purpose is to provide data, analyses, and interpretations to support management of the Clear Lake aquifer. To meet this purpose, the MBMG addressed two objectives:

1. Determine how much groundwater withdrawal from the Clear Lake aquifer is possible without detrimental effects to surface water or excessive water-level declines in wells.
2. Determine if groundwater or energy development have affected water quality in the Clear Lake aquifer.

Major tasks to accomplish these objectives include:

- i. Evaluate data from previous geologic and hydrogeologic reports along with other data.*

Data collection by the MBMG and SCCD since the 1980s have included exploratory test drilling, monitoring well construction, aquifer testing, water-level monitoring, and water-quality sampling. These data combined with data collected during this project (2013–2015) provided information to construct a conceptual hydrogeologic model of the Clear Lake aquifer and support calibration of the South Medicine Lake numerical model.

- ii. Develop a conceptual model for the South Medicine Lake focus area.*

This project focused on the South Medicine Lake area because of concerns that irrigation withdrawals were affecting water levels in Medicine Lake, decreasing flows in Big Muddy Creek, and creating detrimental well interference. The conceptual model includes information on the geologic framework and groundwater flow system for the Clear Lake aquifer. The conceptual model provided the basis for construction of the numerical groundwater flow model.

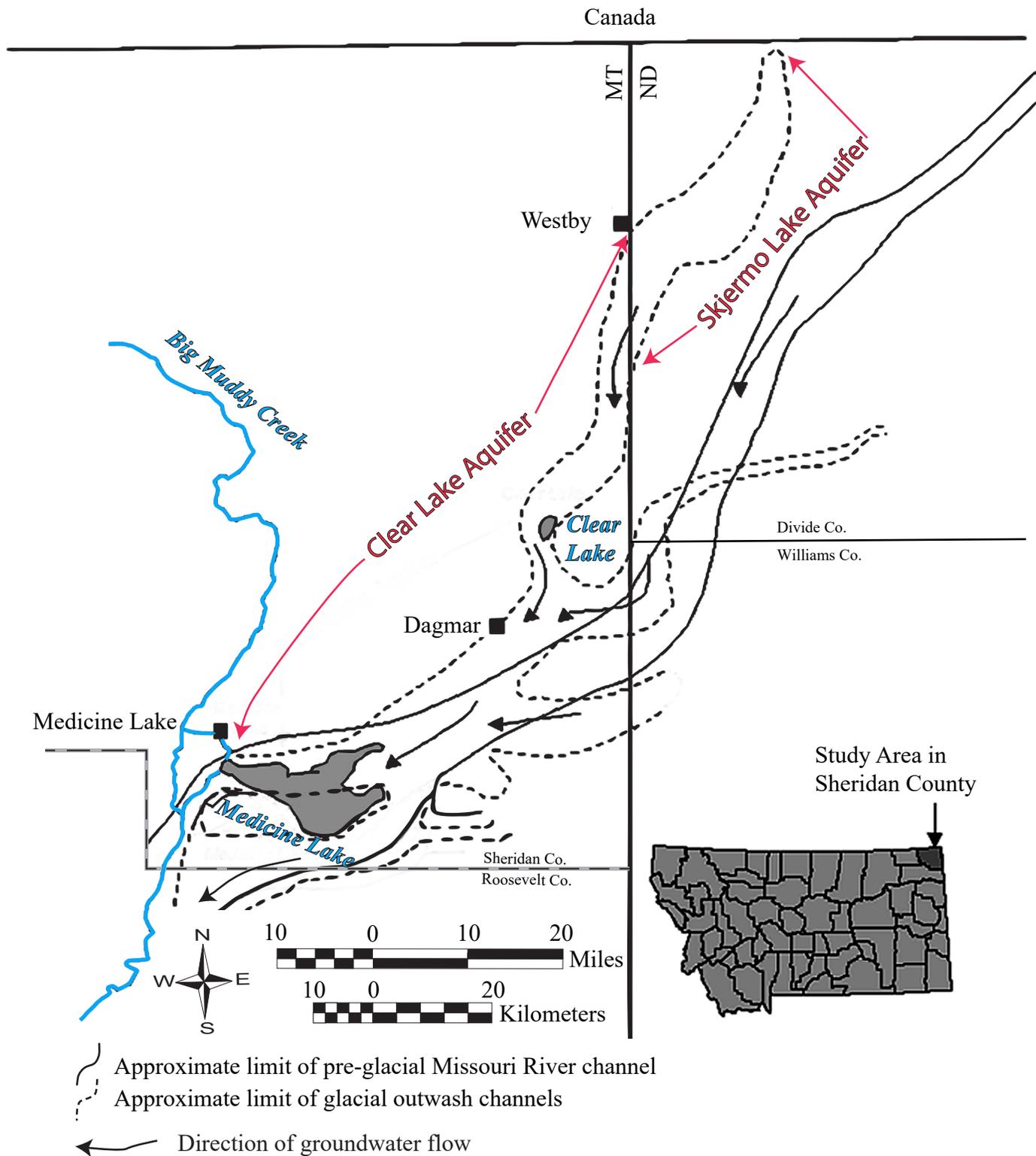


Figure 1. Sand and gravel deposits in buried valleys of the ancestral Missouri River and glacial outwash channels form the Clear Lake aquifer in Montana. The outwash part of the aquifer is called the Skjermo Lake aquifer in North Dakota (modified from Donovan, 1988).

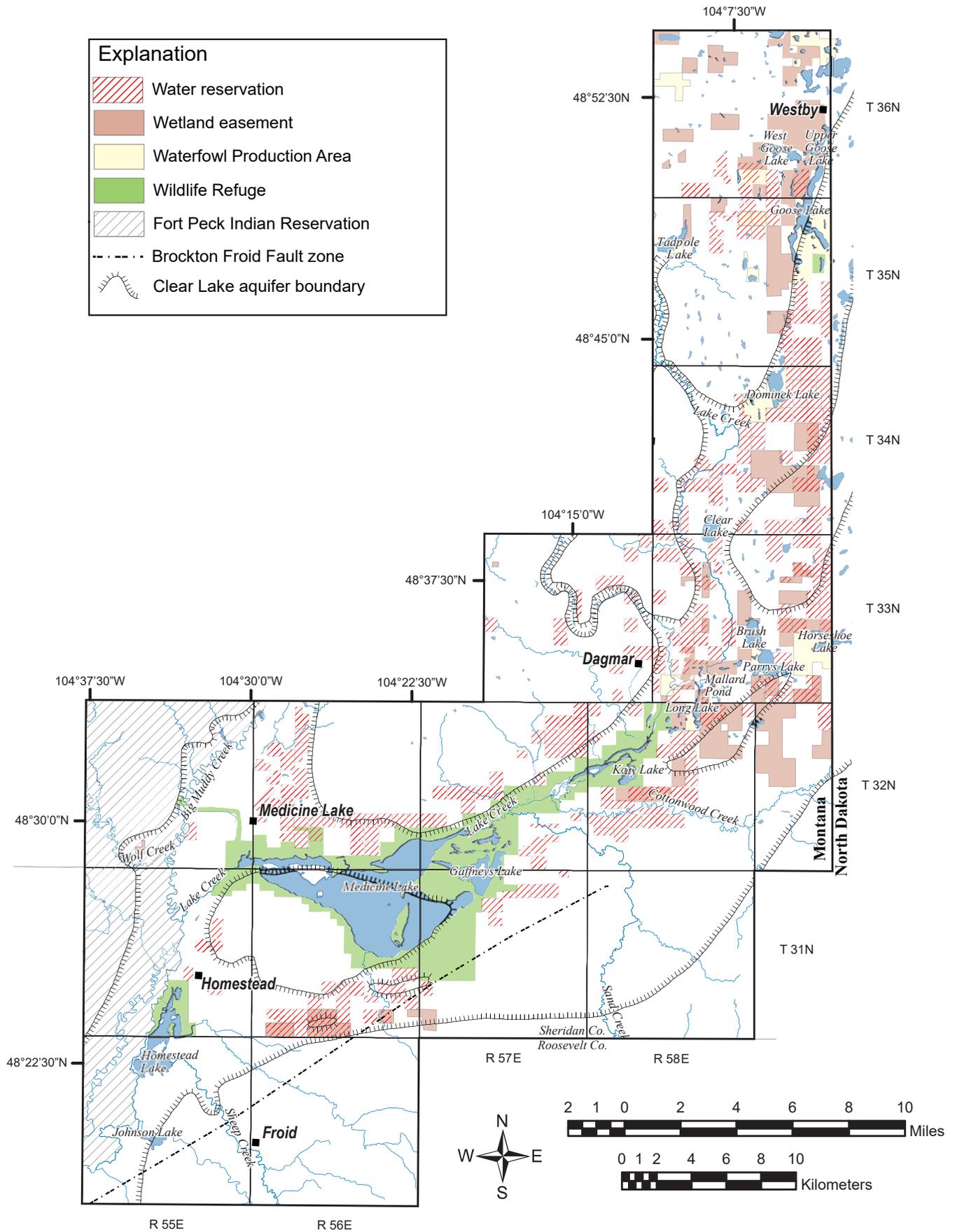


Figure 2. The Clear Lake aquifer is located in northeastern Montana. This map displays the aquifer extent, SCCD water reservation, and USFWS properties.

*iii. Develop a numerical groundwater flow model of the South Medicine Lake focus area.*

The numerical model provides a tool to guide water-management decisions by simulating increased water use from potential new wells. Detailed information on model construction, calibration, and water-use scenario results are presented in the companion report by Chandler and Reiten (2019).

*iv. Model water-use scenarios in the South Medicine Lake focus area.*

The five water-use scenarios simulated with the groundwater flow model included pumping at 2015 volume, no pumping, and three increased pumping scenarios.

*v. Evaluate variability of water quality with respect to irrigation development and potential impacts of releases associated with hydrocarbon extraction.*

Not all of the groundwater in the Clear Lake aquifer is suitable for irrigation. In some areas naturally elevated sodium produces high salinity hazards for irrigation, and clay soils exacerbate these problems, reducing crop production. Historic releases of chloride brines coproduced with hydrocarbons plus hydrocarbon compounds have been detected in samples from lakes, wetlands, streams, and groundwater associated with the Clear Lake aquifer. Identifying contaminated areas and those with elevated salinity will help guide future development.

### **Project Location**

The Clear Lake aquifer study area encompasses 586 mi<sup>2</sup> in northeastern Montana (fig. 2). The numerical model area comprises 73 mi<sup>2</sup>, mostly in the South Medicine Lake management area (fig. 3).

### **Resource Management**

A Technical Advisory Committee (TAC) made up of representatives from the SCCD, MBMG, Department of Natural Resources and Conservation (DNRC), USFWS, Fort Peck Tribes, Natural Resource Conservation Service (NRCS), and Sheridan County Planning Board was established to review requests for new appropriations and their potential effect on water resources. The SCCD through the TAC manages irrigation development to reduce or eliminate potential detrimental effects resulting from new groundwater

use. Detrimental effects include well interference, which may reduce the performance of nearby irrigation wells or make them unusable. Such impacts have not been observed as of 2016.

In the management process, SCCD has conducted or contracted test drilling, water-level monitoring, water-use monitoring, water-quality analyses, data compilation, mapping, and hydrogeologic interpretation required to assess irrigation development from the Clear Lake aquifer. The scrutiny applied to permitting irrigation wells is much greater now than prior to establishment of the SCWR.

The development of the South Medicine Lake groundwater model (Chandler and Reiten, 2019) through this project added another management tool. The model is useful to simulate hydrologic conditions associated with additional irrigation development in the South Medicine Lake focus area. The TAC, assisted by hydrogeologists, will be able to simulate new groundwater withdrawals using the model, and evaluate potential impacts. The USFWS may use project results to evaluate their water rights and to assess how releases from hydrocarbon extraction may affect water quality of wetlands and lakes. The Fort Peck Tribes may use information from this report and model to assess effects from irrigation to flows in Big Muddy Creek.

### *Sheridan County Water Reservation*

Irrigation development of the Clear Lake aquifer started in the mid-1970s under pre-SCWR Montana water permits. Groundwater levels declined during the 1980s, evoking concern among resource managers that the new groundwater withdrawals caused these declines. The 1980s also coincided with a period of drought. To address these concerns, negotiations among the SCCD, USFWS, and DNRC resulted in the creation of the SCWR in 1995. Administered by the SCCD, the SCWR reserved 15,479 acre-ft of water for irrigation purposes, with initial development limited to 5,809 acre-ft. In addition to managing water developed under the SCWR, the SCCD monitors and evaluates water use at wells supplying 26 pre-reservation water permits. By 2002 the pending applications and developed water authorizations had approached the 5,809 acre-ft cap of the SCWR. At a 10-yr review in 2004, the SCCD requested (and DNRC raised) the water development cap to 10,000 acre-ft.

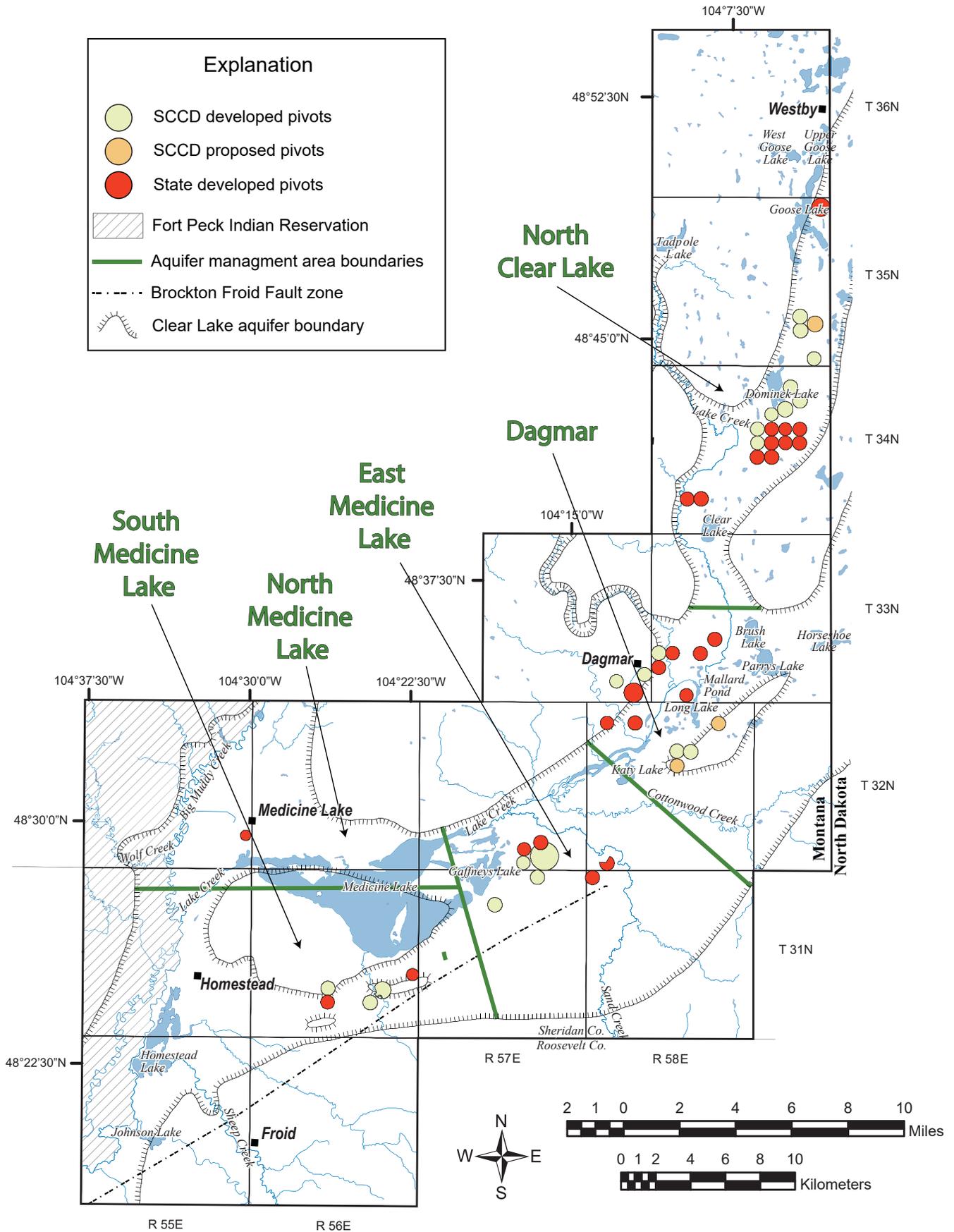


Figure 3. The Clear Lake aquifer is divided into five management areas based on irrigation development patterns, general aquifer conditions, and proximity to locations of recharge and discharge. The pivot symbol sizes approximate the acreage irrigated.

Data requirements for each permit application have tempered the growth in water development. Applicants must document that water quantity, water quality, and soil types are adequate for the proposed irrigation project and that the project is technically feasible. Hydrogeologists from the MBMG assist the TAC review of applications. Hydrographs from nearby monitoring wells are used to determine water-level trends and to assess potential impacts of irrigation projects on wetlands, which are classified by USFWS regarding their significance for habitat and the local environment.

The initial hydrogeologic investigation program started with test-hole drilling, water-level monitoring, water-quality sampling, aquifer testing, water-use monitoring, and numeric modeling. These efforts provided scientific data for the water reservation proposal and initial hydrogeologic report (Donovan, 1988). Monitoring efforts were limited until 1995 when the SCCD, DNRC, and USFWS realized that continued water-level and water-quality monitoring were essential to evaluate the effects of increased irrigation development. The MBMG reestablished monitoring for the SCWR and stored the water-level and water-quality data in the MBMG Groundwater Information Center (GWIC) database. State and Federal funding have periodically allowed additional test drilling to delineate aquifer extent and thickness, and expand the monitoring network.

#### *Water Use in the SCWR*

Figure 3 shows the aquifer management areas and status of irrigation development in the Clear Lake aquifer. The map displays irrigation pivots permitted under State water rights and the SCWR-permitted pivots. Each irrigation water right includes a point of diversion (irrigation well or wells) and a point of use (pivot or pivots). The sizes of the pivots on the map indicate the approximate areas permitted for irrigation. Groundwater use increased from about 600 acre-ft/yr when first reported in 1980 to nearly 6,000 acre-ft/yr in 1998. Although the volume of water authorized has increased since 1998, the annual use has not exceeded the 1998 volume (Reiten, 2002). Because power costs to run the irrigation system are large, producers only irrigate to supplement precipitation when needed. Therefore, the timing and volume of rainfall during a growing season affects irrigation groundwater withdrawals. The SCWR typically authorizes 2 ft annually,

but farmers rarely apply over 1 ft of water even in dry years.

#### *Aquifer Management Areas*

We divided the Clear Lake aquifer into five distinct management areas based on development patterns, general aquifer conditions, and proximity to recharge and discharge (fig. 3). The management areas from northeast to southwest (downgradient) are designated North Clear Lake, Dagmar, East Medicine Lake, North Medicine Lake, and South Medicine Lake. There are 26 State water permits in these management areas authorizing 7,889 acre-ft, and 21 SCWR permits authorizing 5,848 acre-ft of withdrawal each year.

#### **Previous Investigations**

Members of the Lewis and Clark expedition recorded the first geologic observations in this region in 1805–1806 as they traveled along the Missouri River. They described general geologic/geomorphic features including coalbeds, clinker, salts, burning coalbeds, erosional features, and glacial erratics along with descriptions of the region's physiography and topography. In the late 1800s and early 1900s, additional field studies described the physiography and geology of northeast Montana (Warren, 1868; Upham, 1904; Leonard, 1916). At about the same time, Beekly (1912), Bauer (1914), Collier and Thom (1918), and Alden (1924) published on topics ranging from regional mapping (emphasizing Quaternary geology) to coal resource investigations. The first published observations of the ancestral Missouri River Valley and its buried valley alluvial deposits came from this phase of geologic field studies. Researchers also described the Flaxville gravel and its relation to other terrace deposits in the region.

Mapping of glacial deposits since the early 1900s has progressively improved the understanding of the numerous advances and retreats of continental glaciers during the Pleistocene. Calhoun (1906) developed the first map depicting the extent of glaciation in northeastern Montana. Alden (1932) mapped glacial geology of eastern Montana and surrounding areas. Howard (1960) mapped Cenozoic deposits emphasizing the Pleistocene in Montana and North Dakota, centered on the confluence of the Yellowstone and Missouri Rivers. Colton and others (1961) mapped glacial geology in eastern Montana. As researchers incorporated new regional information, they developed a better chronol-

ogy of glacial advances. Mapping by Clayton and others (1980) in North Dakota emphasized the Quaternary geology, glacial chronology, and landforms associated with glacial deposits. Fullerton and Colton (1986) interpreted the glacial geology and chronology of the Montana plains. Bergantino (1986) compiled a map of the Quaternary geology covering the study area. Fullerton and Colton (1986), Colton and others (2006a,b), and Fullerton and others (2007) developed Quaternary geologic maps and the current accepted chronology of glaciation in the region.

Geologic and hydrologic investigations in the Medicine Lake area resulted from proposed water-resource development associated with reservoir construction in the late 1940s. Wilson (1948) proposed the first reservoir on Big Muddy Creek. A proposal to construct a reservoir at Medicine Lake soon followed (Kirchen, 1949) and included an electrical resistivity study to identify the buried valley aquifer (Edwards, 1951). Witkind (1959) published a geologic map covering most of the present study area as part of these reservoir investigations.

During the 1950s and 1960s, groundwater research emphasized county or provincial studies of aquifers to determine water availability. In Montana, research focused on groundwater resources in the Missouri River Valley (Swenson, 1955). Levings (1986) investigated groundwater conditions in the Plentywood area as a result of proposed water use for a newly constructed coal power plant in Canada. The new water reservation for the Clear Lake aquifer required detailed hydrogeologic data and interpretations as reported in Donovan and Bergantino (1987) and Donovan (1988). Schuele (1998) produced a groundwater model to evaluate the water balance and impacts of irrigation development on lakes and wetlands. Both the Donovan (1988) and Schuele (1998) models simulated the entire aquifer system to forecast aquifer response to climate and land-use changes. Although the models provided water-budget estimates, they were too simplistic to serve as management models and did not evaluate the impacts of increased use at specific locations.

Donovan and Rose (1994) modeled the water chemistry of lakes and concluded that the flow of shallow and intermediate groundwater from the Clear Lake aquifer into the lakes explains the lakes' water chemistry. The lake water chemistry ranges from fresh water to highly concentrated brines.

In North Dakota, research focused on Divide and Williams Counties along Montana's northeast border. Armstrong (1965, 1967) mapped the geology and evaluated the hydrology of buried valley aquifers. Hansen (1967) and Freers (1970) mapped the geology of Divide and Williams counties bordering Montana. The Skjermo Lake aquifer is the continuation of the Clear Lake aquifer into North Dakota. A series of annual reports by Wanek (1983–1996), in addition to a groundwater model (Wanek, 1984), describe the Skjermo Lake aquifer and water use. Shaver and Pusc (1992) described longitudinal and transverse hydraulic barriers in buried valley aquifers and their ability to alter groundwater flow and availability.

In Canada, aquifers associated with buried valleys of the ancestral Missouri and Yellowstone Rivers have experienced extensive water withdrawals, monitoring, and modeling (Cummings and others, 2012). Christianson and Parizek (1961) summarized studies of these aquifers near Estevan, Saskatchewan. van der Kamp and Maathuis (2012) evaluated excessive drawdown in buried valley aquifers of the Estevan Valley. Smerdon and others (2005) investigated and modeled the hydraulic connection between a buried valley aquifer and a large irrigation reservoir.

Since the early 1980s the MBMG has provided hydrogeologic data and interpretations to the SCCD. MBMG hydrogeologists have served on the SCCD (TAC) since 1995. Responsibilities include assisting with the maintenance of the monitoring program, interpreting water-level data, conducting and interpreting aquifer tests on new irrigation well applications, and coordinating with the USFWS and Fort Peck Tribes on SCWR activities. The SCCD and MBMG collect, annually compile, and enter monitoring data into the MBMG GWIC system. These data provide background for the water permit application process (SCCD, unpublished file data). Reiten (2002) summarized the results of the monitoring program. The TAC determined that dividing the aquifer into separate management areas would improve development decisions. Reiten (2002) included hydrographs from monitoring and irrigation wells that demonstrate short- and long-term trends in groundwater levels, a discussion of aquifer conditions, effects of pumping on groundwater levels and surface-water resources, and potential for additional development for each management area.

## Regional Setting

### *Physiography*

Big Muddy Creek and its tributaries overlie the Clear Lake aquifer and flow into the Missouri River 26 mi south of Medicine Lake. An exception is the Brush Lake internally drained area located northeast of Dagmar (plate 1). This area has poorly defined surface drainage, and many of the lakes and wetlands have no surface inlets or outlets. Big Muddy Creek is the largest stream in the area and forms the study area western boundary. Other streams that overlie the aquifer include Lake Creek, Cottonwood Creek, and Sand Creek.

Numerous lakes and wetlands overlie the Clear Lake aquifer area. Several of the larger named lakes (from north to south) are Goose Lake, Dominek Lake, Clear Lake, Brush Lake, Mallard Pond, Long Lake, Katy Lake, Gaffney Lake, Medicine Lake, and Homestead Lake. Medicine Lake is the largest of these, covering an area of about 8,214 acres at the full pool level (1,933 ft above mean sea level as reported by the USFWS). At this level, Medicine Lake holds 48,580 acre-ft of water. It is shallow, with a maximum depth of about 10 ft at full stage. Artificial controls regulate water levels in Medicine Lake. Big Muddy Creek is diverted to the lake during high streamflows to increase lake stage.

The topography of the study area is one of relatively low relief. Elevations range from 2,100 ft in the northern area near Westby to about 1,930 ft in the southwest, where the aquifer underlies Big Muddy Creek. Most of the modern streams are underfit and occupy relatively broad valleys derived from glacially diverted streams or meltwater channels.

### *Geology and Drainage Development*

The landscape of northeastern Montana has evolved with Tertiary (Flaxville) and Quaternary (Wiota) deposition and erosion by rivers that flowed to the northeast in the geologic past. These fluvial systems formed a series of gravel-capped terraces that step down in elevation from the northwest to the southeast. The terraces slope towards the northeast, where they underlie glacial till. Plate 2 depicts the regional geology, physiography, and drainage in the Wolf Point quadrangle in northeast Montana. The Flaxville plateaus are highly dissected and their edges show

distinct dendritic erosional patterns exposing underlying fine-grained bedrock. While the Clear Lake aquifer occupies a relatively small part of this area, this view aids in placing the aquifer within the regional geologic setting and landscape.

During Miocene–Oligocene time, the ancestral Missouri River flowed to the northeast down the regional paleo-slope from just north of the Canadian border towards Hudson Bay and ultimately the Labrador Sea (fig. 4; Cummings and others, 2012). Tectonic uplifting in Canada and stream piracy south of the border forced the channel to the southeast. Alluvial terrace deposits formed at each new stream position (fig. 5). The Cypress Hills gravel (1 in fig. 5) is a remnant of the oldest fluvial deposit (Eocene–early Miocene). The river system shifted south into Montana during the Oligocene and Pliocene, depositing successive lower terraces of the Flaxville Formation (2). During the Pleistocene, several terraces developed as the alluvial system migrated southeast (Donovan and Bergantino, 1987). These ancestral Missouri River alluvial deposits are referred to as the Wiota gravel (3) in this report. Green and blue lines show the approximate locations of buried scarps of the Wiota gravel. Although the Wiota gravel as described by Jensen and Varnes (1964) was not projected into the subsurface, we use this name to describe subsurface deposits of the basal gravel in the buried valley (4). The gravel is identical to higher terrace deposits and it forms the final “Wiota” deposition in the buried valley (5). The Brockton–Froid Fault zone appears to have tectonically controlled the southeast migration of the ancestral Missouri River.

The limits of Quaternary glacial ice advance as described by Fullerton and others (2004) are also featured in plate 2. The unglaciated areas associated with the Flaxville terraces are notable by the relatively high elevation and flat surfaces. Plate 2 also identifies several broad swales: these are defined by areas of low relief within topographic lowlands. The Manning Lake swale and the Medicine Lake swale trend southwest to northeast and delineate the preglacial Missouri River Valley. These swales are 6–8 mi wide and are filled by alluvium and glacial deposits, including outwash, lake sediment, and till. The north- to south-trending Clear Lake swale is 1–4 mi wide, and is filled by glacial deposits, including outwash, lake sediment, and till.

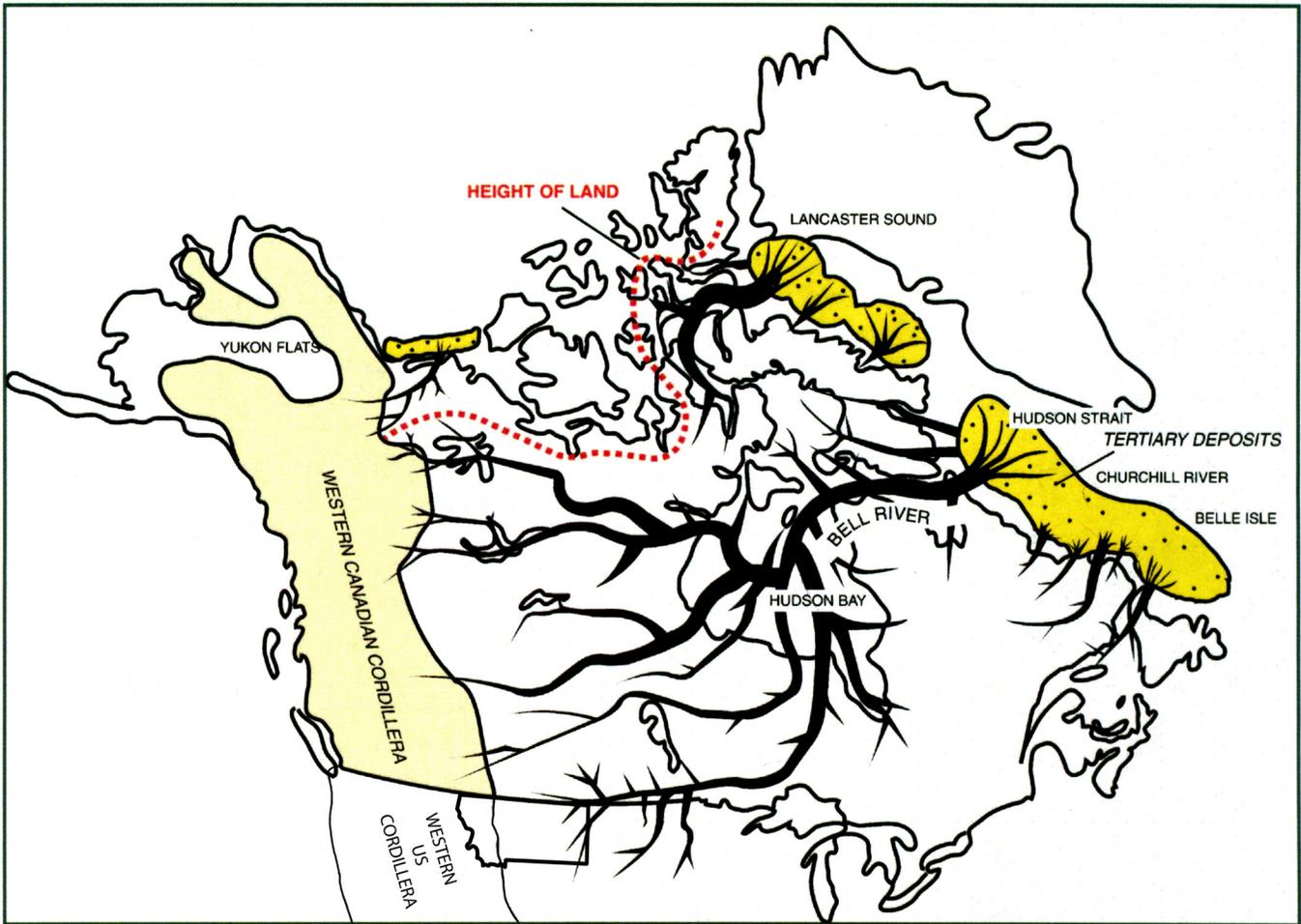


Figure 4. Conceptual depiction of Tertiary (preglacial) drainage in northern United States and Canadian prairies. Adapted from Cummings and others (2012) with permission.

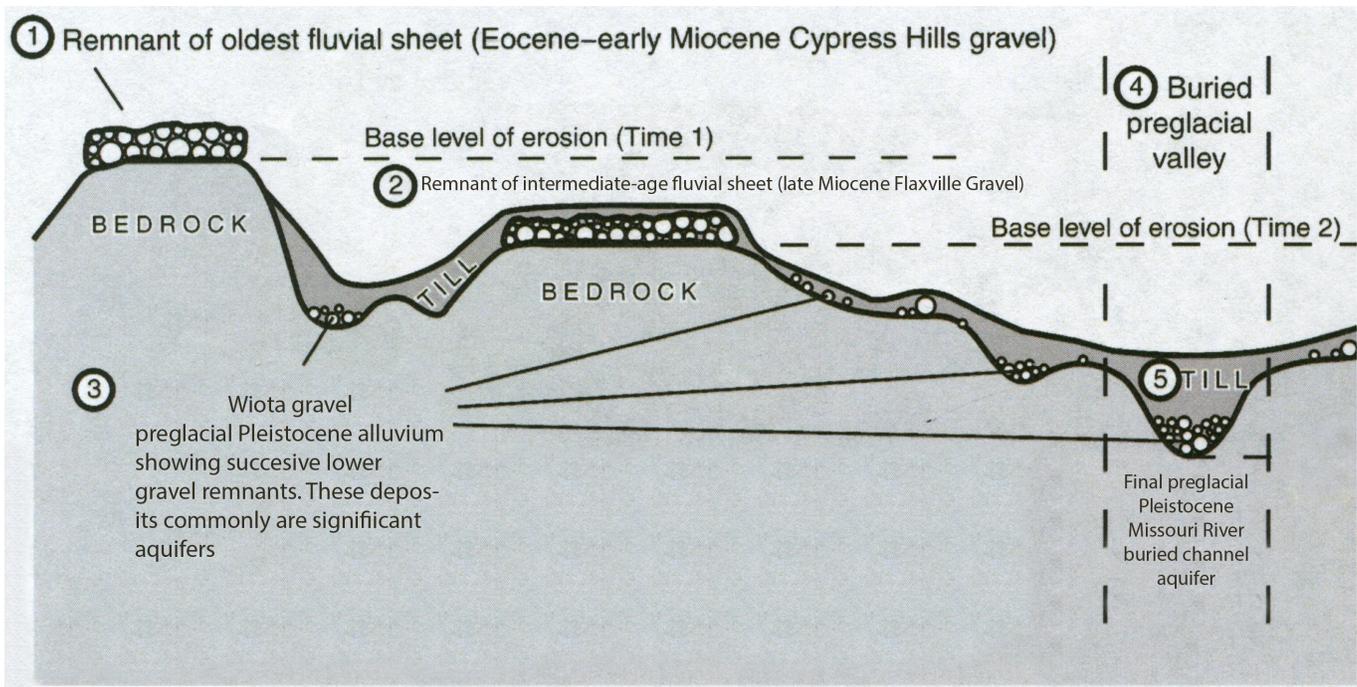


Figure 5. Illustration depicting the stratigraphic position, relative age, and relative position of Rocky Mountain derived quartzite-rich gravel during Tertiary erosion in the Great Plains of northern Montana and Canada. Adapted from Cummings and others (2012) with permission.

Quaternary glaciers dammed and eventually diverted the Missouri to its present course (Cummings and others, 2012). These diversions commonly occupied the low-lying swales, but flow was generally opposite the regional paleo-slope or around the margins of glacial ice. Multiple advances and retreats of glacial ice buried the alluvial gravels beneath glacial till, preglacial lake clays, fine-grained deltaic deposits, and coarse-grained glacial outwash. Glacial meltwater deposited outwash in these low-lying swales multiple times. Meltwater incised channels 200 to 300 ft deep into some of the larger drainages, and these channels were subsequently filled with glacial sediments.

Advancing glaciers diverted several of these drainages at a variety of scales (plate 2). A large-scale diversion of the Missouri River east of Culbertson blocked the river, forming the Bainville diversion, leaving the Bainville swale in the former river position. The Antelope diversion and swale result from a medium-scale diversion. Numerous smaller scale diversions appear on the map, including the Coalridge channel near Dagmar, and the unnamed channel west of the Big Muddy near Plentywood, which trends generally southwest crossing Wolf Creek and Smoke Creek.

The Clear Lake aquifer is composed of two zones: (1) the lower pre-glacial alluvium (Wiota gravel) deposited by the northeast-flowing ancestral Missouri River in the Medicine Lake swale, and (2) upper glacial outwash sand and gravel deposited by southwest-flowing meltwater streams in the Clear Lake swale and adjacent part of the Medicine Lake swale. The outwash deposits overlie the alluvial deposits, older glacial deposits, and Fort Union bedrock depending on the specific location. In some places, locally extensive aquitards composed of fine-grained glacial till and lake clay separate the aquifer zones. In other locations the upper outwash and lower alluvial zones are hydraulically connected. The aquifer in our study area extends from Westby to the southwest where it directly contacts alluvial and outwash sand and gravel underlying Big Muddy Creek. Groundwater flow in the Clear Lake aquifer is generally towards the south or southwest.

### *Climate*

Eastern Sheridan and Roosevelt Counties have a semiarid continental climate, characterized by cold,

dry winters; cool, moist springs; moderately hot, dry summers; and cool, dry autumns. January is generally the coldest month with an average low temperature of  $-2^{\circ}\text{F}$  and July the warmest month with an average high temperature of  $84^{\circ}\text{F}$  (based on 1960–2015 data). At Westby, near the northern extent of the aquifer, the average precipitation is about 14.2 in per year, with most of the precipitation falling during the growing season from May through August (fig. 6; WRCC, 2018, Westby, Montana Station 248777).

Near the southern extent of the aquifer at Medicine Lake, the average precipitation is about 13.8 in per year (WRCC, 2015 Medicine Lake, Montana Station 245572). During the 11-yr period from 1979 to 1990, an extended drought (1980s drought) affected water resources in northeastern Montana. In 9 of the 11 drought years, the annual precipitation was below the 1960–2015 average (fig. 7). Extreme drought, defined here by annual precipitation less than 10 in, occurred during 7 of these 11 years. Shallow wetlands dried and lake levels declined. Likewise, aquifer recharge was diminished, resulting in groundwater-level declines.

Although potential evaporation typically exceeds precipitation in this area, estimates of evaporation vary widely. The Medicine Lake 3 SE weather station ( $48.29^{\circ}$ ,  $-104.27^{\circ}$ ) reports a long-term pan evaporation average of 39.30 in for the period of record 1911–2005 (WRCC). A comparable station with a long-term Class A evaporation pan at Sidney, 50 mi south, was used in previous studies. Long-term data from the Sidney site report average annual evaporation of 33.14 in, period of record 1910–2005 (WRCC). Evaporation from a Class A pan at the U.S. Department of Agriculture research farm near Froid indicated 53% higher evaporation rates than at Sidney. The average evaporation at the Froid station was 52 in/yr from 1984 to 1988, during extreme drought conditions. In contrast, the Sidney station reported 34 in of annual evaporation over the same time period (Donovan, 1988). Wind and hot temperatures contribute to an average of 6 to 12 in of monthly evaporation from May to August at the Froid site. For the modeling effort conducted as part of this study, Chandler and Reiten (2019) assumed a potential evapotranspiration rate of approximately 3 ft/yr.

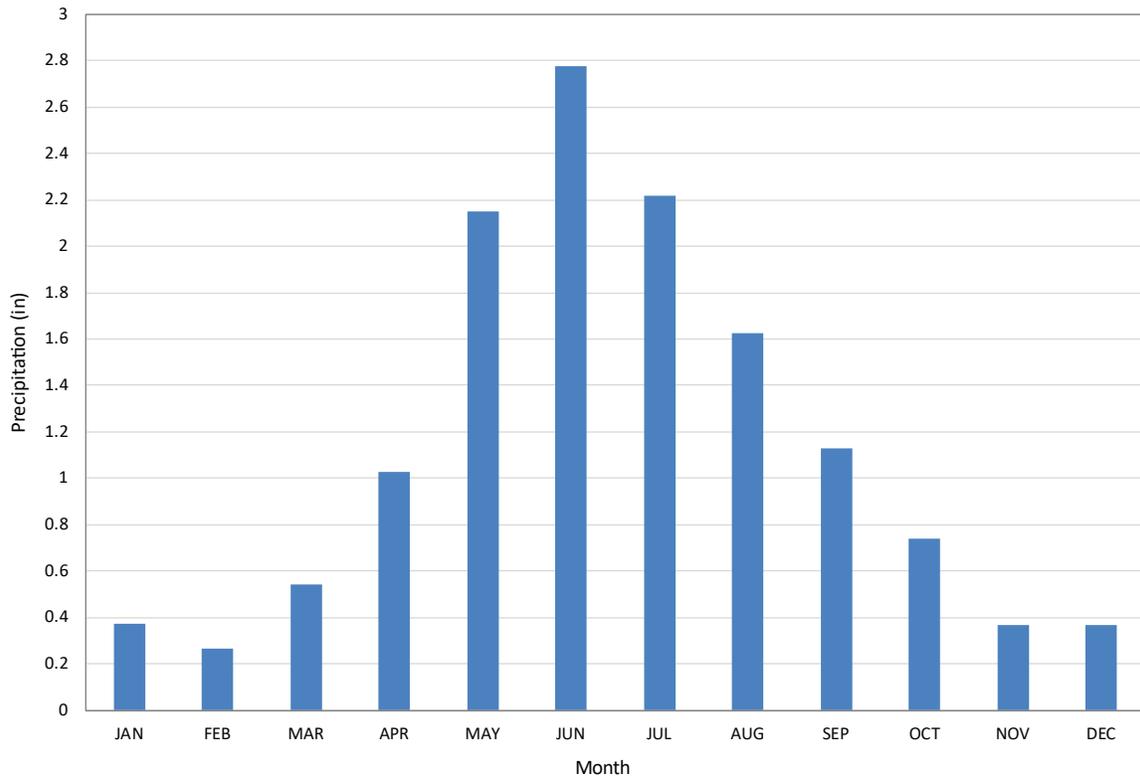


Figure 6. Average monthly precipitation for Medicine Lake Based on the period 1960–2015. <http://www.wrcc.dri.edu/cgi-bin/cliMAIN.pl?mt5572> [Accessed March 1, 2016].

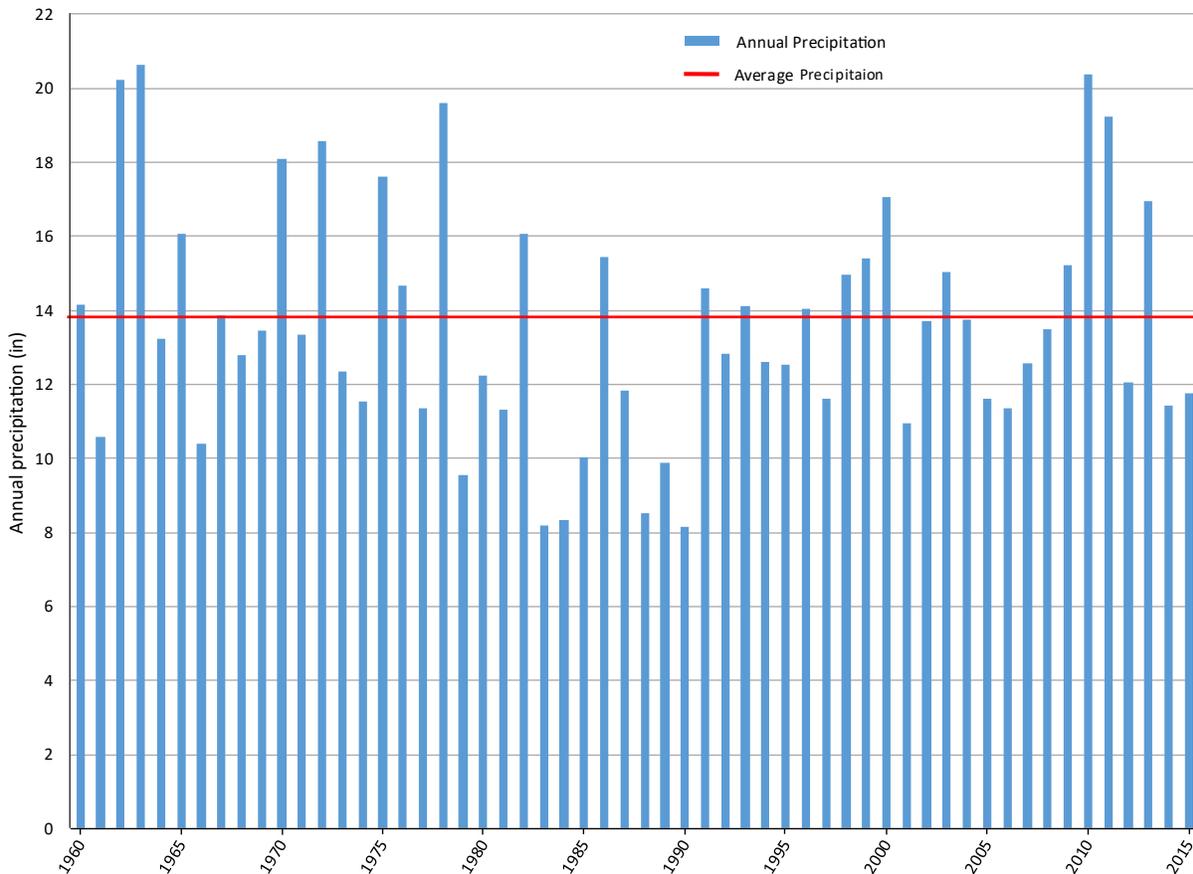


Figure 7. Annual precipitation for Medicine Lake from 1960 to 2015. <http://www.wrcc.dri.edu/cgi-bin/cliMAIN.pl?mt5572> [Accessed March 1, 2016].

## METHODS

### Data Compilation and Management

MBMG hydrogeologists compiled and verified Clear Lake aquifer data collected from 1980 to 2016. These data include lithologic and well completion logs, water chemistry analyses, and aquifer test reports stored in GWIC. These data are accessible through the GWIP project pages (<http://mbmaggwic.mtech.edu/sqlserver/v11/menus/menuProject.asp?mygroup=GWI&myroot=BWIPCL&ord=1&>).

### Geologic (Lithologic) Mapping

The parent materials of the different soil series (Richardson and Hanson, 1977) were used to construct a surficial geologic map based on the NRCS soil survey of Sheridan County. NRCS soil maps were used in ArcGIS to classify specific geologic units based on the parent material for that soil type. We conducted limited field investigation to verify the map units and confirm the geologic interpretations. The resulting lithostratigraphic map emphasizes the hydrogeologic characteristics of each mapped unit.

### Hydrogeologic Framework

Lithologic logs were simplified and entered into a digital format. This included verification and correction of well locations as needed. These logs show heterogeneous lithologic sequences common in glaciated terrains, where large lateral and vertical stratigraphic variation exists between closely spaced boreholes or within a single borehole. Lithologic descriptions were translated to major hydrogeologic units by combining materials with similar hydrologic properties. These modifications brought consistency to the wide range of terminology used by drillers and geologists. The simplified logs provided the information for developing stratigraphic cross sections, constructing aquifer maps, and groundwater modeling.

### Water Use

Water-use records were compiled for the project area. Data sources included electric utility records, flowmeters, and engine hour meters. Water-level responses to pumping in observation wells were used to determine when nearby irrigation wells were operated. Most methods for estimating water use require knowledge of the average pumping rate at each irrigation well. In the past, pumping rates were estimated by the pump installer or read directly from inline flowme-

ters. However, the accuracy of flowmeters decreases with time, and the SCCD used an ultrasonic flowmeter to improve accuracy in these measurements. The ultrasonic flowmeter used in this study was factory calibrated and the manufacturer claims pumping rate accuracy to  $\pm 1$  percent of the actual flow.

### Monitoring Network

#### *Groundwater*

Data from a network of 189 wells were used for this project. These included observation wells, stock and domestic wells, and irrigation wells. These wells, shown on plate 1, are listed in appendix A.

The frequency of water-level measurements was variable. Observation wells adjacent to irrigation wells have data loggers programmed to record at 15-min intervals. Data loggers in other wells recorded water levels hourly. Water levels in irrigation wells are measured twice a year, typically with a steel tape. Late fall measurements record water levels during recovery from the recently completed irrigation season, and early spring measurements show recovered levels, prior to the onset of irrigation.

#### *Surface Water*

Prior to 2008, water levels in lakes and wetlands were monitored with staff gages. Most of these gages failed after a few years from ice jacking and damage from floating ice. These have been replaced with shallow wells constructed along shorelines of lakes and wetlands. The SCCD monitors 11 of these shallow wells to record surface-water fluctuations.

#### *Site Surveys*

The elevations and coordinates of all monitoring and irrigation wells were surveyed with a Leica 1200 GNSS GPS system. The well survey utilized a base station and rover to provide inch accuracy for altitude and the latitude–longitude at each location.

### Water Quality

The GWIC database contains 383 analyses collected in this area since 1960 (appendix B). Most samples were collected using MBMG Standard Operating Procedures (SOPs). Water was purged from the wells using pumps, bailers, or airlift methods. Purging continued until about three casing volumes of water was removed or until temperature, specific conductivity

(SC), and pH had stabilized. Surface water was collected by pumping or grab sampling from the wetland surface.

Several agencies collected water-quality samples contemporaneous with this GWIP project. The US-FWS sampled surface water for inorganic constituents at 34 sites, and the SCCD sampled 27 wells as part of a coalbed-methane monitoring program. We also accessed results from 147 samples collected from wells and wetlands during previous MBMG, USFS, Department of Environmental Quality (DEQ), and DNRC-funded projects.

Samples submitted to the MBMG lab were field processed using the MBMG sampling protocol. This requires filtered and preserved samples for dissolved constituents and unfiltered samples for total recoverable constituents. All samples were cooled with ice or refrigeration to below 42°F prior to analysis. Analytes tested at the MBMG Lab included major anions and cations and a standard list of trace elements.

Samples collected by well drillers or well owners since 1960 also provide some water-quality data. These samples, typically sent to private labs, were analyzed for specific conductance (SC), total dissolved solids (TDS), and the sodium absorption ratio (SAR). The parameters tested were limited to those needed to determine the usability of the groundwater for irrigation. We assume that these samples were untreated and collected at the wellhead or an available tap.

### Potentiometric Surface Map

A potentiometric surface map was constructed using water levels measured and water levels taken at other time periods but adjusted based on records from nearby wells completed at similar depths. Water levels measured in April–May 2015 were recovering from the 2014 irrigation season and did not appear to be affected by spring recharge prior to the 2015 irrigation season. The locations without April–May 2015 data were estimated by taking the non-pumping water-level data measured closest to the April–May timeframe and adjusting it in accordance with trends in nearby wells completed in the same zone. Water levels changed slightly during these months and the adjustments made to data from wells without April–May 2015 measurements were in the 2–5 ft range.

### Aquifer Test Data

Aquifer test data from previous investigations and water-permit applications (Donovan, 1988; Schuele, 1998; Reiten, 2002) were compiled for this investigation. These include single well tests (production well only) and multiple well tests where observation wells were available. Aquifer parameters determined from these tests are based on various analytical methods. Additional aquifer tests in the South Medicine Lake model area were analyzed by the MBMG using transducer data from observation wells located next to irrigation wells (appendix C). Normal irrigation operation resulted in 32 fortuitous aquifer tests during the 2015 growing season. Aquifer parameters estimated from these fortuitous tests used the Cooper–Jacob method only (Cooper and Jacob, 1946).

## RESULTS—REGIONAL CONDITIONS IN THE CLEAR LAKE AQUIFER

This section presents results and findings for the regionally extensive Clear Lake aquifer. Reiten (2002) summarized the history, status, and potential for additional irrigation development in the five management areas. Here, we present the hydrogeologic framework of the region, including aquifer delineation and properties, the nature of groundwater recharge and discharge, and surface water–groundwater interactions. The South Medicine Lake focus area is discussed in more detail in a later section of this report.

### Hydrogeologic Framework

The Clear Lake aquifer in Montana and equivalent Skjermo Lake aquifer in North Dakota lie within a buried valley that extends approximately 40 mi. Twelve miles of the aquifer are in North Dakota and the remaining 28 mi are in Montana (fig. 1). The Clear Lake aquifer includes the lower alluvial zone (Wiota gravel) that underlies the Medicine Lake swale (see plate 2). The upper outwash zone lies within the Clear Lake swale and parts of the Medicine Lake swale. Although the ancestral Missouri River deposits differ lithologically from overlying outwash deposits, these zones form the Clear Lake aquifer. The width of the Clear Lake aquifer ranges from 1.5 mi near Clear Lake where the channel narrows (referred to as the Clear Lake constriction) to more than 8 mi wide east of Medicine Lake (fig. 2).

Figure 8 shows the surficial geology in the project area. The Paleocene Fort Union Formation is the oldest unit mapped. The Fort Union Formation consists of fine-grained mudstones and siltstones interbedded with sandstone and coalbeds. The fine-grained units form aquitards and thin coalbeds and sandstones form low-yield aquifers. Most of the area is covered by Quaternary glacial till. The till is a poorly sorted mixture of clay, silt, sand, and gravel. This till generally restricts recharge at the surface and forms aquitards in the subsurface. Quaternary glacial outwash is the dominant surficial unit in the Clear Lake swale from Westby to several miles southwest of Dagmar. It is composed of blue to light brown fine sand grading to coarse gravel. This upper outwash zone forms moderate to high-yield regions of the aquifer. Quaternary glacial lake deposits composed of gray to dark blue silty clay and clayey silt form a minor surficial geologic unit. Lake deposits in the subsurface form extensive and commonly thick aquitards. Holocene alluvium forms minor surficial deposits along smaller streams including Lake Creek, Cottonwood Creek, and Sand Creek. Extensive alluvial deposits occur along Big Muddy Creek. These deposits range from light brown to dark grayish brown silt, silty clay, silty sand, and silty sand and gravel. Quaternary eolium deposits are largely brown to reddish brown, well-sorted fine to medium sand and range from 2 to 50 ft thick. The eolium, referred to in this report as sand hills, covers several square miles southeast of Medicine Lake where it forms a groundwater recharge area.

The aquifer materials are composed of three major lithologic components. These materials were transported from the west and north, and also include locally derived rocks. Although present in the subsurface, neither the Flaxville gravel nor the Wiota gravel are exposed in this area (fig. 8). Rocks eroded from the Rocky Mountains and transported from the west include distinctively colored (purple, red, green, and tan) quartzite, probably from the Belt Supergroup, and less abundant volcanic gravel (from the Bears Paw Mountains). These are the dominant lithology of the alluvial Flaxville and Wiota gravels. The Wiota gravel contains reworked Flaxville gravel and minor amounts of Canadian Shield rocks; the Wiota gravel is largely indistinguishable from the Flaxville gravel. Northern source rocks include igneous and metamorphic rocks from the Canadian Shield and Paleozoic carbonates transported to the study area by glaciers. Local source

rocks include mudstone, sandstone, siltstone, and coal derived from the Fort Union Formation. The glacial outwash contains all three lithologic components, but is dominated by light-colored Paleozoic carbonates. In most areas, fine-grained lake clay or glacial till separates the upper outwash and lower alluvial zones of the aquifer.

The bedrock elevation contour map (fig. 9) depicts a 200- to 300-ft-deep buried valley extending from Homestead northeast towards the North Dakota border. The buried valley splits into two channels separated by a bedrock ridge near Medicine Lake. The northern channel contains both the upper outwash zone and lower alluvial zone of the aquifer. The southern channel comprises the lower alluvial zone, although hydraulically connected thin layers of outwash gravels overlie the lower alluvium. East of Medicine Lake, the two channels merge and continue in an easterly direction. Groundwater flow barriers form at the edge of buried valleys, where highly transmissive sand and gravel of the Clear Lake aquifer transitions to fine-grained bedrock of the Fort Union Formation. Several bedrock highs form ridges within buried valleys, and these can hydraulically isolate parts of the aquifer system similar to those observed in North Dakota by Shaver and Pusc (1992).

The buried valley-fill is highly heterogeneous across the study area. Complex stacking of these deposits indicates the buried valleys were reactivated during multiple glaciations. In all locations, the buried valley-fill is predominantly fine-grained sediments; these glacial till and glacial lake deposits form the aquitards. Remnants of a higher Wiota terrace form a separate aquifer capping the bedrock ridge in the South Medicine Lake area.

The Clear Lake aquifer thickness is highly variable, ranging from 10 to 250 ft. Figure 10 depicts areas of continuous sand and gravel greater than 100 ft thick located near the middle of the cross section. These highly transmissive zones, referred to as "central plugs" by Donovan (1988), are shown near well 44455, and may yield very large volumes of water to wells. Areas of the aquifer that are thinner but are hydraulically connected to these transmissive zones typically show greater aquifer transmissivity than would be expected based on gravel thickness reported in the well log.

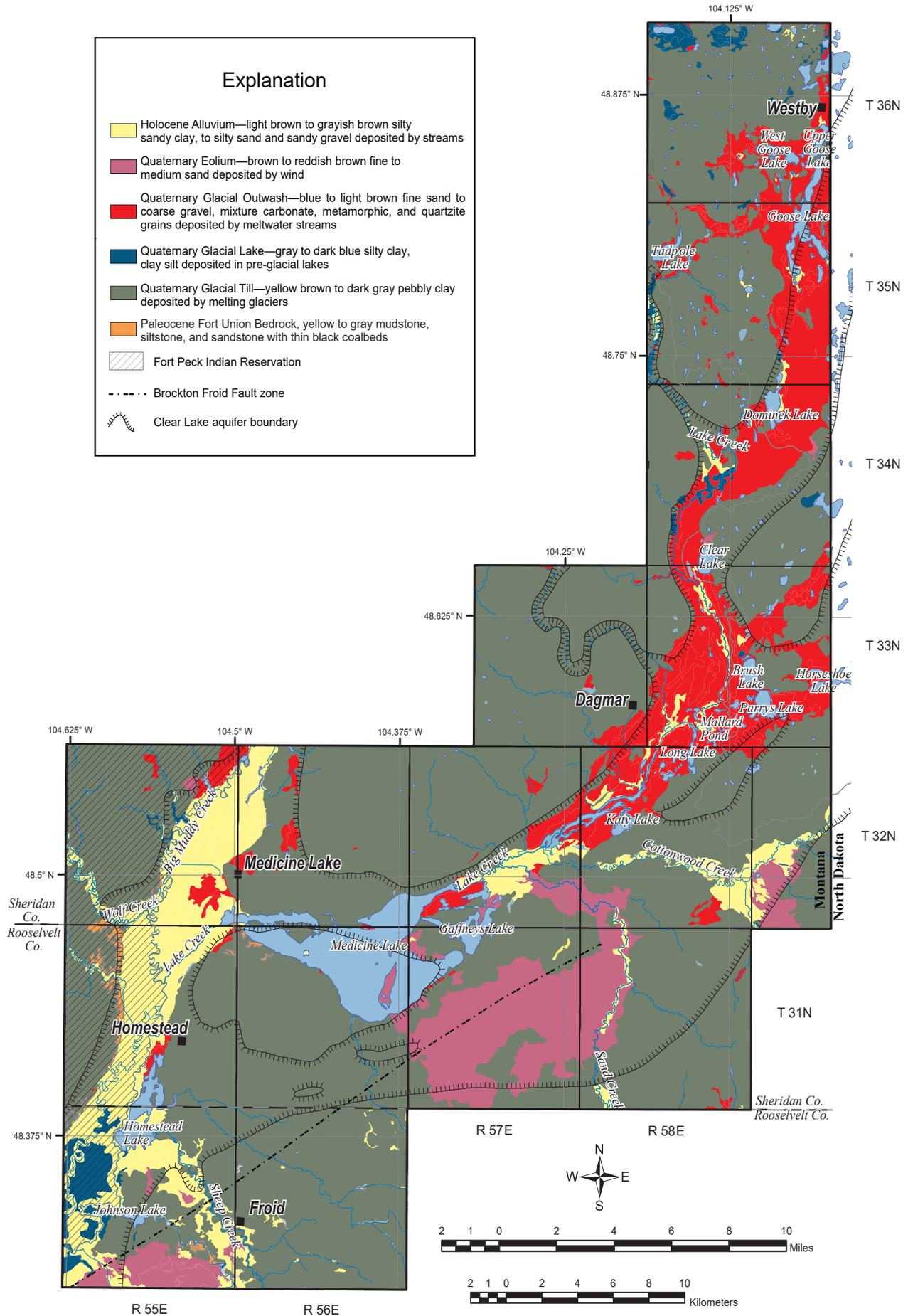


Figure 8. Surficial geology of the Clear Lake aquifer.

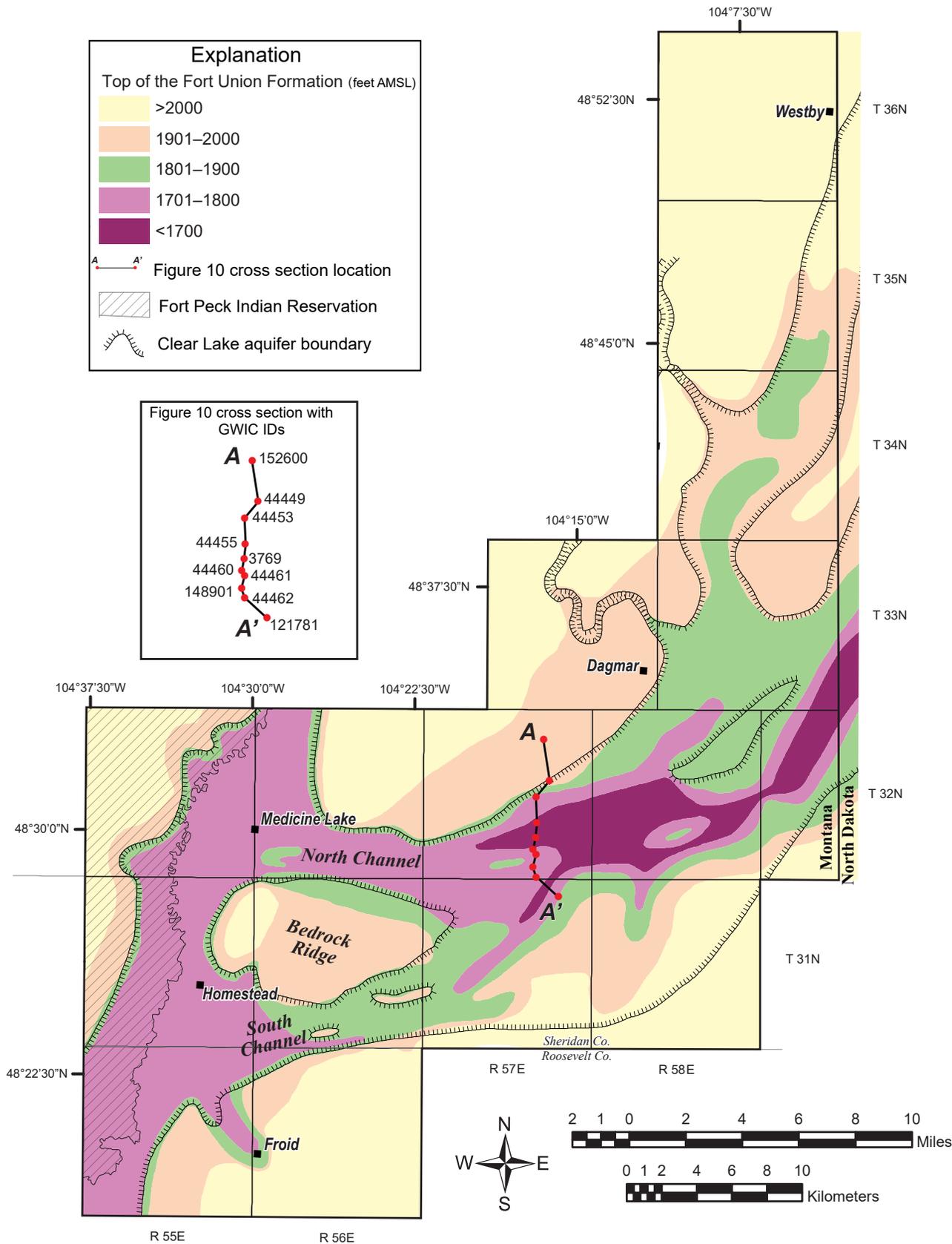


Figure 9. The top of the Fort Union Formation bedrock forms the base of the buried valley containing the Clear Lake aquifer. Contours are based on the elevation at 273 control points.

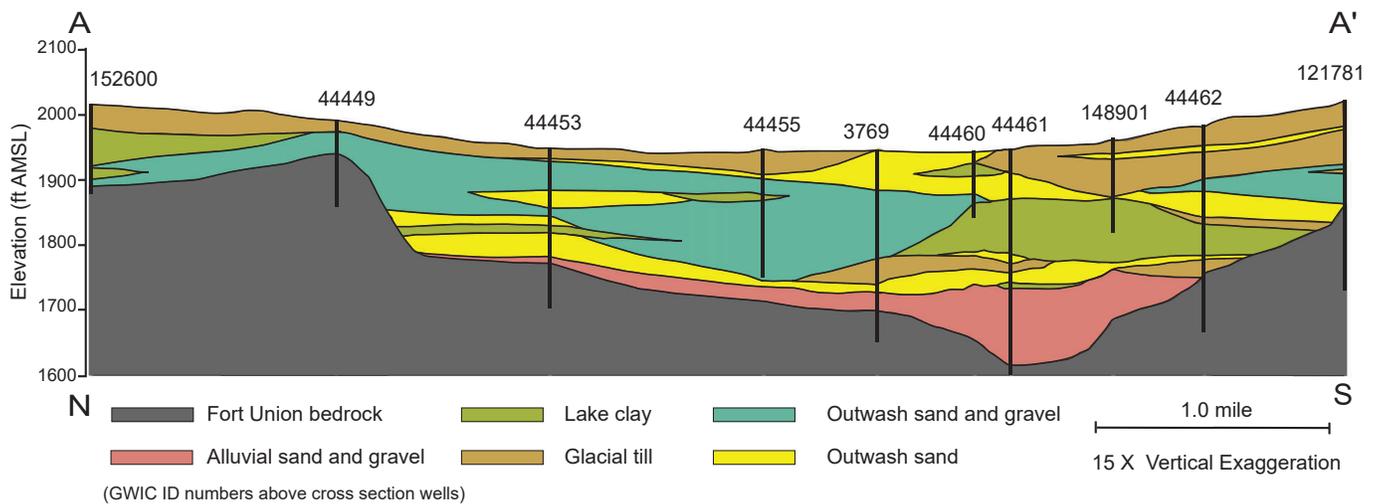


Figure 10. Geologic profile across the Clear Lake aquifer east of Medicine Lake. See figure 9 for location of the cross section.

Highly productive wells completed in relatively thin sand and gravel layers are explained using a schematic of glacial stratigraphy (fig. 11). A well at point A on figure 11 completed in *glacial outwash 1* has direct connection to the highly transmissive *glacial outwash 3*. A well at point B completed in *glacial outwash 2* has thicker outwash aquifer than the well at point A, but lower productivity because of limited storage in the aquifer. A well at point C encounters till and clay, but not the aquifer. The upper alluvial terraces are limited in aerial extent. Wells completed in the terraces, such as point D, may support domestic or stock uses, but typically lack the yield needed for irrigation. Direct recharge from precipitation and snowmelt and unconfined conditions occur in areas where the glacial outwash is exposed at the surface (fig. 11, *glacial outwash 3*).

### Groundwater/Surface-Water Connection

Results of this study support previous interpretations of the groundwater/surface-water connections of the Clear Lake aquifer (Donovan, 1988; Schuele, 1998; Reiten, 2002). The deposits form a heterogeneous system; the hydraulic connections among the aquifer, lakes, and wetlands vary from location to location. The land surface overlying the Clear Lake aquifer contains wetlands, sloughs, and lakes. Many of these water bodies are in kettles, which are depressions in the land surface formed when stagnant glacial ice melted and left behind an undulating topography of hills and depressions. These kettles formed the lakes and wetlands in the region today.

Complexly layered stratigraphy includes fine-grained glacial till (pebbly clay loam), lacustrine (silt

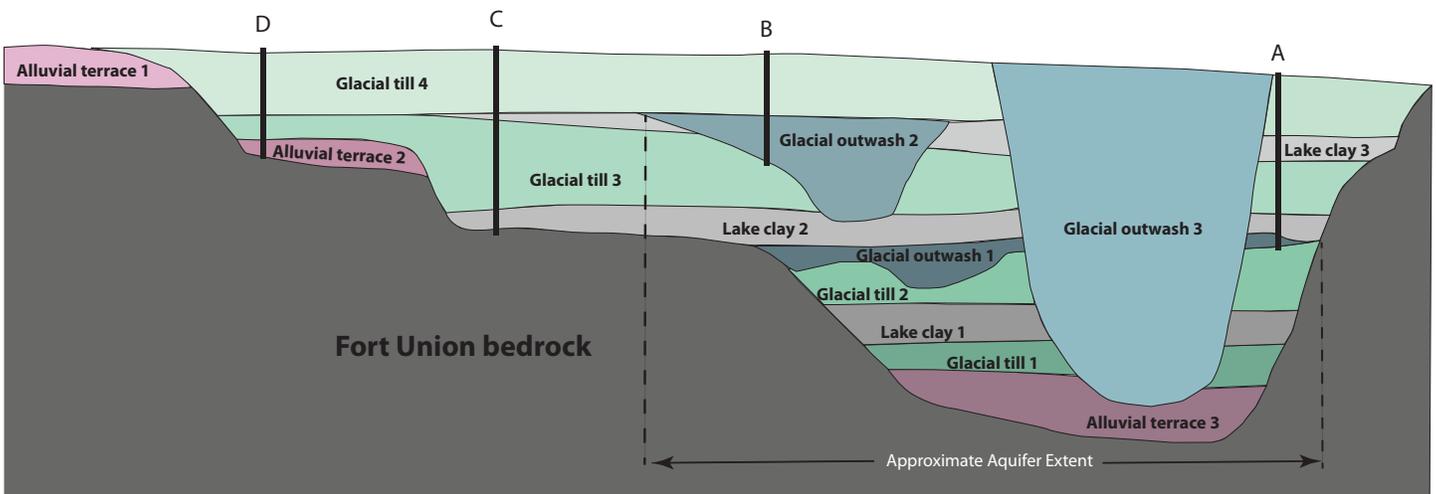


Figure 11. A schematic of glacial stratigraphy typical for the Medicine Lake area illustrates connectivity between deposits of different ages and heterogeneous sediments over short horizontal and vertical distances.

and clay) deposits, and slack water sediment (fine sand, silt, and clay). Where present, these materials form aquitards that restrict groundwater flow. Lakes and wetlands display a wide range of hydraulic connection to the underlying aquifer. Some are separated from the aquifer by fine-grained materials, others are separated by thick sodium sulfate salt deposits (up to 70 ft thick in Miller Lake near Westby; Murphy, 1996), and others are in good hydraulic connection with the underlying aquifer. The aquitards control the flow of water between the underlying aquifer and surface water in lakes and wetlands.

Groundwater is the primary source of recharge to the lakes, with minor contributions from direct precipitation and overland flow. Donovan (1988) categorized many of the lakes as open-, closed-, or restricted-outflow type lakes based on water quality. Closed or restricted-flow lakes allow very little outflow and have poor water quality because evaporation has increased the mineral concentration in the lake water. Flow in open-outflow lakes is not restricted, which results in better water quality. In either case, evapotranspiration concentrates dissolved minerals in surface water compared to nearby groundwater. In 1990, Reiten (1992) resampled lakes and wetlands that were initially sampled in 1984, before the extended drought of the 1980s. This showed 20 of 22 lakes had increases in SC ranging from 9 percent to 128 percent. In general, dissolved solids increased more in the closed-outflow lakes compared to open-outflow lakes.

Water quality is a good proxy for degree of hydraulic connection between the Clear Lake aquifer and lakes. Lakes with relatively fresh water indicate a good connection to the aquifer. Mallard Pond, an open-flow lake, had a TDS of 1,039 mg/L in 2005. A sample at the same time from a 10-ft-deep shoreline well (161782) completed in outwash had a TDS of 392 mg/L. The TDS of four closed or restricted-flow lakes sampled in 1990 varied from 6,090 mg/L (Brush Lake), 11,700 mg/L (Clear Lake), 28,800 mg/L (SE Goose Lake), and 264,000 mg/L (Horseshoe Lake). The aquifer adjacent to these lakes did not have the same poor water quality. Evapotranspiration causes lake levels to decline, inducing low-TDS groundwater flow to the lake, which improves surface-water quality.

The upper outwash zone on the north side of Medicine Lake has a direct hydraulic connection to the lake as indicated by similar water quality and level. The

hydraulic connection between the lake and the lower alluvial zone appears to be limited by aquitards in both the north and south channels. Confining beds ranging in thickness from 37 to 184 ft are present near Medicine Lake; these aquitards overlie the lower alluvial zone and limit the movement of water. A discussion of the aquifer–lake hydraulic connection of the south channel is covered later in this report (see Hydrogeologic Framework in the section South Medicine Lake Focus Area).

Groundwater from the Clear Lake aquifer discharges to the aquifer underlying Big Muddy Creek. Low-permeability clay sediments ranging from 25 to 150 ft separate Big Muddy Creek from the underlying lower alluvial zone. Monitoring well water levels in this area show little fluctuation, indicating a constant upward gradient towards Big Muddy Creek. As a result, groundwater flow is restricted and relatively constant from the aquifer to the creek. Groundwater seepage helps maintain the baseflow in the creek, but the seepage rate does not appear to be affected by climatic fluctuations or groundwater withdrawals.

### Aquifer Properties

Hydraulic properties of the aquifer, including transmissivity, hydraulic conductivity, and storage coefficients, were compiled from earlier studies and analyses completed during this investigation. Data from several aquifer tests reported by Donovan (1988) and Schuele (1998) demonstrate the range and variability of the hydraulic properties of the Clear Lake aquifer. Tables 1 and 2 summarize these tests from outwash and alluvial deposits, respectively. Figure 12 shows the locations of the aquifer tests reported by Donovan (1988) and Schuele (1998). Estimates of hydraulic properties prepared by the MBMG within the South Medicine Lake focus area are presented below.

Transmissivity ranges from 960 ft<sup>2</sup>/d to 147,000 ft<sup>2</sup>/day in the upper outwash zone, with a geometric mean of 23,800 ft<sup>2</sup>/d (table 1). Hydraulic conductivity ranges from 60 ft/d to 6,280 ft/d, with a geometric mean of 960 ft/d. Donovan (1988) and Schuele (1998) estimated storage coefficients for the six aquifer tests where observation well data were available (table 1). Storage coefficient values indicate confined to unconfined conditions, and range from 0.0003 to 0.05 with an arithmetic mean of 0.015.

Table 1. Aquifer parameters for wells completed in the upper outwash zone of the Clear Lake aquifer.

GWIC ID	Latitude	Longitude	Well Yield (gpm)	Test Duration (h)	Aquifer Thickness (ft)	Transmissivity (ft <sup>2</sup> /day)	Hydraulic Conductivity (ft/day)	Storage Coefficient
3770	48.49775	-104.28107	110	24.0	75	56,400	750	NA
3773	48.55502	-104.18186	85	1.1	16	960	60	NA
3869	48.58408	-104.14289	85	5.5	30	45,900	1,560	NA
3871	48.58553	-104.14540	85	5.8	23	87,000	3,780	NA
3858	48.61384	-104.14441	85	5.8	7	31,900	4,550	NA
3862	48.59823	-104.09945	85	5.0	43	20,600	480	NA
3947	48.68554	-104.14395	85	6.0	20	26,300	1,310	NA
3866	48.60633	-104.04819	80	2.7	17	7,900	460	NA
3868	48.59207	-104.04802	77	7.0	9	23,000	2,560	NA
3767	48.49082	-104.45426	16	4.0	12	5,900	490	NA
3854	48.56920	-104.19780	21	0.0	18	113,000	6,280	NA
3861	48.60401	-104.10028	85	6.0	14	36,100	2,580	0.0003
3870	48.58656	-104.14710	1,000	24.0	61	147,000	2,410	0.0030
46003	48.66799	-104.15805	1,000	1.6	49	39,400	800	0.0500
150614	48.48351	-104.29082	550	9.6	29	3,240	110	0.0010
45327	48.56610	-104.16020	1,000	1.5	25	12,100	480	0.0030
3859	48.59520	-104.13860	960	5.5	94	67,000	710	0.0090
					Mean	42,600 <sup>1</sup>	1,730 <sup>1</sup>	0.0200 <sup>2</sup>
					Median	31,900	800	0.0060
					Minimum	960	60	0.0003
					Maximum	147,000	6,280	0.0500
					Geometric Mean	23,800	960	0.0050

<sup>1</sup>Geometric mean.<sup>2</sup>Arithmetic mean.

Note. Data compiled from Donovan (1988) and Schuele (1998).

Table 2. Aquifer properties for wells completed in the lower alluvial zone of the Clear Lake aquifer.

GWIC ID	Latitude	Longitude	Well Yield (gpm)	Test Duration (h)	Aquifer Thickness (ft)	Transmissivity (ft <sup>2</sup> /day)	Hydraulic Conductivity (ft/day)	Storage Coefficient
3867	48.59160	-104.04800	52	6	9.8	2,800	280	NA
3779	48.49806	-104.17297	47	6	36.1	16,400	460	NA
3766	48.49082	-104.45425	28	5	13.1	1,300	100	NA
3677	48.38955	-104.45278	50	4	75.5	5,380	70	NA
155931	48.48496	-104.27037	Variable	2	29.5	2,160	70	0.0029
43116	48.40409	-104.43808	550	6	13.1	4,250	320	NA
150615	48.48481	-104.26836	Variable	5.3	29.5	5,580	190	0.0001
3769	48.49788	-104.28109	Variable	1.6	29.5	6,930	240	0.0003
142105	48.56724	-104.19445	1,250	74	52.5	6,000	110	0.0009
45315	48.59520	-104.13860	1,000	1.6	19.7	8,840	450	0.0002
					Mean	5,970 <sup>1</sup>	230 <sup>1</sup>	0.0009 <sup>2</sup>
					Median	5,480	210	0.0003
					Min	1,310	70	0.0001
					Max	16,430	460	0.0029
					Geometric Mean	4,770	180	0.0004

<sup>1</sup>Geometric mean.<sup>2</sup>Arithmetic mean.

Note. Data compiled from Donovan (1988) and Schuele (1998).

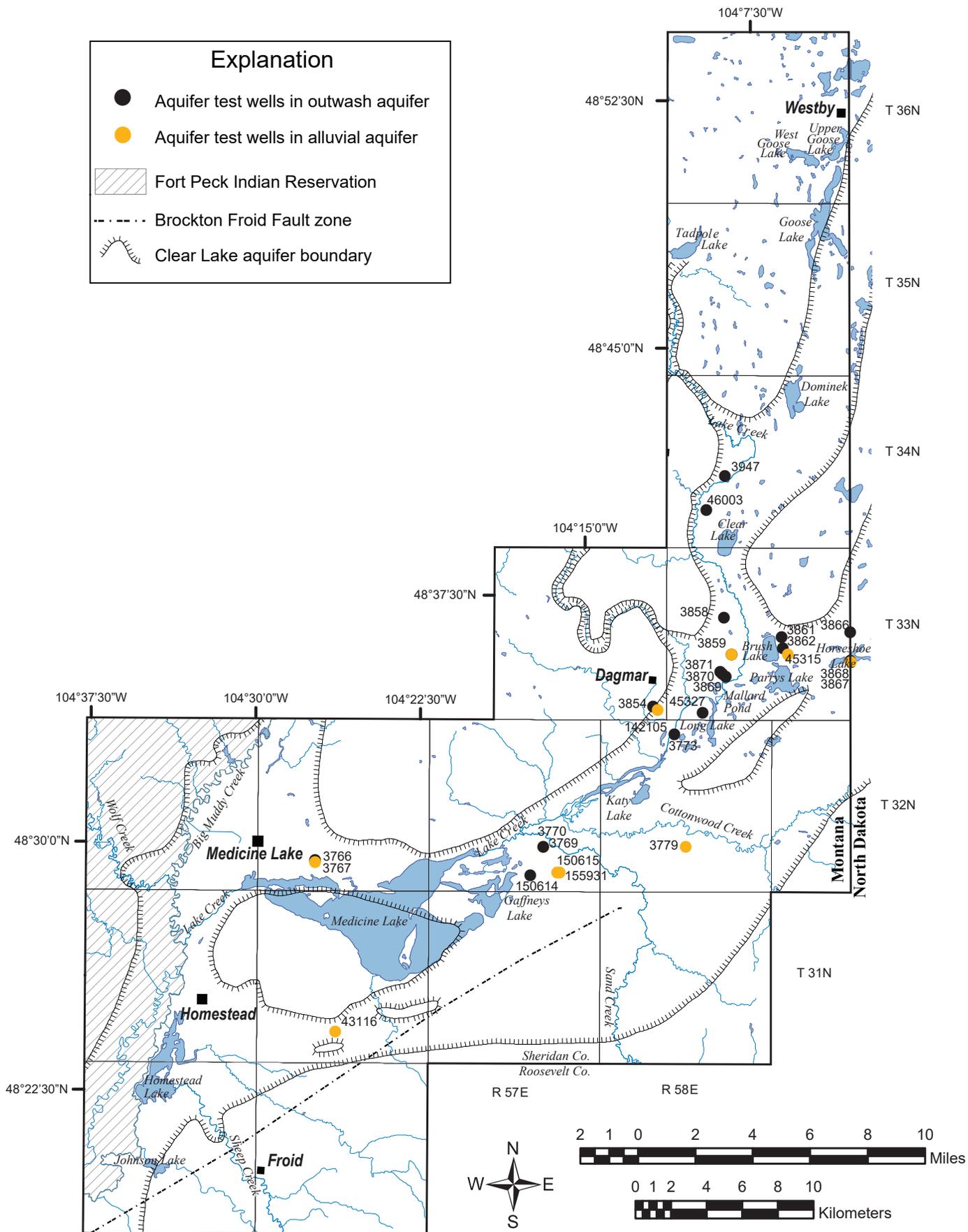


Figure 12. Locations of aquifer tests listed in tables 1 and 2.

Testing of the lower alluvial zone of the Clear Lake aquifer indicates transmissivity ranges from 1,300 ft<sup>2</sup>/day to 16,400 ft<sup>2</sup>/d, with a geometric mean of 4,770 ft<sup>2</sup>/d (see table 2). Hydraulic conductivity ranges from 70 ft/d to 460 ft/d, with a geometric mean of 190 ft/d. Storage coefficients were calculated for 5 of the 10 aquifer tests and ranged from 0.0001 to 0.0003; the arithmetic mean was 0.00089. These data indicate that water in the alluvial portion of the Clear Lake aquifer is under confined to leaky confined conditions.

Although the ranges of these properties overlap for the upper outwash zone and the lower alluvial zone of the Clear Lake aquifer, the medians show significant differences (fig. 13). The lower alluvial zone of the Clear Lake aquifer is generally more confined, with hydraulic conductivity values about an order of magnitude less than the upper outwash zone. As a result, irrigation wells constructed in the upper outwash zone of the Clear Lake aquifer are typically more produc-

tive and produce less drawdown than wells completed in the lower alluvial zone.

### Groundwater Flow

Water levels measured during April and May 2015 were used to create a potentiometric surface of the Clear Lake aquifer (fig. 14). Groundwater flows from northeast to southwest from the North Dakota border near Westby towards Dagmar. Groundwater gradients increase from 0.0007 north of the Clear Lake constriction to 0.002 at the constriction and 0.001 below, near Dagmar.

Groundwater flow in the Clear Lake aquifer to the south of Medicine Lake is generally to the west. A groundwater mound underlies the sand hills (fig. 14) southeast of Medicine Lake. The high point of the mound forms a groundwater divide, with part of the flow to the north at a steep gradient (0.008) until it joins the westerly flow of the northern channel. Southwest of this groundwater divide, flow is westerly

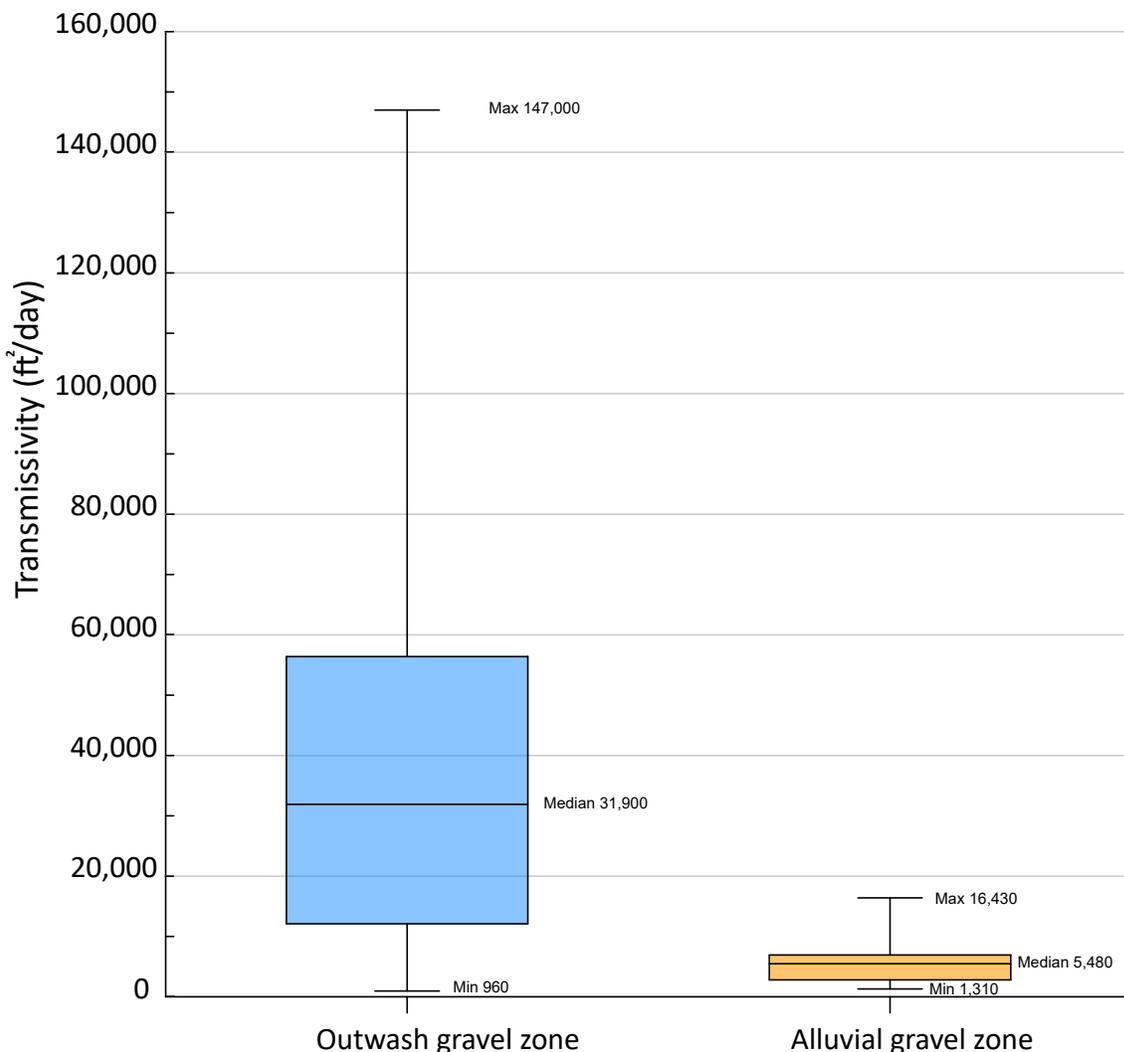


Figure 13. The transmissivity of the upper outwash zone is significantly higher than that of the lower alluvial zone in the Clear Lake aquifer ( $p < 0.01$ , two-tail  $t$ -test assuming unequal variances).

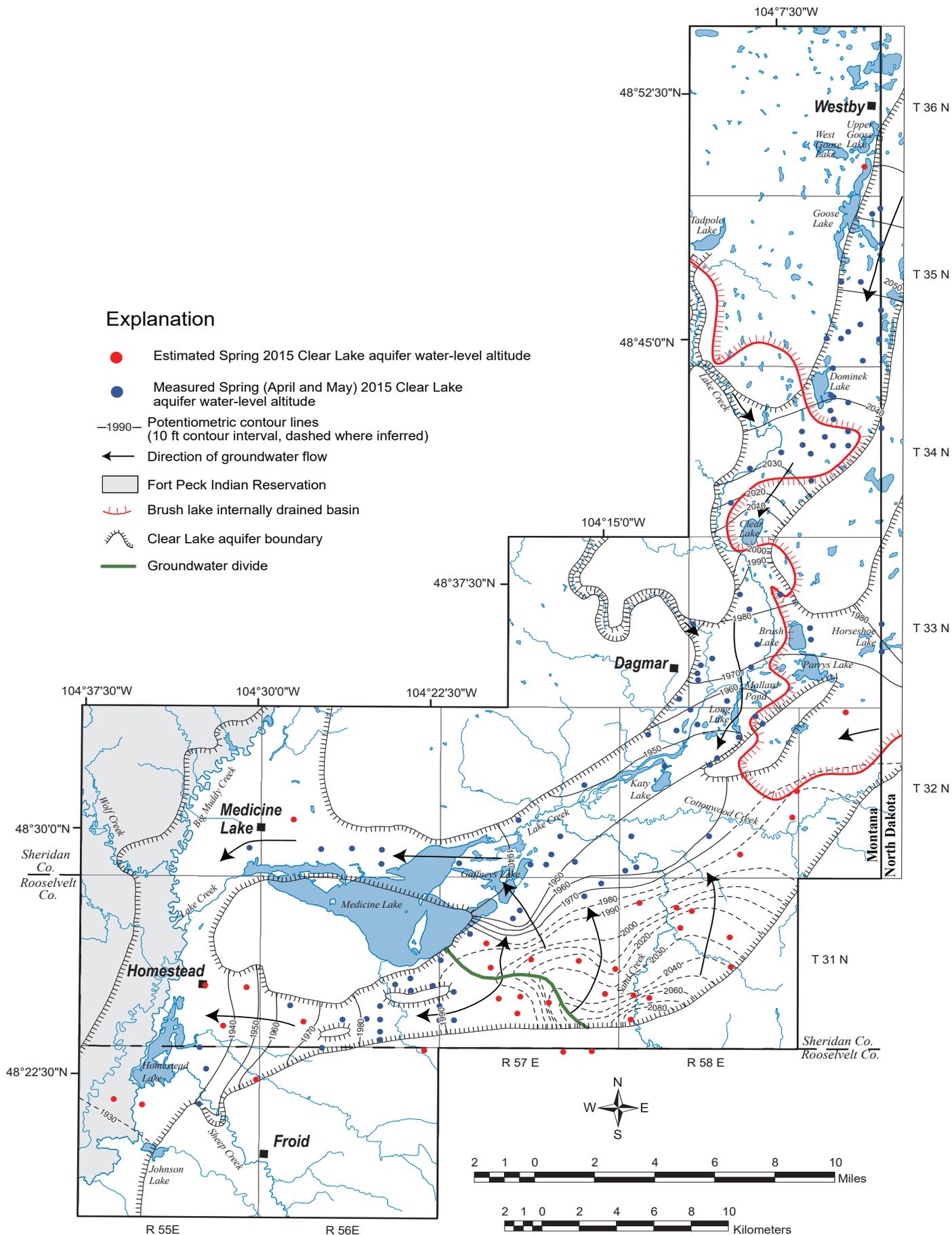


Figure 14. Potentiometric surface of the Clear Lake aquifer.

towards the Big Muddy Creek Valley. Relatively low hydraulic gradients (0.0007) extend throughout most of the South Medicine Lake area. Hydraulic gradients steepen to about 0.002 between this area and undeveloped portions of the aquifer underlying the Big Muddy Creek Valley.

In the center of the East Medicine Lake area (location shown in fig. 3), the upper outwash zone and lower alluvial zone are in direct hydraulic connection and form an area of high transmissivity (Donovan, 1988). Even within the highly transmissive area, vertical hydraulic gradients are upward, with heads in wells completed in the lower alluvial zone typically a few feet higher than wells completed in the upper outwash zone. Lateral gradients range from 0.0002 to 0.0001 from the East Medicine Lake area to Big Muddy Creek.

### Water Use

Figure 15 summarizes the estimated water use for irrigation from the Clear Lake aquifer. Because of the annual variation in crop moisture demands and precipitation, the annual water use fluctuates widely. Since 1994, water use has ranged from about 3,000 acre-ft to 6,000 acre-ft. High pumping costs promote water conservation, and producers limit irrigation to supplement normal growing season precipitation.

Water-use estimates depend on accurate pumping-rate measurements. Figure 16 compares pumping rates measured with an ultrasonic flowmeter to original reported rates. The original pumping rates are from a variety of measurement methods, including in-line flowmeters, well-driller reports, and designed rates for sprinkler-head nozzles. The original reported pumping rate was greater in 28 of the 37 well systems measured with the ultrasonic flowmeter in 2014 and 2015. The measured rate was lower in all of the wells with reported rates greater than 950 gpm. The 9 wells with higher reported rates tended to show greater differences between reported and measured values. Possible causes of the discrepancies are: sand degrading the impellers of the flowmeters, flowmeters breaking, decline in pump efficiency, or incorrect estimates of the original pumping rate. Current measurements of water use are more accurate than past estimates because of these improved measurement methods.

Water use in 2015 ranked 13th highest out of the 36 years recorded (fig. 15). The total volume pumped was 4,087 acre-ft, which is 352 acre-ft greater than the long-term average. Over this period, the mean (3,735 acre-ft) and median (3,758 acre-ft) were similar, indicating that annual water use is normally distributed over time.

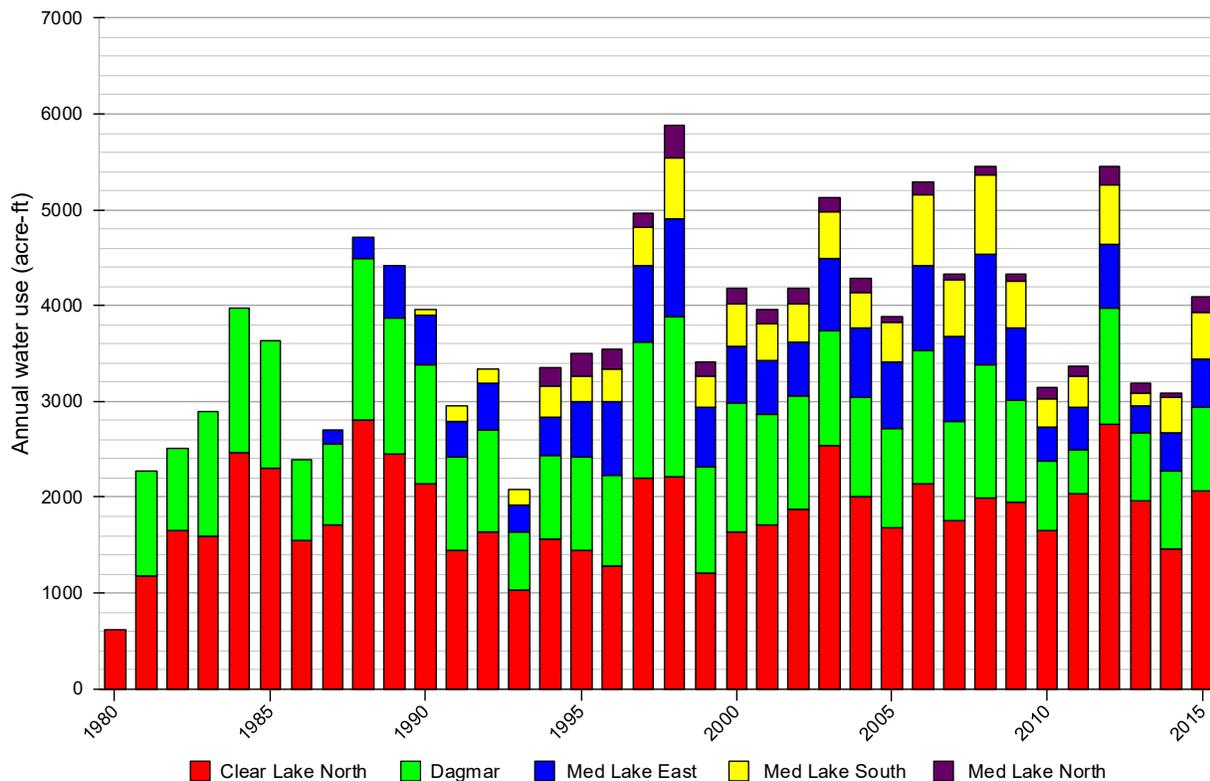


Figure 15. Estimated annual irrigation water use (acre-ft) from the Clear Lake aquifer by management area.

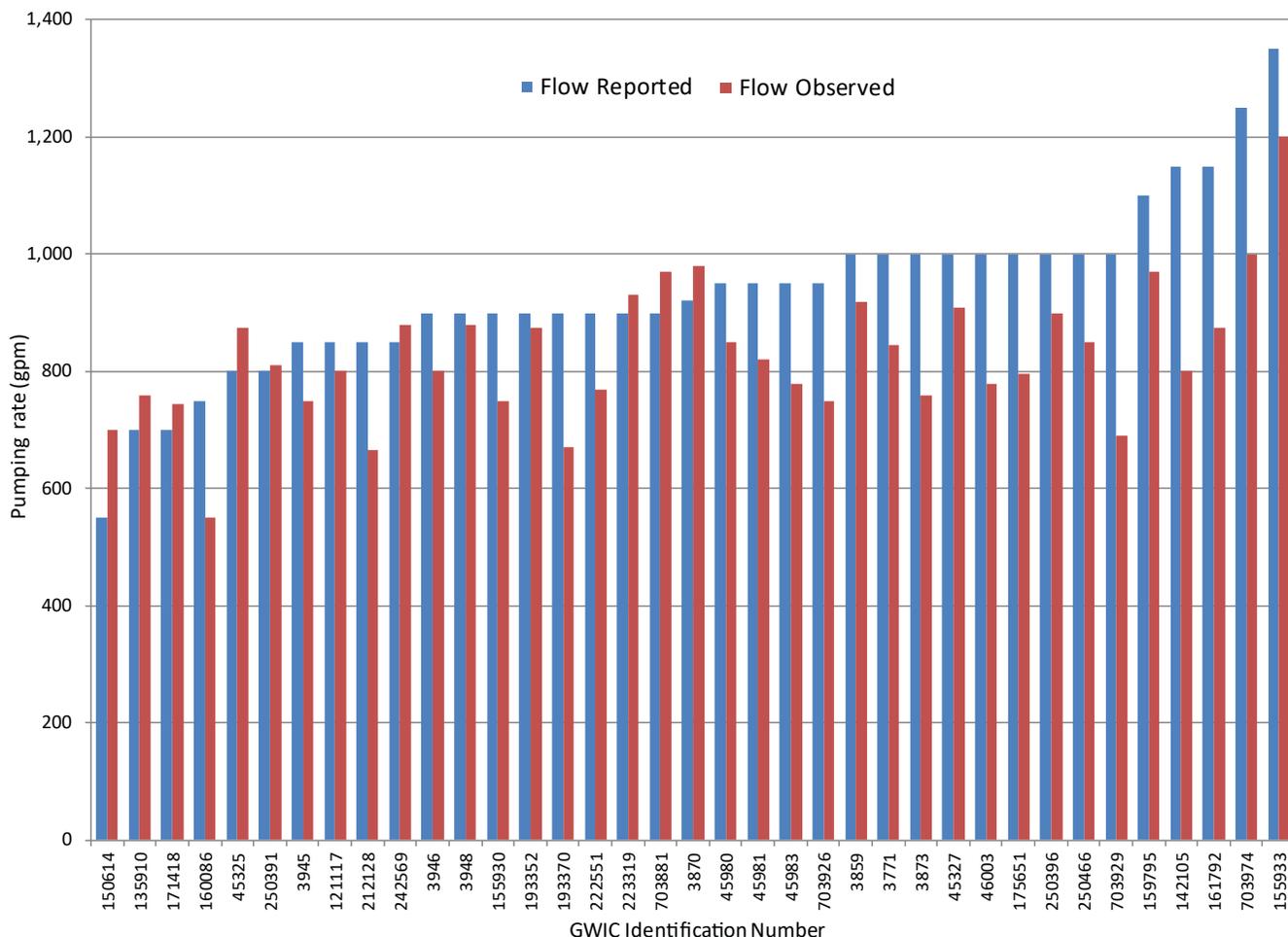


Figure 16. Comparison of reported pumping rates to those measured using an ultrasonic flowmeter shows that most of the wells pump at rates lower than reported.

### Groundwater Response to Climate and Pumping

Prior to this project, the primary tools for assessing irrigation impacts were observations of climate, water-use, and water-level changes. Water levels in monitoring wells near irrigation wells decline in response to irrigation withdrawals. These declines are greatest in tightly confined parts of the lower alluvial zone. In this setting, seasonal water-level declines can extend several miles from a pumping well. The water-level response is lower at monitoring wells in unconfined or leaky confined parts of the aquifer; often the response is limited to less than  $\frac{1}{2}$  mi from the pumped well. Following the irrigation season, water levels recover, approaching the original static water level before the next irrigation season. Long-term water-level trends follow precipitation trends, with declines apparent during the extended drought of the 1980s. As illustrated below, aquifer water levels have now recovered to those prior to irrigation development.

Representative hydrographs in each of the five management areas illustrate the influence of precipita-

tion and irrigation on groundwater levels (plate 3). A more detailed discussion of the management areas is presented in Reiten (2002). Hydrographs shown on plate 3 demonstrate the variability of water-level fluctuations in the Clear Lake aquifer. Wells completed in the lower alluvial zone of the Clear Lake aquifer typically have greater fluctuations than those completed in the upper outwash zone of the aquifer.

Moving downgradient from the northeast, near Westby, the hydrographs at wells 149610, 3944, and 3947 show similar timing of water-level fluctuations, although their magnitudes vary (plate 3). Water levels at these wells declined about 5–7 ft from 1982 to 1993 during the 1980s drought. The rate of decline slowed, responding to higher precipitation and recharge in the 1990s and 2000s. Above-normal snowmelt and precipitation in 2009 and 2011 resulted in abrupt water-level rises.

The drought response in the 1980s was not as severe in the southern part of the study area. In the Dagmar and East Medicine Lake Management areas, the

water-level decline was 2–4 ft, although the decrease was partially masked by the annual drawdown and recovery caused by irrigation. Water-level fluctuations at wells 3871, 45292, 3779, 3769, and 3770 show this pattern. The response at well 3871 (completed in the upper outwash zone) ranges from 2 to 5 ft and is attributed to seasonal irrigation drawdown. In contrast, the water-level response at well 45292 (completed in the lower alluvial zone) ranges from 10 to 20 ft.

Although proximity to a production well influences patterns of drawdown and recovery, so do differences in transmissivity and storage in the glacial outwash compared to the alluvium. Similar responses are shown at monitoring wells 3769 (lower alluvial zone) and 3770 (upper outwash zone). These wells are next to each other in the East Medicine Lake area and both respond to irrigation pumping from wells completed in the alluvial and outwash zones. Monitoring well 3779, 3 mi from the nearest irrigation well pumping from the deep alluvial gravel, shows a typical confined aquifer response.

The South Medicine Lake area has some of the largest water-level fluctuations observed in the Clear Lake aquifer. The annual water-level fluctuations at well 150940 range from 10 to 40 ft. High water-use years produce a larger fluctuation. The large response at this well is attributed to its proximity to several irrigation wells and to the hydraulic boundary located  $\frac{1}{4}$  to  $\frac{1}{2}$  mi north.

The only well in the South Medicine Lake Management area with long-term monitoring is 3677. This well is in the lower alluvial gravel and it displays an unusual water-level response compared to other wells in this area. It is about 100 ft deeper than the other wells and appears to be in a deepened region of lower alluvial zone. Based on water-level responses and water chemistry, it also appears hydraulically separated from the irrigation wells by a longitudinal flow barrier. The subdued annual drawdown and recovery at this site are probably caused by cumulative impacts of irrigation withdrawals from distant wells. In contrast, other monitoring wells in this area exhibit an immediate or nearly immediate response to pumping.

### **Regional Groundwater Recharge and Discharge**

This section presents a discussion of the recharge and discharge processes that affect the Clear Lake aquifer, based largely on understanding of the geologic

setting, the location of surface-water features in the landscape, and their role in the hydrologic system. The surficial geologic materials overlying the aquifer greatly influence groundwater recharge rates. Donovan (1988) mapped the highest recharge rates of about 9.3 in/yr in areas of glacial outwash or sand hills (fig. 8). Intermediate recharge rates of about 1.3 in/yr apply to Quaternary alluvium and glacial till (Donovan, 1988). Schuele (1998) estimated that  $\frac{1}{3}$  of the long-term average precipitation recharged the aquifer and the rest was lost to evapotranspiration. The Fort Union Formation aquifers have the potential to provide recharge to the Clear Lake aquifer through interaquifer flow. The hydraulic heads are higher in the Fort Union Formation surrounding the Clear Lake aquifer, suggesting inflow, but the recharge volume it provides is unknown.

The timing and rate of precipitation are not uniform across the landscape. Summer thunderstorms are intermittent, with no rain in one area and a deluge of several inches falling in another area. Snowfall collects in low areas, such as road ditches, while hilltops are blown free of snow. As a result, when the snowpack melts, recharge rates vary across the landscape. As snow accumulates on the prairie, the potential for infiltration depends on the thickness of frozen ground. Thick snow can insulate the ground, reducing the frost depth and extent. As the snow melts, downward infiltration of water has a greater potential to recharge the aquifer in areas with less frozen ground.

The discharge area for the Clear Lake aquifer is the Big Muddy Creek Valley between Homestead and Froid. A large volume of groundwater discharges to the lakes and wetlands as relatively high rates of evapotranspiration from these surface-water bodies drives the hydrologic flow system. Lakes alone cover about 15,075 acres overlying the aquifer (Donovan, 1988). There are several thousand additional acres of wetlands. Estimates of total annual evapotranspiration over the entire Clear Lake aquifer from previous models have ranged from 50,000 acre-ft (Donovan, 1988) to 150,000 acre-ft (Schuele, 1998). Recharge and discharge occur in close proximity where the surface glacial outwash (areas of high recharge rates) are adjacent to lakes and wetlands (areas of high evapotranspiration rates).

### **Water Quality**

Water-quality information was compiled from previous projects and samples collected as part of this

project. The water quality of the Clear Lake aquifer is important for irrigation development, since the areas with high SAR water are not suitable for irrigation. Water quality was also examined to evaluate groundwater–surface water interactions and presence or absence of brine water co-produced with oil and gas development.

The water quality is highly variable in different zones of the Clear Lake aquifer and associated surface-water bodies. Table 3 summarizes analyses of water-quality samples in surface water, Clear Lake aquifer, Fort Union aquifer, the Fox Hills–Hell Creek aquifer, and Mesozoic (for example, Dakota Formation) and Paleozoic (Madison Group) reservoirs associated with hydrocarbon extraction. Appendix B lists all of the samples and water-quality data. Piper diagrams were constructed from 369 water samples.

Oil and gas exploration and development have targeted Paleozoic hydrocarbon reservoirs underlying most of northeast Montana. As a byproduct from

hydrocarbon development, high-salinity wastewater is produced and then injected into Mesozoic reservoirs for disposal. Surface-water and groundwater influenced by this co-produced water (brines) have elevated chloride concentrations and fall in the yellow shaded area on the Piper plots (fig. 17). Samples from such sources potentially have contamination index (CI) values greater than 0.035. Reiten and Tischmak (1993) developed the CI as a screening tool for identifying oilfield contamination. The CI is the ratio of chloride concentration (mg/L) to specific conductance (SC,  $\mu\text{S}/\text{cm}$ ). A value of 0.035 was empirically established as a lower limit of brine contamination.

Seventy-five surface-water samples were collected from lakes, streams, and wetlands overlying the Clear Lake aquifer. The lakes and wetlands are typically less than 10 ft deep. One exception is Brush Lake (east of Dagmar, shown in fig. 3), which reaches depths greater than 60 ft. The streams are slow moving and are perennial to intermittent. Lakes and streams along

the Lake Creek drainage are controlled by headgates allowing managers to divert water when desired to enhance wildlife habitat. A diversion canal moves water from Big Muddy Creek to Medicine Lake for lake-level control. High rates of evapotranspiration in the summer concentrate the salts dissolved in surface-water bodies. Magnesium, sodium, bicarbonate plus carbonate, and sulfate dominate the dissolved constituents found in the surface water. Six samples are dominated by chloride anions, which we attribute to brine contamination. Chloride concentrations in 20 samples exceeded the lower CI limit of 0.035.

The upper outwash zone ranges from near-surface sand and gravel deposits to depths greater than 200 ft. A total of

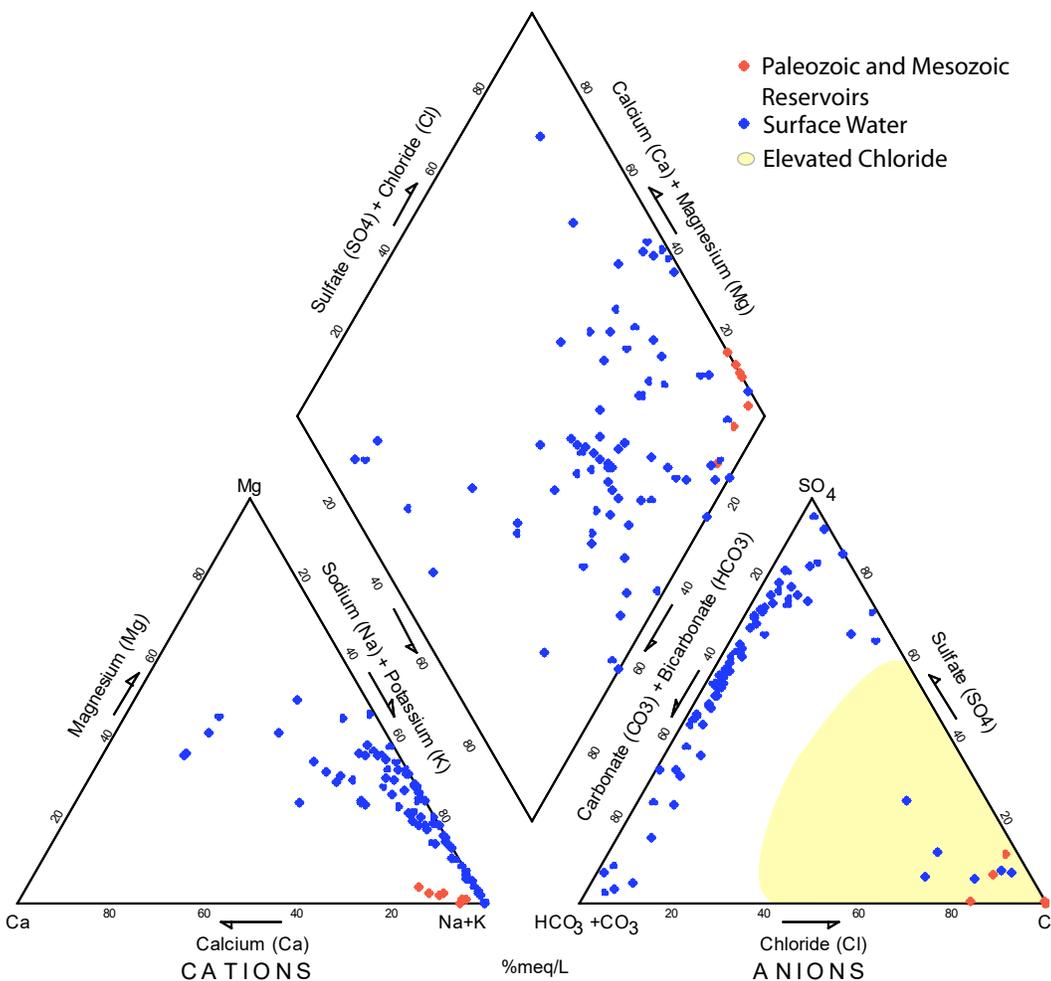


Figure 17. A Piper diagram of the surface-water samples and the samples from Mesozoic and Paleozoic reservoirs shows that elevated chloride (yellow shading) can be used to identify water impacted by brine contamination.

Table 3. Summary of inorganic water quality from groundwater and surface water associated with the Clear Lake aquifer.

Lab	pH	Lab SC	Ca (mg/L)	Mg (mg/L)	Na (mg/L)	K (mg/L)	Fe (mg/L)	Mn (mg/L)	SiO <sub>2</sub> (mg/L)	HCO <sub>3</sub> (mg/L)	CO <sub>2</sub> (mg/L)	SO <sub>4</sub> (mg/L)	Cl (mg/L)	F (mg/L)	TDS (mg/L)	SAR
<b>Surface Water</b>																
<i>n</i> = 75 samples																
Reported and Detections	70	69	75	75.0	75.0	75.0	42.0	49.0	61.0	75.0	64.0	75.0	75.0	55.0	75	75.0
Maximum	10.09	174,957	699.0	1,940.0	108,000.0	1,050.0	1.0	2.7	33.7	71,000.0	54,800.0	93,300.0	10,500.0	10.5	282,003	4,229.4
Minimum	7.23	327	1.6	11.7	12.1	4.0	0.0	0.0	0.2	68.1	1.5	6.6	5.5	0.0	204	0.3
Average	8.92	14,786	70.8	299.5	5,848.9	136.9	0.2	0.1	8.6	2,574.0	1,493.0	8,379.8	888.0	0.5	18,174	130.4
Median	9.05	5,210	22.9	165.5	1,134.0	47.8	0.1	0.0	6.9	780.2	171.9	1,820.0	90.6	0.2	4,162	15.6
Standard Deviation	0.69	26,425	119.8	363.6	15,407.7	201.6	0.2	0.4	7.3	8,323.5	6,885.2	18,168.8	1,850.1	1.5	42,482	543.1
<b>Clear Lake Aquifer Outwash</b>																
<i>n</i> = 116 samples																
Reported and Detections	116	116	116	116	116	116	109	101	101	115	3	116	116	96	116	116
Maximum	8.50	33,100	1,663.0	857.0	5,000.0	32.0	13.6	4.1	31.0	1,525.5	21.6	1,147.0	12,770.0	1.2	20,879.0	196.4
Minimum	6.74	509	0.7	0.3	2.1	0.9	0.0	0.0	9.2	98.2	2.0	9.0	1.0	0.1	284.6	0.1
Average	7.57	1,927	123.6	60.3	226.4	5.7	2.6	0.5	23.9	650.6	9.9	337.0	133.4	0.3	1,229.5	5.7
Median	7.51	1,434	104.8	48.8	135.6	5.3	2.5	0.3	25.0	612.5	6.0	288.5	11.0	0.2	1,031.5	2.6
Standard Deviation	0.34	3,105	151.1	79.1	467.3	3.0	2.3	0.7	4.1	268.6	8.4	243.9	1,179.1	0.2	1,889.7	18.2
<b>Clear Lake Aquifer Alluvium</b>																
<i>n</i> = 54 samples																
Reported and Detections	54	54	54.0	54.0	54.0	54.0	48.0	48.0	49.0	53.0	5.0	52.0	53.0	47.0	54.0	54.0
Maximum	8.73	2,999	204.0	129.1	678.0	15.9	6.1	2.8	33.2	1,683.4	1,233.0	666.2	75.5	3.4	2,030.3	62.0
Minimum	7.08	845	3.4	2.0	24.4	2.3	0.1	0.0	10.4	513.0	2.0	0.4	1.2	0.1	430.7	0.5
Average	7.84	1,670	98.5	48.1	245.5	5.8	2.0	0.3	24.9	848.1	263.2	287.0	11.3	0.7	1,145.1	6.8
Median	7.85	1,692	102.1	44.5	175.9	5.5	1.5	0.2	25.7	775.9	19.2	251.4	7.0	0.5	1,121.6	3.7
Standard Deviation	0.32	516	47.1	21.5	164.6	2.0	1.7	0.4	4.5	281.1	485.3	186.9	14.4	0.6	395.3	9.0
<b>Fort Union Aquifer</b>																
<i>n</i> = 16 samples																
Reported and Detections	16	16	16	16	16	16	14	13	15	15	3	16	16	16	16	16
Maximum	9.30	3,839	333.0	191.0	1,040.0	10.1	3.7	3.5	25.3	1,430.7	27.8	1,604.0	103.7	5.3	2,647.9	84.4
Minimum	7.12	1,030	2.0	0.9	54.0	2.0	0.1	0.0	6.4	497.0	20.2	1.2	0.8	0.1	670.8	0.9
Average	7.94	2,234	82.7	49.9	401.9	5.1	1.0	0.4	14.5	869.4	22.9	504.2	18.4	1.4	1,484.7	28.2
Median	7.82	2,044	49.2	36.3	334.8	5.2	0.3	0.0	13.5	749.0	20.6	454.0	6.3	0.4	1,342.6	8.6
Standard Deviation	0.54	789	91.2	52.7	283.4	2.4	1.2	0.9	6.9	298.2	3.5	391.0	26.0	1.6	542.0	30.4
<b>Fox Hills–Hell Creek Aquifer</b>																
<i>n</i> = 2 samples																
Reported and Detections	2	2	2.0	2.0	2.0	2.0	1.0	1.0	2.0	2.0	2.0	1.0	2.0	2.0	2.0	2.0
Maximum	8.47	3,699	4.1	1.0	865.2	2.7	0.2	0.3	10.7	1,106.1	30.5	2.6	766.6	2.7	2,170.7	109.7
Minimum	8.21	3,340	3.1	0.9	804.5	1.9	0.2	0.3	9.3	994.4	12.1	2.6	594.3	2.3	1,975.8	94.4
<b>Mesozoic and Paleozoic Reservoirs</b>																
<i>n</i> = 4 samples																
Reported and Detections	4	0	4	4	4	3	1	0	0	4	0	4	4	0	4	4
Maximum	7.80	na	13,800	2,760	110,800	4,200	26	na	na	2,560	na	780	202,950	na	330,965	252.9
Minimum	5.20	na	170	36	5,700	100	26	na	na	12	na	68	7,800	na	15,135	103.6
Average	6.44	na	8,065	1,405	77,550	2,190	26	na	na	727	na	447	138,163	na	227,731	208.0
Median	6.37	na	9,145	1,413	96,850	2,270	26	na	na	168	na	470	170,950	na	282,412	237.8
Standard Deviation	0.94	na	5,335	972	42,892	1,675	0	na	na	1,061	na	283	79,104	na	128,469	61.2

116 groundwater samples from the outwash were analyzed. Ions of calcium, magnesium, bicarbonate, and sulfate dominate the dissolved constituents. The cations range from nearly equal parts of calcium and magnesium to water high in sodium. The anions are dominated by bicarbonate plus carbonate, with a few samples high in sulfate. Four samples are dominated by chloride anions (fig. 18). We attribute the increased chloride concentrations and elevated CI to leaks and spills from facilities related to hydrocarbon development. Groundwater in the relatively shallow outwash zone is more vulnerable to near-surface brine releases than groundwater in the deeper alluvial zone. Chloride concentrations in 15 samples showed CI values greater than 0.035.

The upper outwash zone has good water quality over most of the Clear Lake aquifer, with the best-quality water from wells completed in relatively shallow outwash. Commonly, these wells are close

to areas with surface deposits of sand and gravel, where precipitation recharge is likely. Multiple permeable layers in the upper outwash zone are isolated by confining layers of till or lake clay. Water in the deeper outwash commonly has higher SAR and SC than in the shallow. This water-quality stratification can pose a problem when developing irrigation supplies. Situations arise where potential pumping rates of the deeper, poor-quality water are high enough for irrigation development while the shallower, better-quality water does not produce enough water to support irrigation.

Most of the lower alluvial zone water samples are from wells completed in deep deposits, but samples from upper alluvial terrace aquifers are also included in this group (fig. 11). Fifty-four water samples from alluvial gravels were analyzed. Sodium, calcium, bicarbonate plus carbonate, and sulfate ions dominate the dissolved constituents associated with the alluvium. The cations show water dominated by calcium plus lesser

amounts of magnesium to water dominated by sodium. Anions are dominated by bicarbonate and carbonate (fig. 18). One sample slightly exceeded the lower brine CI impact limit.

The inflow from bedrock often has poorer quality than water in the upper outwash or lower alluvial zones. Mixing of water from Fort Union sandstone and coal aquifers can significantly degrade water quality of the Clear Lake aquifer, primarily by increasing in the sodium concentration. This inflow can elevate the sodium adsorption ratios to levels above irrigation standards.

Wells completed in the Fort Union range in depth from less than 100 ft below the land surface to 350 ft. Sixteen samples from the Fort Union aquifer were analyzed; they showed a

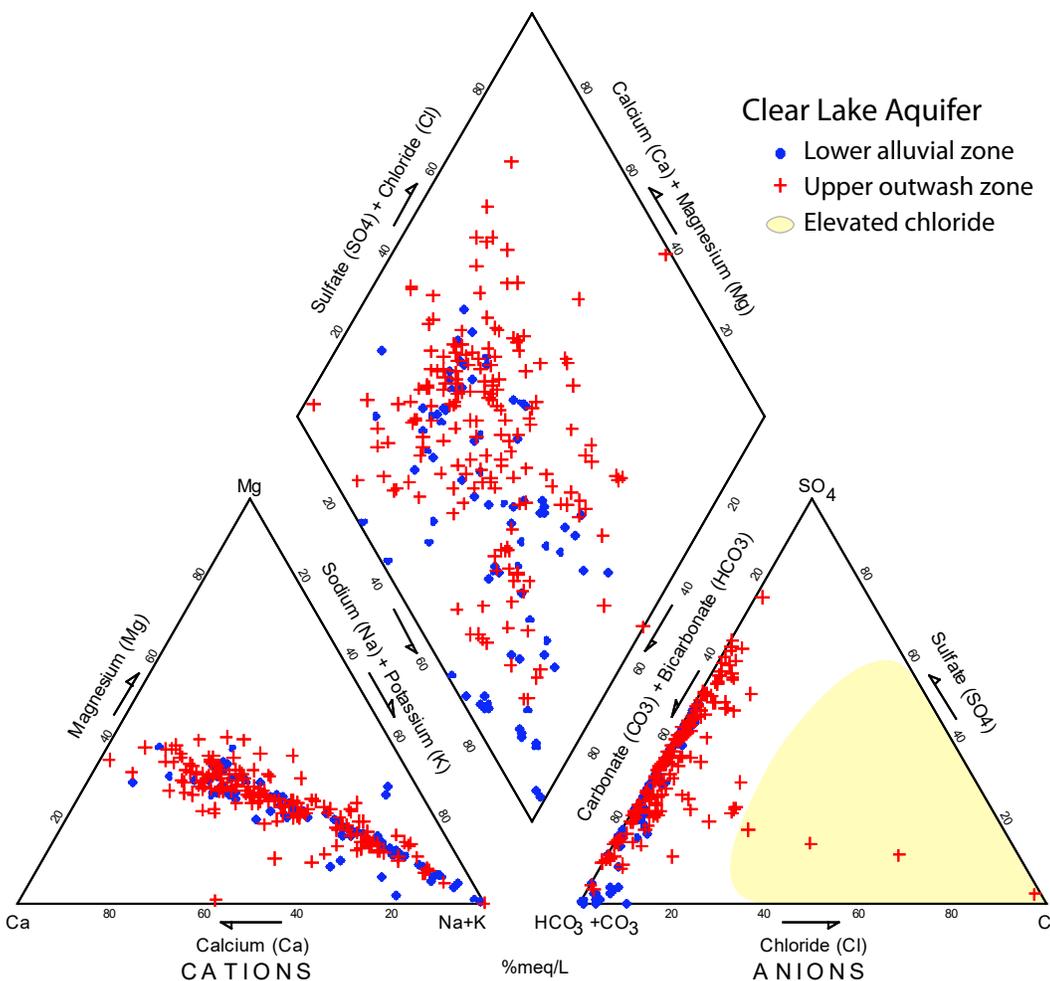


Figure 18. A Piper diagram of the Clear Lake aquifer water samples shows that there is more variability in the quality of the upper outwash zone than in the lower alluvial zone. Samples with elevated chloride (yellow shading) levels appear to be impacted by coproduced brine water from oil production.

wide range in water qualities (see fig. 19, table 3). Cations range from a calcium–magnesium mix to sodium, and anions range from sulfate-rich, to bicarbonate–carbonate. Chloride was uncommon in the Fort Union samples. The CI slightly exceeded the lower brine impact limit in only one sample. It is unlikely that brine contamination caused this exceedance.

Deeper bedrock aquifers have the potential to degrade the water quality of the Clear Lake aquifer where groundwater discharges from bedrock to the alluvial and outwash deposits. The Fox Hills–Hell Creek aquifer is about 800 to 1,200 ft deep in this area. Because of its depth, the Fox Hills–Hell Creek aquifer is not widely developed; only two wells near the study area are completed in this aquifer. Four water samples were analyzed from these two wells. Ions of sodium, bicarbonate, and chloride dominate the dissolved constituents (fig. 19). The elevated chloride in samples from the Fox Hills–Hell Creek aquifer causes a mod-

erately high CI. This results from naturally occurring chloride concentrations and a relatively low SC, not because of brine contamination. Fox Hills–Hell Creek aquifer water discharging from flowing wells can give the appearance of brine contamination where it discharges to surface water or shallow fresh water aquifers.

Groundwater from formations deeper than the Fox Hills–Hell Creek is not used for irrigation or stock water in this area, but is produced as wastewater during oil and gas production. Groundwater from Mesozoic and Paleozoic reservoirs associated with hydrocarbon development is high in sodium and chloride ions. The average concentration based on four water samples was 77,550 mg/L for sodium and 138,000 mg/L for chloride (table 3). Figure 19 shows the contrast in major ion chemistry between the deep Mesozoic and Paleozoic reservoirs, the Fox Hills–Hell Creek aquifer, and Fort Union aquifers. Samples from the Mesozoic and Paleozoic reservoirs cluster near the maximum of the trilinear axes for sodium and chloride.

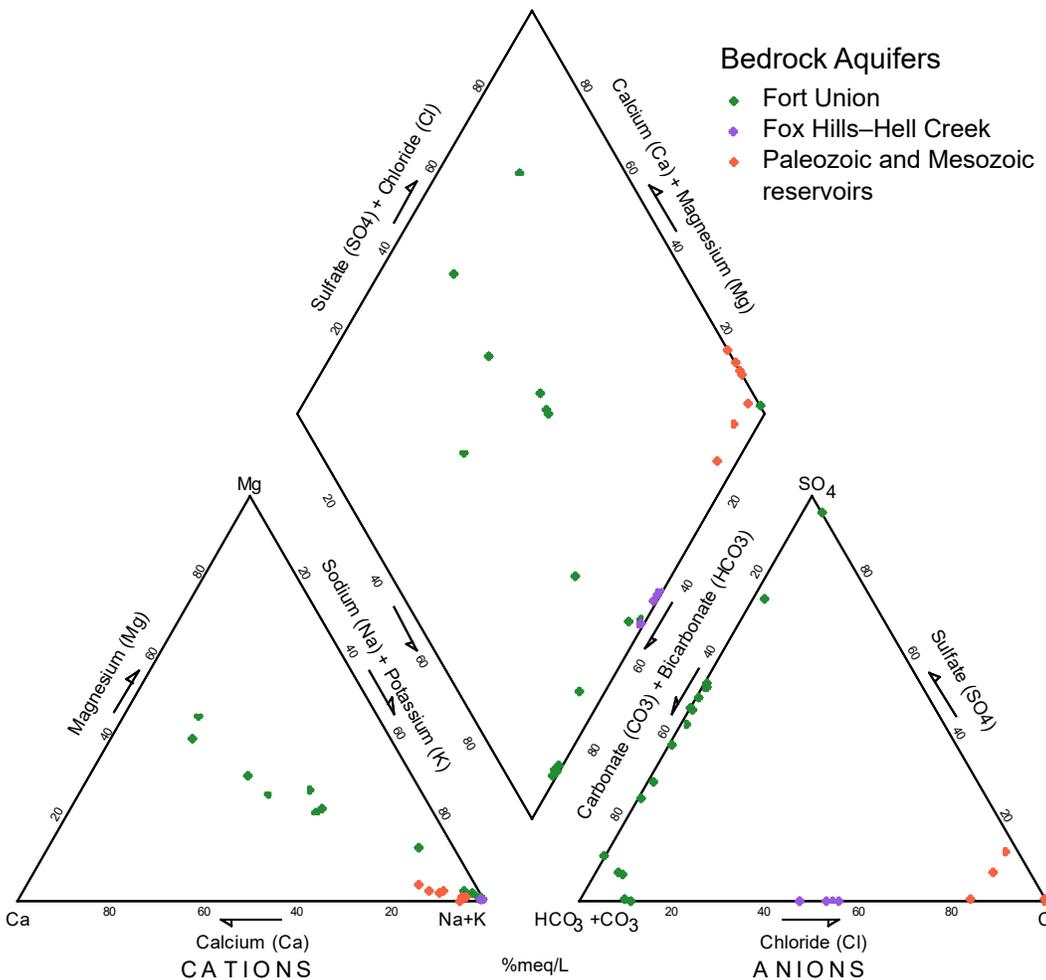


Figure 19. A Piper diagram of the bedrock aquifer samples shows the Fort Union Formation aquifers have the greatest variability in water quality. The Fox Hills–Hell Creek samples show elevated chloride (yellow shading), but are distinct from the Mesozoic and Paleozoic reservoir samples, which are strongly sodium chloride brine waters.

cluster near the maximum of the trilinear axes for sodium and chloride.

Poor water quality in parts of the Clear Lake aquifer can be a significant problem for irrigators; therefore the usability of groundwater for irrigation is often characterized by comparing the SC and SAR of the water. The SAR is the ratio of calcium plus magnesium to sodium with concentrations measured in milliequivalents per liter:

$$SAR = [Na] / \sqrt{\frac{[Ca] + [Mg]}{2}}$$

A plot of the SC versus the SAR (fig. 20) describes irrigation hazard potential (USDA, 1954). The SC increases with dissolved solids in the water and the SAR increases with

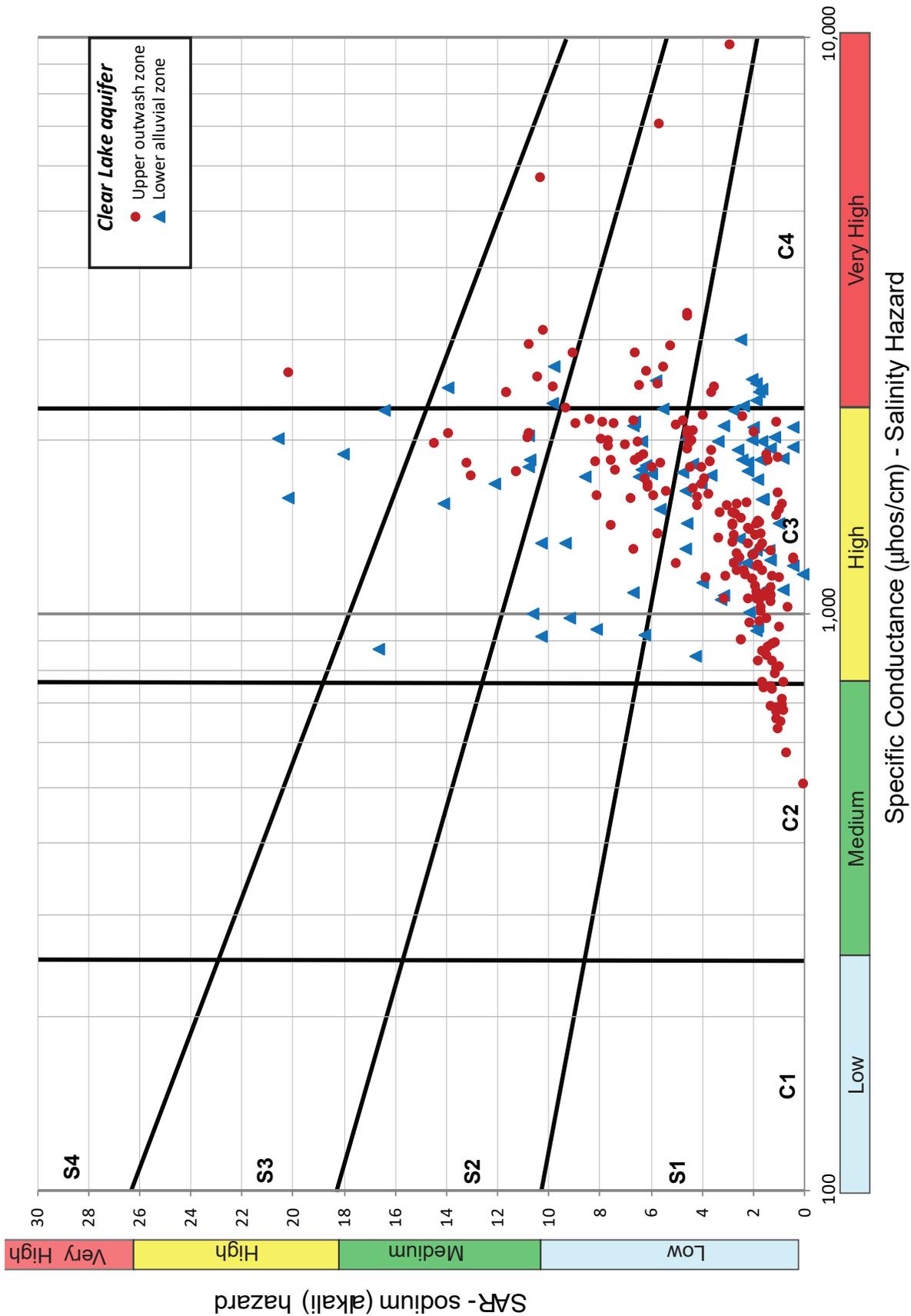


Figure 20. Plot of water samples from the Clear Lake aquifer using the USDA irrigation-water classification diagram (USDA, 1954).

an increase in sodium relative to the alkaline earth cations (calcium and magnesium). The groundwater SC commonly encountered in this area ranges from 750 to 2,250 micro-Siemens/cm ( $\mu\text{S}/\text{cm}$ ). In this SC range, water poses a high-sodium (alkali) hazard at SAR greater than about 10–13; a medium hazard at SAR ranging from 4 to 10; and a low hazard below SAR of 4. In northeastern Montana, high-salinity hazards with SAR greater than 13 resulted in soil damage and abandonment of irrigation systems.

Figure 21 depicts areas of good and marginal water quality based on SAR in the upper outwash zone (fig. 21A) and lower alluvial zone (fig. 21B). The lower alluvial zone is more likely to have poor water quality based on SAR (fig. 21A). The buried alluvial valleys incise deeper into Fort Union bedrock than buried glacial outwash channels, and the buried alluvial valleys are more likely to intercept discharge from saturated coal and sandstone beds of the Fort Union aquifers (Seyoum and Eckstein, 2014).

As the high SC and SAR groundwater from the Fort Union Formation discharges into the buried valley aquifer, the bedrock groundwater mixes with the Clear Lake aquifer and decreases the SAR. This is illustrated by the SAR and SC from samples at wells 3612, 221637, 221634, and 169217 (figs. 21, 22). Water from the Fort Union bedrock well 3612, south of the buried valley, has an SAR value of 65.3. About 2 mi north, well 221637, completed 288 ft deep in the lower alluvial zone, has an SAR of 62. The SAR steadily decreases to the north as the high SAR bedrock groundwater is diluted. About  $\frac{1}{4}$  mi north of 221637, the SAR at 221634 is 10.3. The SAR decreases to 2.1 at well 169217, 1 mi farther north. There apparently is enough high SAR water flowing from Fort Union aquifers to degrade the buried valley groundwater quality over large areas near the aquifer boundaries. Fresh, low SAR water enters the buried valley aquifer as precipitation recharges the sand hills area, to the east of the cross section in figure 22. This recharge results in water quality in the center of the aquifer suitable for irrigation development.

Groundwater can become unusable for irrigation if brines produced during hydrocarbon extraction infiltrate the aquifers. The TDS of water in deep aquifers associated with hydrocarbons can reach up to 10 times the TDS of seawater (Reiten and Tischmack, 1993). The average SAR of the brines produced with hydro-

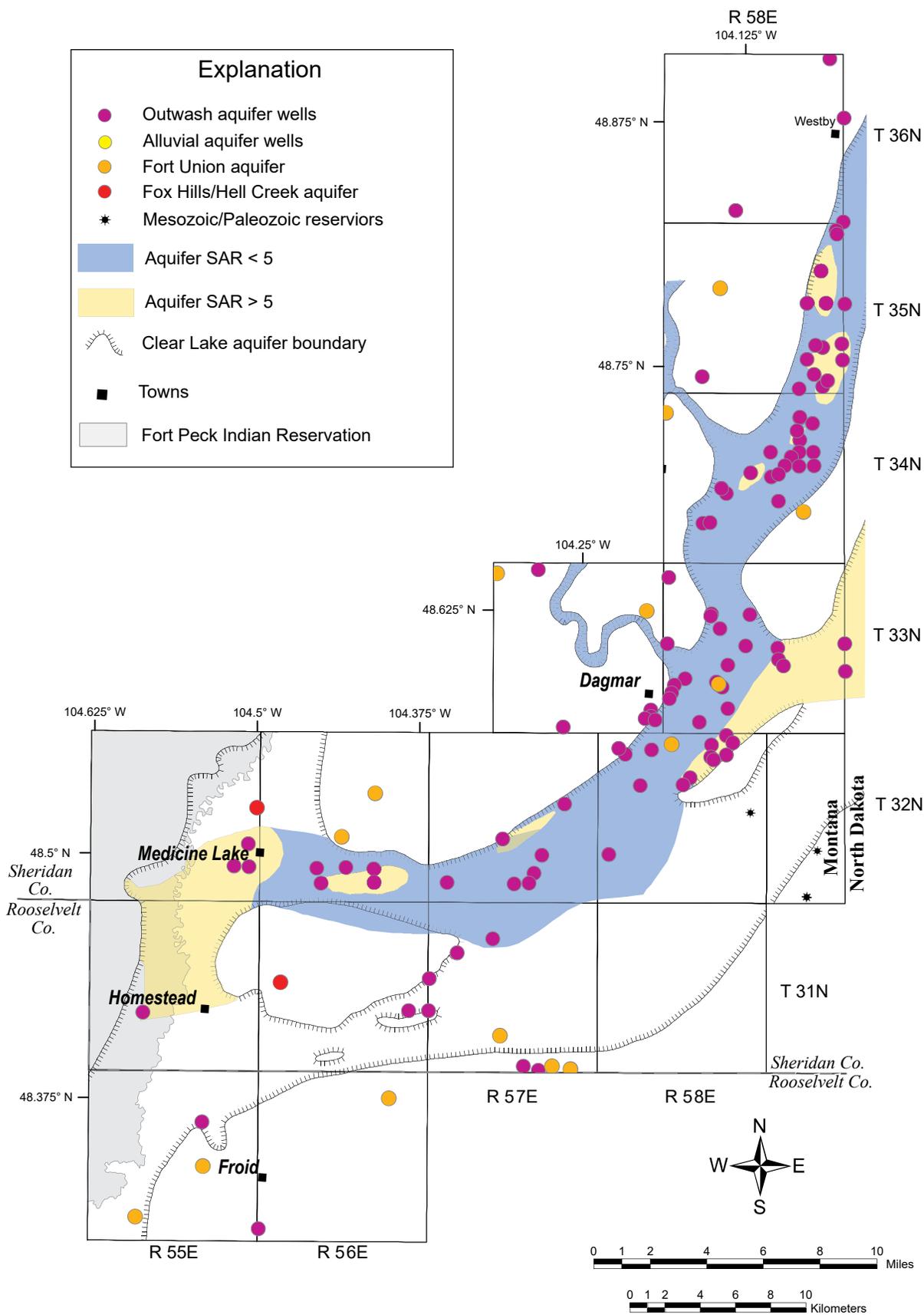
carbons is about 208 (table 3). Only small volumes from leaks and spills of produced water have been documented infiltrating irrigation supplies (Reiten, 2002). Because of high chloride concentrations in the produced water, spikes in chloride concentrations affect groundwater prior to major changes in SAR or SC. Samples collected from an irrigation well (3859) west of Brush Lake increased in chloride concentration from 32 mg/L in 1982 to 100 mg/L in 1995. In 2015, a sample from this well showed a small decrease in the chloride concentration, to 92 mg/L. Although the SC has increased about 20% since 1982, the SAR has remained stable at 1.9. While it appears that produced waters have affected the Clear Lake aquifer water quality at this site, the groundwater remains suitable for irrigation.

## SOUTH MEDICINE LAKE FOCUS AREA

The South Medicine Lake area was selected for a focused investigation because of concerns that irrigation withdrawals were affecting the water level in Medicine Lake and flows in Big Muddy Creek, and that additional groundwater development for irrigation could have detrimental effects on surface water. Irrigation wells in the South Medicine Lake area are 150–250 ft deep with static water levels 40 to 125 ft below the surface. The aquifer has variable water quality and is slow to recharge. Pumping for irrigation produces large drawdowns in some areas with the potential for well interference. Extensive test drilling and monitoring in this isolated branch of the Clear Lake aquifer provided hydrogeologic information for groundwater-flow models (Chandler and Reiten, 2019).

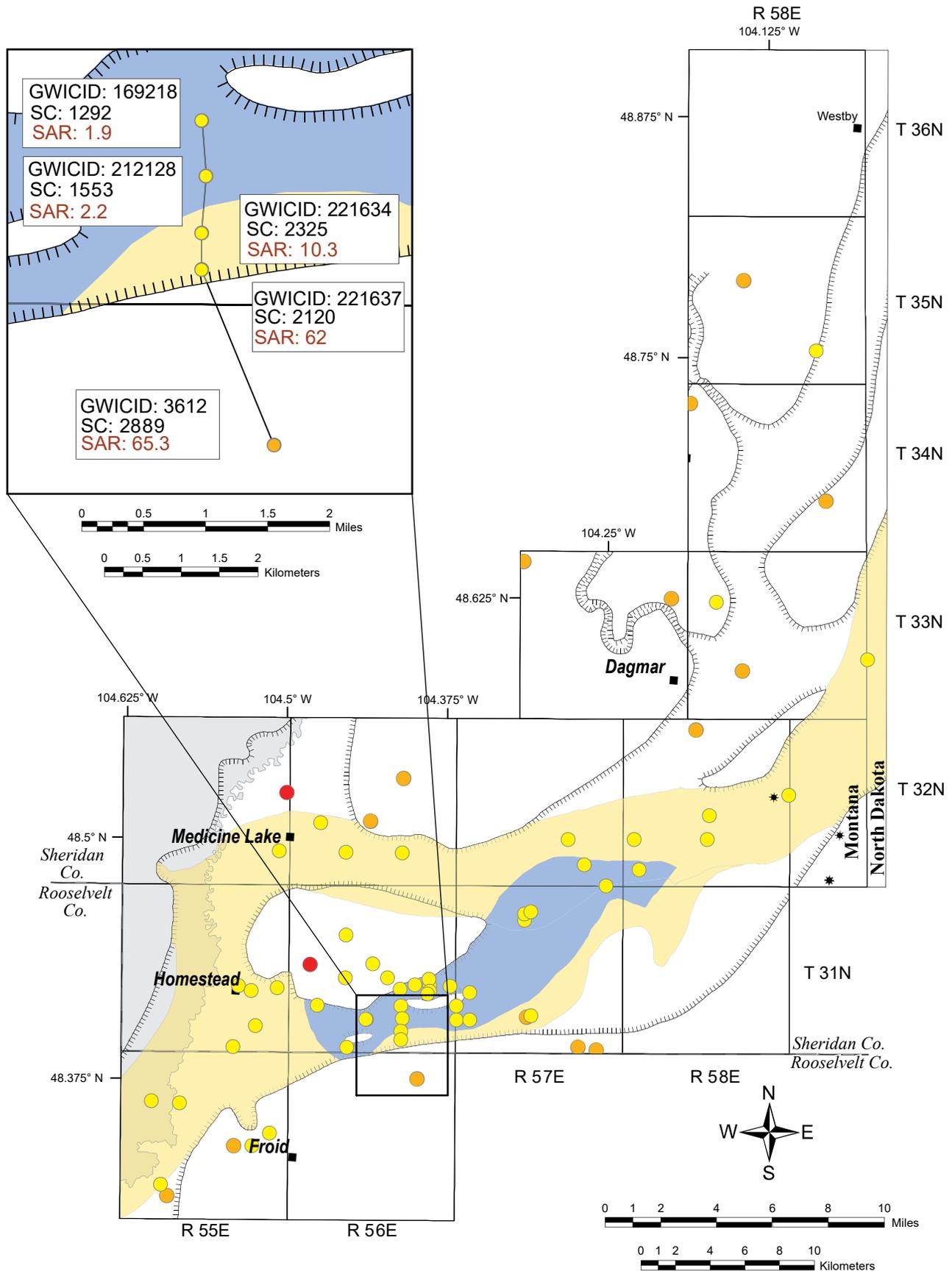
### Hydrogeologic Framework

The Clear Lake aquifer in the focus area occupies a west-to-east-trending buried valley located south of Medicine Lake. The aquifer is bounded to the north by a bedrock ridge that underlies the southern part of Medicine Lake. This effectively separates the south channel from the deeper north channel of the aquifer (fig. 23). The south channel boundary is a buried bedrock escarpment formed by erosion. A map of the elevation of the bedrock surface shows the incised valley, which is now filled by the aquifer materials and overlying sediments (fig. 23). The southern boundary crosses the surface expression of the Brockton–Froid Fault zone (Bergantino and Wilde, 1999; fig. 24).



A. Outwash Gravel SAR

Figure 21. Clear Lake aquifer water quality based on sodium adsorption ratio. (A) Outwash gravel. B on following page.



### B. Alluvial Gravel SAR

Figure 21, continued. (B) alluvial gravel.

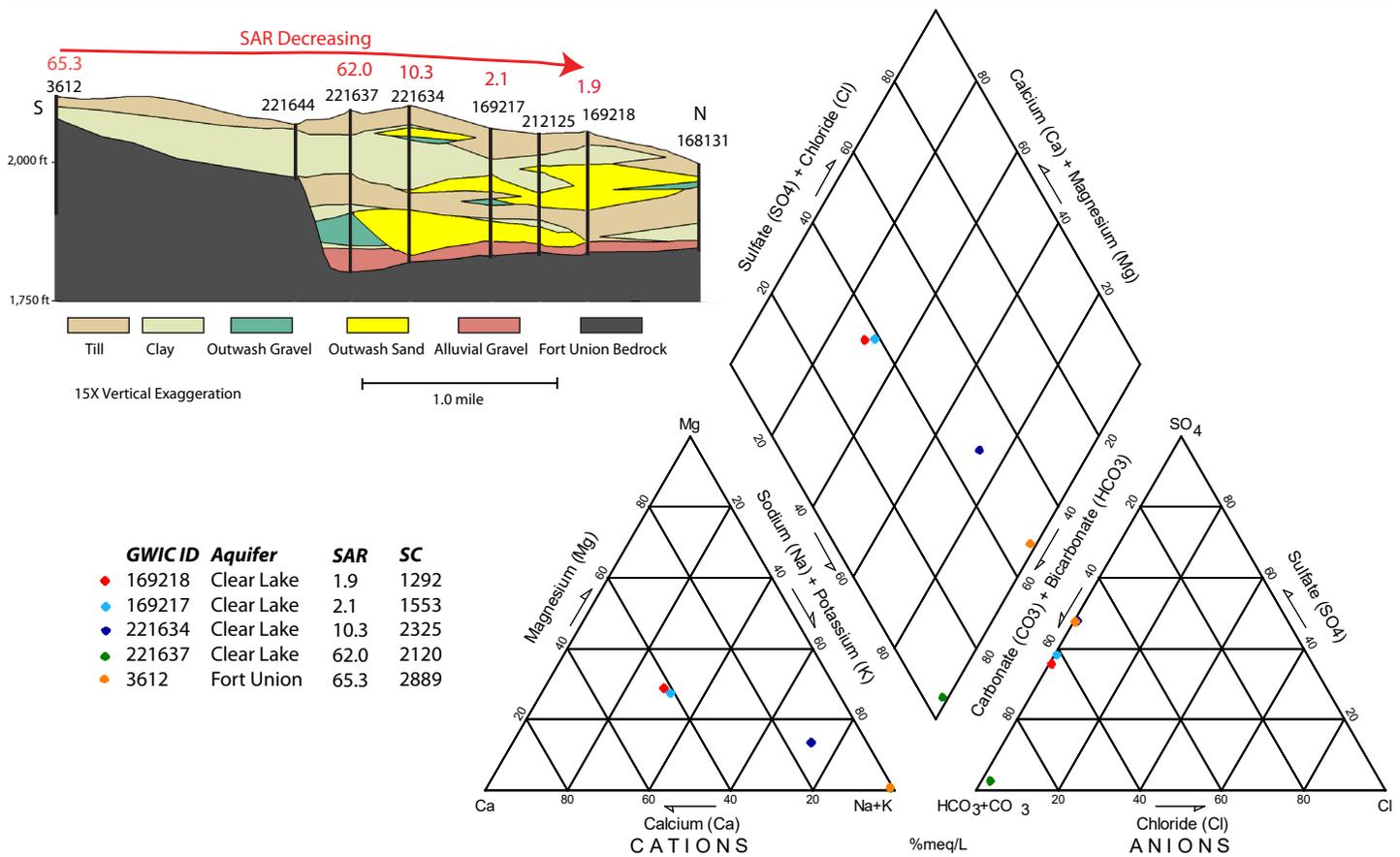


Figure 22. Piper plot showing water-quality differences near the southern boundary of the Clear Lake aquifer and cross section showing stratigraphy and changing SAR conditions in the aquifer. Well locations are on figure 21B, inset.

fault offsets deep basement rocks (Rogers and others, 1985) as well as Quaternary glacial sediments (Colton, 1963).

Lithologic logs and interpretations of hydrogeologic conditions were used to define the model boundaries of the Clear Lake aquifer. Four cross sections were constructed perpendicular to the buried valley (fig. 24). The farthest east cross section profiles three distinct alluvial channels separated by low-permeability bedrock boundaries (fig. 24, cross section A–A'). A bedrock high forms a hydraulic barrier separating the lower alluvial zone underlying the northern part of Medicine Lake from the lower alluvial zone in the focus area. Drilling density is insufficient to locate this boundary, but well logs indicated a thinning of the aquifer as shown in cross section A–A' at well 169219.

Other hydrogeologic observations provided useful information for identifying aquifer boundaries. Well 169219 is completed in sand and gravel of the Clear Lake aquifer. Water-level fluctuations show a strong response at this well to pumping at irrigation well 121117 about 1 mi to the southwest, confirming that

well 169219 has a good hydraulic connection to the Clear Lake aquifer (fig. 25). It typically has flowing artesian conditions with a potentiometric surface about 35–40 ft above the level of Medicine Lake (cross section A–A'). If the aquifer had a good hydraulic connection to the lake at this well, aquifer heads would equilibrate with the lake level and the well would not flow. These observations indicate confined conditions and a barrier boundary between well 169219 and the lake. Well 169220 is another flowing well located about 2.2 mi northeast of irrigation well 121117. Figure 25 compares the water-level fluctuations at well 43113 (25 ft NE of well 121117), well 169219 (5,035 ft NE), and well 169220 (11,670 ft NE); these wells respond to pumping at irrigation well 121117.

In the focus area, the aquifer splits into two southern channels. Bedrock highs form longitudinal hydraulic barriers, which restrict lateral flow within the aquifer (fig. 23). Shaver and Pusc (1992) identified similar longitudinal barriers in a buried valley aquifer near Minot, North Dakota. These longitudinal barriers compartmentalize buried valleys and isolate parts of the aquifer. Shaver and Pusc (1992) observed that these

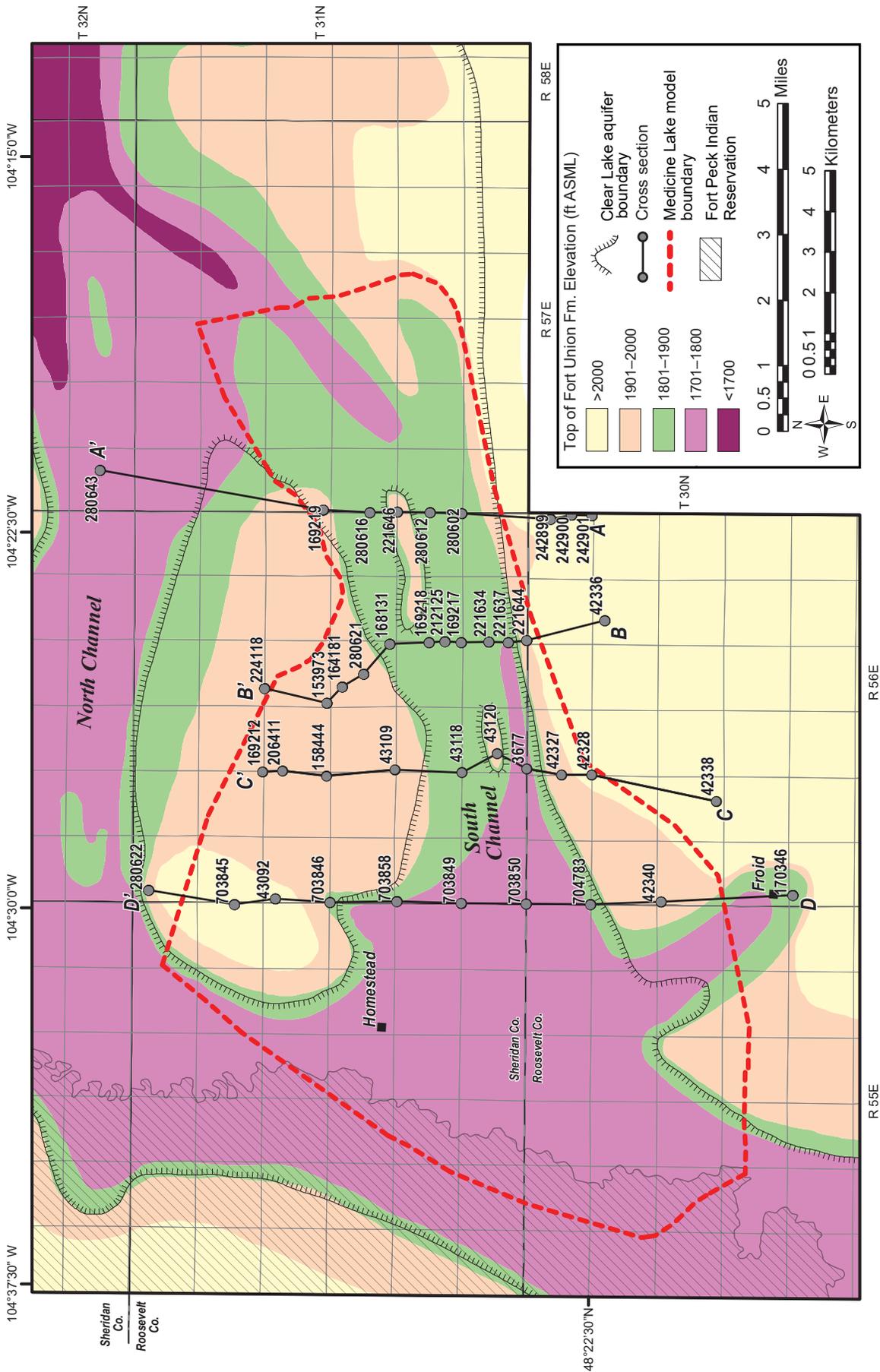


Figure 23. Altitude of the top of the Fort Union Formation showing the South Medicine Lake cross sections.

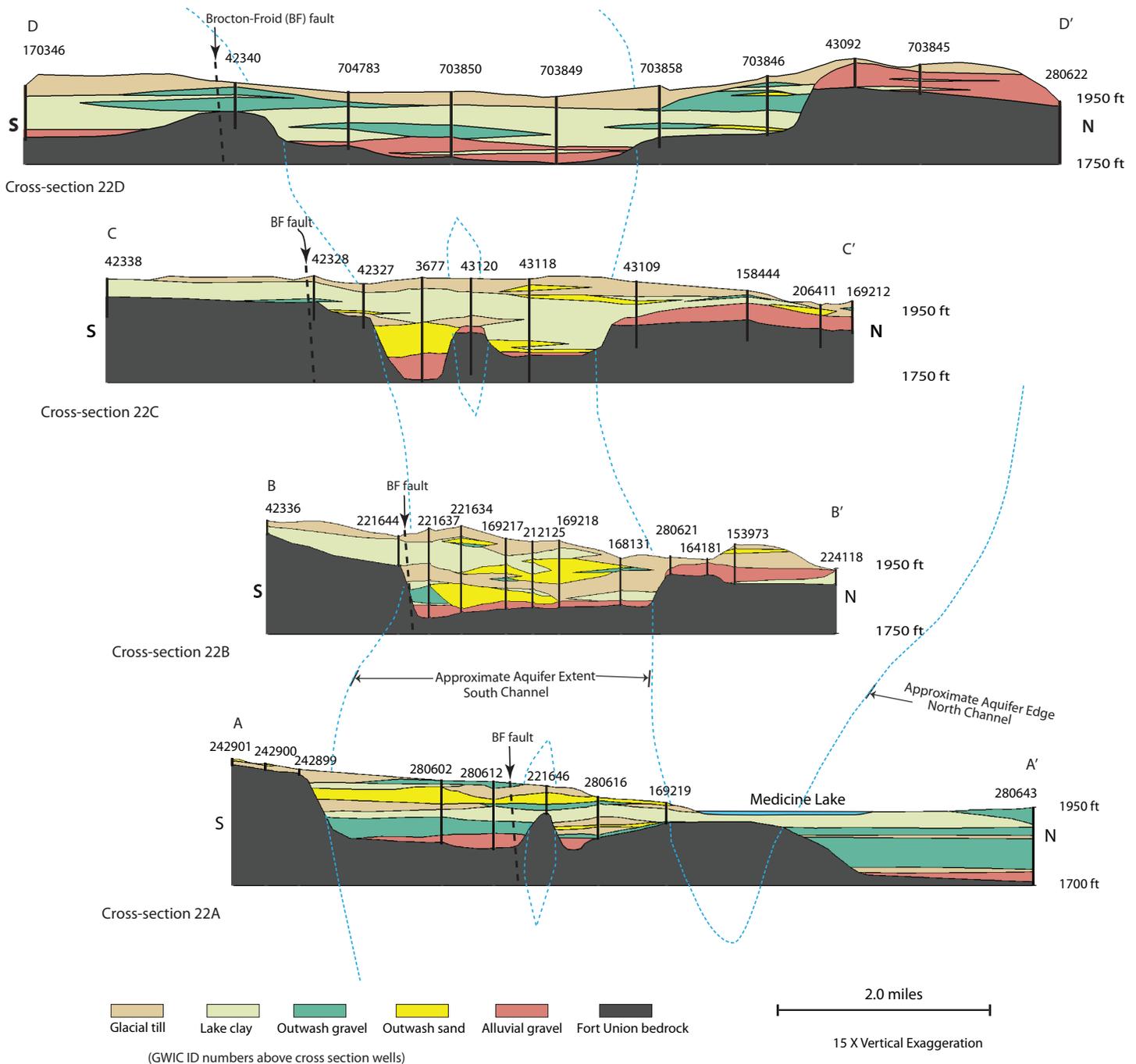


Figure 24. Cross sections constructed perpendicular to the axis of the South Medicine Lake buried valley display current interpretations of the hydrostratigraphy of the Clear Lake aquifer southern channel. (Cross section locations, fig. 23).

longitudinal barriers form at a variety of scales and can completely hydraulically isolate and compartmentalize subchannels or act as partial hydraulic barriers. Cross sections A–A', B–B', C–C', and D–D' (fig. 24) present an interpretation of the channel geometry and aquifer from east to west. Cross section B–B' (fig. 24) profiles the buried valley as one channel bounded on the south by the Brockton–Froid Fault zone (labeled BF in fig. 24). This steep bedrock escarpment bounds the lower alluvial zone and groundwater inflow from the bedrock mixes with the alluvial groundwater. The

northwest edge of the buried valley along cross section B–B' is located between well 168131 and well 280621. North of the buried channel, along this cross section, thick sand and gravel associated with a higher terrace of the ancestral Missouri River overlies Fort Union bedrock. This upper alluvial terrace forms an aquifer, which is hydrologically isolated from the Clear Lake aquifer by low conductivity glacial till and clay. Similar terraces occur along much of the northwest flank of the buried valley.

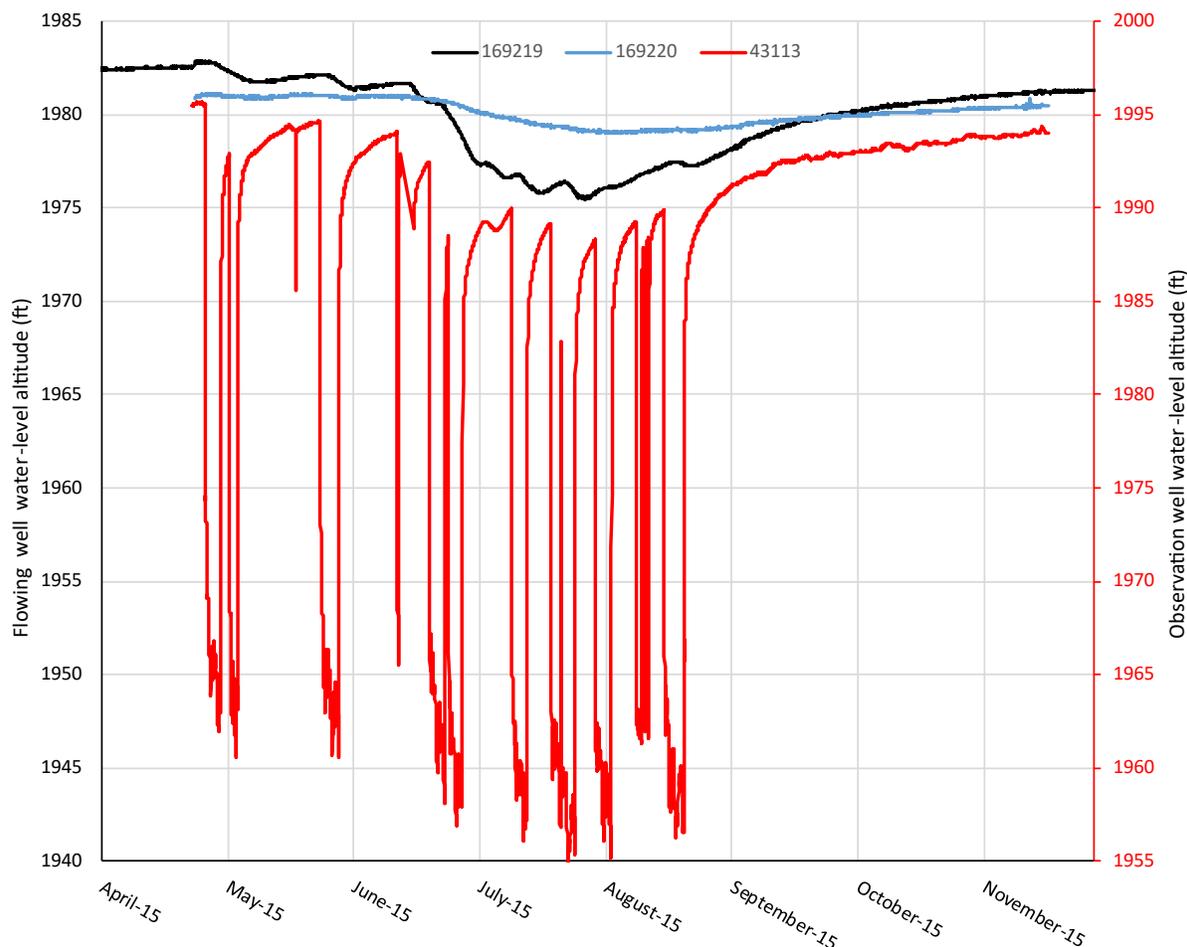


Figure 25. Hydrograph showing response at three wells to pumping of irrigation well 121117. Well 43113 is located 25 ft northeast of the irrigation well. Two wells under flowing artesian conditions were also monitored: well 169219 is 0.95 mi northeast and well 169220 is located 2.2 mi northeast. Right axis for well 43113.

To the west, cross section C–C' (fig. 24) identifies two lower alluvial zone channels separated by a bedrock ridge. The southern channel incises into the Fort Union bedrock to a depth of 350 ft. This is about 100 ft deeper than any other segment of the buried valley aquifer in the south Medicine Lake area. As in other cross sections, relatively steep escarpments incised into the Fort Union bedrock form the lateral boundaries of the buried valley aquifer. A higher elevation alluvial gravel terrace overlies the bedrock ridge (well 43120). The groundwater in this terrace north of the buried valley is unconfined (wells 158444 and 206411). Heads in the terrace are about 20 ft lower than in the lower alluvial zone of the south channel. The upper terrace underlies most of the area between the south buried valley channel and Medicine Lake. Groundwater flow in the upper terrace aquifer is towards the lake and daylights at springs and seeps in the small drainages.

The westernmost cross section D–D' (fig. 24) profiles a wide, buried valley portion of the Clear Lake

aquifer bounded laterally by Fort Union bedrock. The southern boundary closely aligns with the Brockton–Froid Fault zone. South of the fault, a tributary channel to the buried valley projects into the cross section. This may be a lower alluvial channel or an upper outwash zone channel associated with Sheep Creek near Froid. On the north end of this cross section, the upper alluvial terrace aquifer is isolated from the lower alluvial zone of the Clear Lake aquifer.

### Aquifer Properties

Water-level monitoring in 2015 near four irrigation wells provided data to estimate aquifer parameters. Irrigation pumping resulted in drawdown in nearby monitoring wells, which was analyzed as fortuitous aquifer tests. Tables 4–7 summarize these tests, and interpretations are presented in appendix C.

Average parameters calculated from multiple tests at each well represent reasonable estimates of the aquifer conditions. The average transmissivity ranged from about 2,940 ft<sup>2</sup>/d to 9,900 ft<sup>2</sup>/d. The average

Table 4. Summary of 2015 aquifer tests at Nelson 1 (121117), monitoring well 43113.

Test	Test Date Range	Time Pumped (min)	Pumping Rate Q (gpm)	Transmissivity T (ft <sup>2</sup> /day)	Aquifer Thickness (ft)	Hydraulic Conductivity K (ft/day)	Storage Coefficient
1	4/16–4/27	2,055	800	3,203	30	107	0.0026
2	5/2–5/4	2,755	800	3,240	30	108	0.0019
3	5/24–5/29	6,525	800	3,316	30	110	0.0025
4	6/11	495	800	3,203	30	107	0.0024
5	6/20–6/24	5,460	800	3,240	30	108	0.0019
6	6/25–6/28	4,920	800	3,240	30	108	0.0024
7	7/10–7/14	5,520	800	3,240	30	108	0.0019
8	7/20–7/26	8,520	800	3,203	30	107	0.0026
9	7/31–8/4	5,760	800	3,203	30	107	0.0023
10	8/10–8/11	1,755	800	3,240	30	108	0.0047
11	8/11	285	800	3,240	30	108	0.0024
12	8/12	270	800	3,240	30	108	0.0018
13	8/17–8/22	6,855	800	3,240	30	108	0.0015
			Minimum	3,203	30	107	0.0015
			Maximum	3,316	30	110	0.0047
			Average	3,234	30	108	0.0024

Note. Distance:  $r = 25$ , Cooper–Jacob method.

Table 5. Summary of 2015 aquifer tests at Nelson 2 (155930), monitoring well 155929.

Test	Test Date Range	Timed Pumped (min)	Pumping Rate Q (gpm)	Transmissivity T (ft <sup>2</sup> /day)	Aquifer Thickness (ft)	Hydraulic Conductivity (ft/day)	Storage Coefficient
1	6/22–6/24	2,280	750	3,722	22	169	0.00021
2	7/22–7/25	3,570	750	4,556	22	207	0.00015
3	7/3–7/7	5,560	750	4,804	22	218	0.00015
4	7/10–7/13	3,660	750	4,719	22	214	0.00015
5	7/14–7/15	1,800	750	5,285	22	240	0.00017
			Minimum	3,722	22	169	0.00015
			Maximum	5,285	22	240	0.00021
			Average	4,617	22	210	0.00017

Note. Distance:  $r = 700$  ft, Cooper–Jacob method.

hydraulic conductivity ranged from 108 ft/d to 450 ft/d. The average storage coefficient ranged from 0.00017 to 0.017. Storage coefficient estimates from monitoring wells less than 25 ft from the production wells were about one order of magnitude greater than estimates from more distant monitoring wells. These tests likely reflect non-linear flow and are not presented here. A reasonable estimate of storage coefficient in the South Medicine Lake focus area is about 0.0002. The range of estimates are similar to estimates for the lower alluvial zone shown in table 2. The presence of thick fine-grained sediments overlying the aquifer, the low magnitude of the

storage coefficient, and observation of relatively high barometric efficiencies indicate confined aquifer conditions.

Water-level measurements from the 2015 irrigation season identified areas with large amounts of draw-down, which suggests proximity to barrier boundaries and potential well interference. When a cone of depression around a production well contacts a barrier boundary, the rate of change of drawdown doubles (Fetter, 2001). Boundary effects are shown on time-drawdown plots compiled in appendix C4. Figure 26 shows the maximum drawdown observed in moni-

Table 6. Summary of 2015 aquifer tests at Nelson 3 (212128), monitoring well 169217.

Test	Test Date Range	Time Pumped (min)	Pumping Rate Q (gpm)	Transmissivity T (ft <sup>2</sup> /day)	Aquifer Thickness(ft)	Hydraulic Conductivity (ft/day)	Storage Coefficient
1	5/2–5/4	2,805	665	10,413	22	473	0.00048
2	5/29–6/1	4,380	665	9,372	22	426	0.00020
3	6/12–6/13	1,185	665	9,117	22	414	0.00028
4	6/16–6/17	1,620	665	10,898	22	495	0.00013
5	6/22–6/23	1,560	665	10,276	22	467	0.00036
6	6/26–6/30	5,880	665	9,372	22	426	0.00027
7	7/3–7/7	5,565	665	9,602	22	436	0.00020
8	7/10–7/12	2,280	665	10,413	22	473	0.00013
9	7/22–7/26	5,640	665	9,448	22	429	0.00020
			Minimum	9,117	22	414	0.00013
			Maximum	10,898	22	495	0.00048
			Average	9,879	22	449	0.00025

Note. Distance:  $r = 210$  ft, Cooper–Jacob method.

Table 7. Summary of 2015 aquifer tests at Bolstad (175651), monitoring well 43116.

Test	Test Date Range	Time Pumped (min)	Pumping Rate Q (gpm)	Transmissivity T (ft <sup>2</sup> /day)	Aquifer Thickness (ft)	Hydraulic Conductivity K (ft/day)	Storage Coefficient
1	4/28–4/29	1,275	795	2,745	30	152	0.0024
2	5/23–5/29	8,130	795	2,334	30	130	0.052
3	7/31–8/6	9,000	795	3,257	30	181	0.027
4	6/9–6/11	3,183	795	3,165	30	177	0.0024
5	8/10–8/17	10,000	795	3,220	30	179	0.0024
			Minimum	2,334	30	130	0.0024
			Maximum	3,257	30	181	0.0520
			Average	2,944	30	164	0.0172

Note. Distance:  $r = 20$  ft, Cooper–Jacob method.

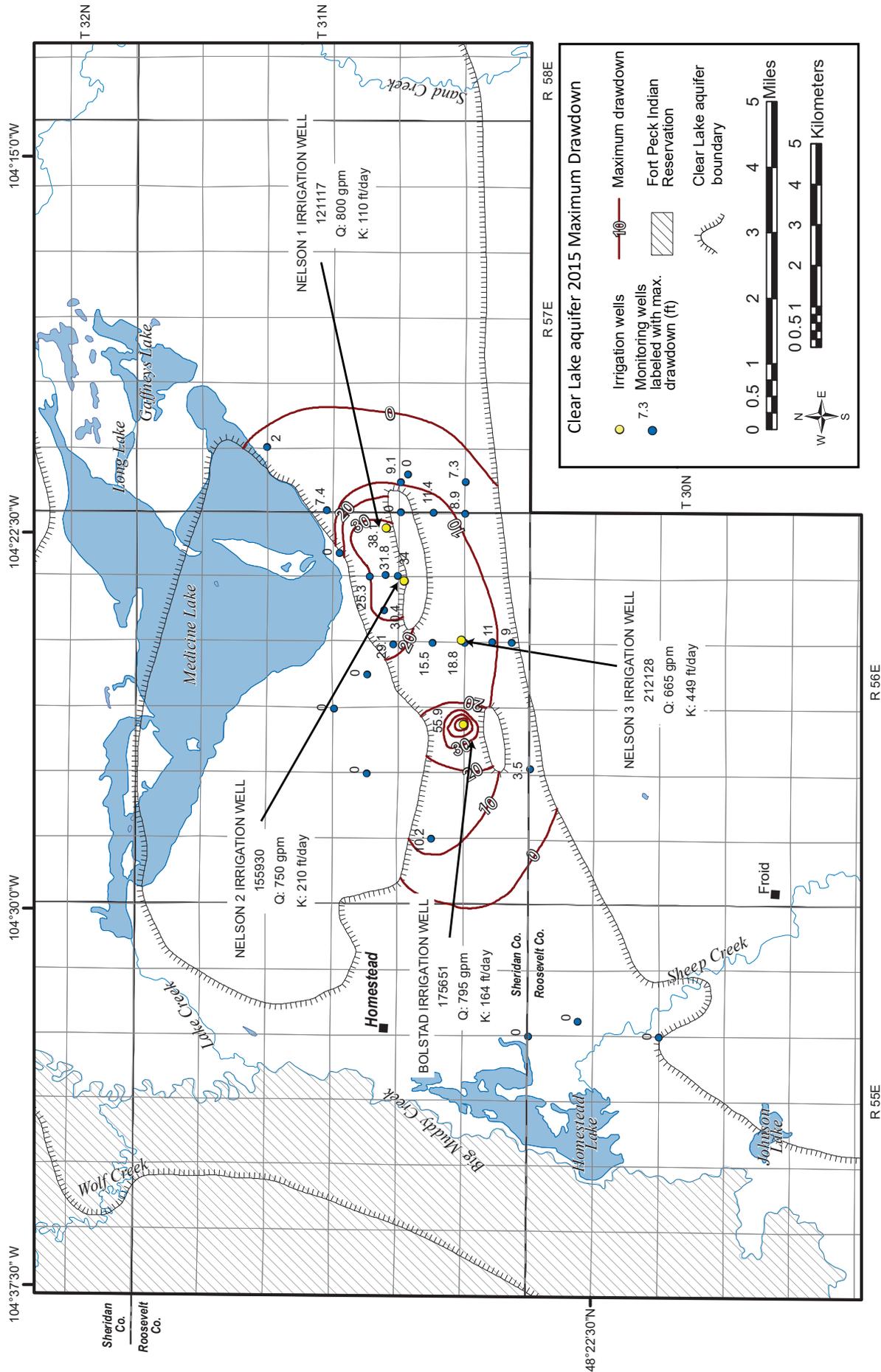
toring wells of the South Medicine Lake area during the summer 2015 irrigation season. The four irrigation wells in this area are identified along with the average pumping rate, volume of water pumped in 2015, and average hydraulic conductivity. Three of the wells—Nelson 1 (well 121117), Nelson 2 (well 155930), and Bolstad (well 175651)—are associated with the northern channel of the focus area. The fourth well (Nelson 3, 212128) is more aligned with the southern channel. The distribution of the aquifer parameters suggests higher transmissivity and hydraulic conductivity associated with the southernmost channel.

The greatest drawdown observed was in the immediate vicinity of the Bolstad well. Data loggers measured over 50 ft of drawdown at the monitoring well located about 25 ft from the Bolstad production

well (well 43116). The largest area of cumulative drawdown was in the north channel of the aquifer near Nelson 1 and Nelson 2 irrigation wells, showing well interference. The large drawdowns are strong evidence for nearby hydraulic boundaries (van der Kamp and Maathuis, 2012).

### Groundwater Flow

The potentiometric map was constructed from water levels measured in April–May 2015 (fig. 14). Water levels were recovering from the 2014 irrigation season and did not appear to be affected by spring recharge prior to the 2015 irrigation season. Groundwater flows from northeast to southwest in much of the South Medicine Lake focus area (fig. 14). The flow is away from the groundwater mound underlying an area of focused recharge at the overlying sand hills.



### Groundwater Recharge and Discharge

Infiltration of rainfall and snowmelt directly recharges the aquifer, and other areas extending several miles laterally likely provide additional recharge to the Clear Lake aquifer through groundwater flux. The permeability of surface materials is a dominant factor controlling recharge rates. This is evident in the sand hills, where the loose fine sand allows rapid infiltration without runoff. The Quaternary eolium mapped in figure 8 shows the extent of sand hills. Other sources of recharge likely include unmapped gravel deposits and the bedrock aquifers along the edges of the buried valley aquifer. The numeric model indicates these are both important, but poorly defined, sources of recharge (Chandler and Reiten, 2019).

Discharge from the South Medicine Lake portion of the Clear Lake aquifer is to deep and shallow deposits in the Big Muddy Creek Valley between Homestead and Froid (fig. 14). Several wells completed in deep gravel deposits indicate a continuation of the aquifer under the Big Muddy Creek Valley, but the extent and hydraulic gradient of this deep aquifer is unknown.

As groundwater flows towards Big Muddy Creek,

there is likewise an unknown volume of upward seepage to the wetlands and the creek. This seepage is a fraction of wetland inflow as water also flows in from deep and shallow sources upgradient to the north and west. In this area south of Homestead, wetlands cover about four sections along the east side of Big Muddy Creek. ET from area wetlands is estimated to remove 3 to 5 ft per year (Donovan, 1988). Based on a four section area (2,560 acres), ET discharge from this part of the wetlands ranges from about 7,700 acre-ft/yr to 12,800 acre-ft/yr. The volume pumped for irrigation (typically less than 600 acre-ft/yr) is 5–7% of the estimated water loss to ET. For pumping to decrease aquifer seepage to the wetlands or creeks, there would have to be a decrease in the deep aquifer heads. We observed no head changes in response to irrigation pumping in deep upgradient wells. Deep aquifer water levels actually increased during the irrigation season of 2015 at well 3605, just east of Homestead Lake (fig. 27). It is unlikely that pumping from the four existing irrigation wells measurably affects water levels in the Homestead area wetlands or Big Muddy Creek.

### Groundwater Modeling and Water-Use Scenarios

The numerical models presented by Chandler and Reiten (2020) were based on detailed stratigraphic

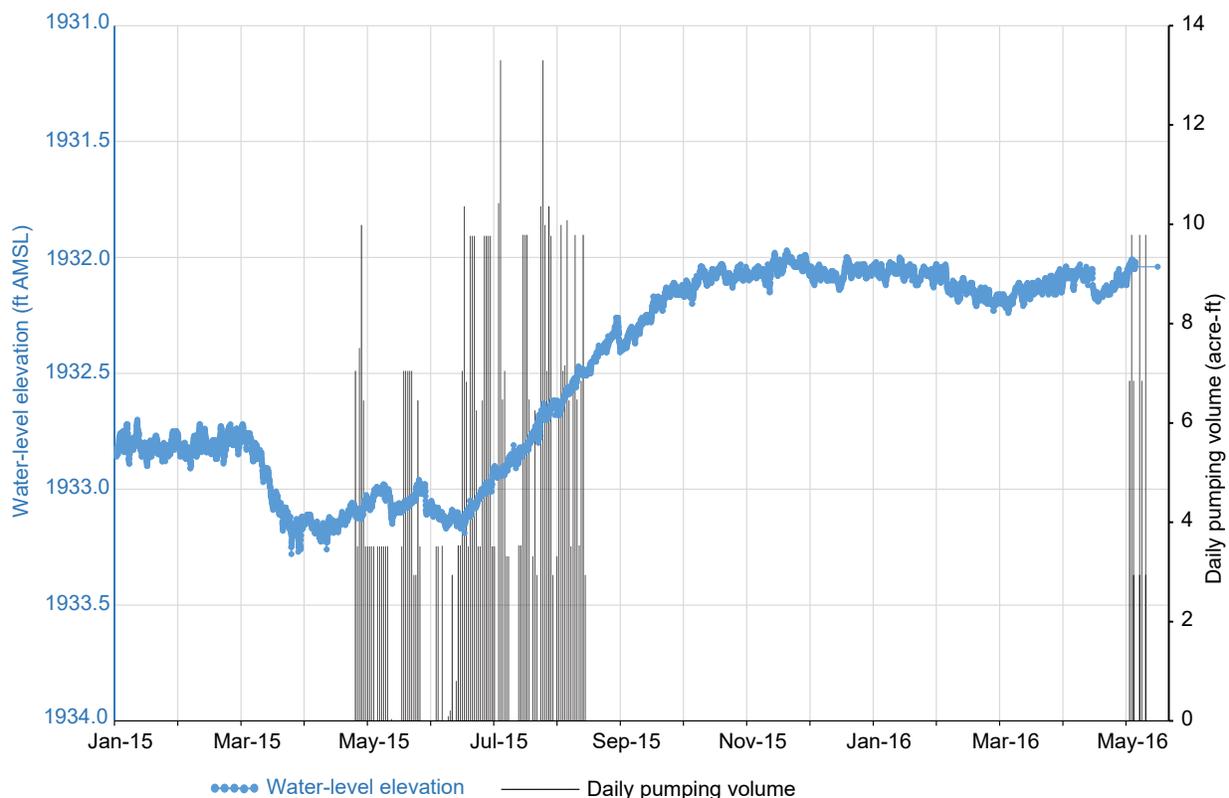


Figure 27. Water levels at well 3605 do not show response to irrigation well pumping from the Clear Lake aquifer south of Medicine Lake. Monitoring well 3605 (185 ft deep) is located south of Homestead in the Big Muddy Creek Valley (plate 3).

modeling, records of irrigation water-use and pumping duration, and aquifer parameters from this study. The model simulates the timing and magnitude of 2015 observed drawdown (Scenario 1). After calibration, a “no use” or “control” output was generated by simulating the system with no irrigation pumping (Scenario 2). Additional model runs simulated increased irrigation water use (Scenario 3) and twice the number of irrigation wells (Scenarios 4 and 5). The head changes resulting from increased use were isolated from those related to boundary condition effects (recharge and ET) by calculating the head difference between “increased use” and “no use” simulations. The head difference results for the time periods of maximum drawdown are contoured for comparison (figs. 28–31). Water-level changes were compared at four model cells in the Big Muddy Creek Valley. Cell 80239 showed the greatest drawdown of the four cells and is therefore mentioned in the scenarios for comparison. Hydrographs of water levels at four model cells near wetlands in the Big Muddy Creek alluvial valley illustrate results of Scenario 5, the greatest water use of these simulations (fig. 32).

### Scenario Results

#### Scenario 1: Four Well 2015 Volume Pumping (483 acre-ft)

The first scenario simulated the pumping schedule recorded in 2015 at the four existing irrigation wells. The results of scenario 1 show well interference between the irrigation wells (fig. 28). Drawdown greater than 30 ft was predicted for the eastmost well 121117. This well is the closest to the barrier boundary between the aquifer and Medicine Lake, and this explains the large drawdown here. Figure 28 shows the maximum drawdown simulated during 2015. The drawdown from the wells intersect, producing cumulative drawdown or well interference as several irrigation wells pump simultaneously. The cumulative drawdown cone extends for more than 4 mi, with a maximum drawdown of 0.35 ft at cell 80239 near Homestead.

#### Scenario 2: “No use” (0 acre-ft)

The “no use” scenario simulated heads before irrigation development and provided a baseline for comparison of the different pumping scenarios. This

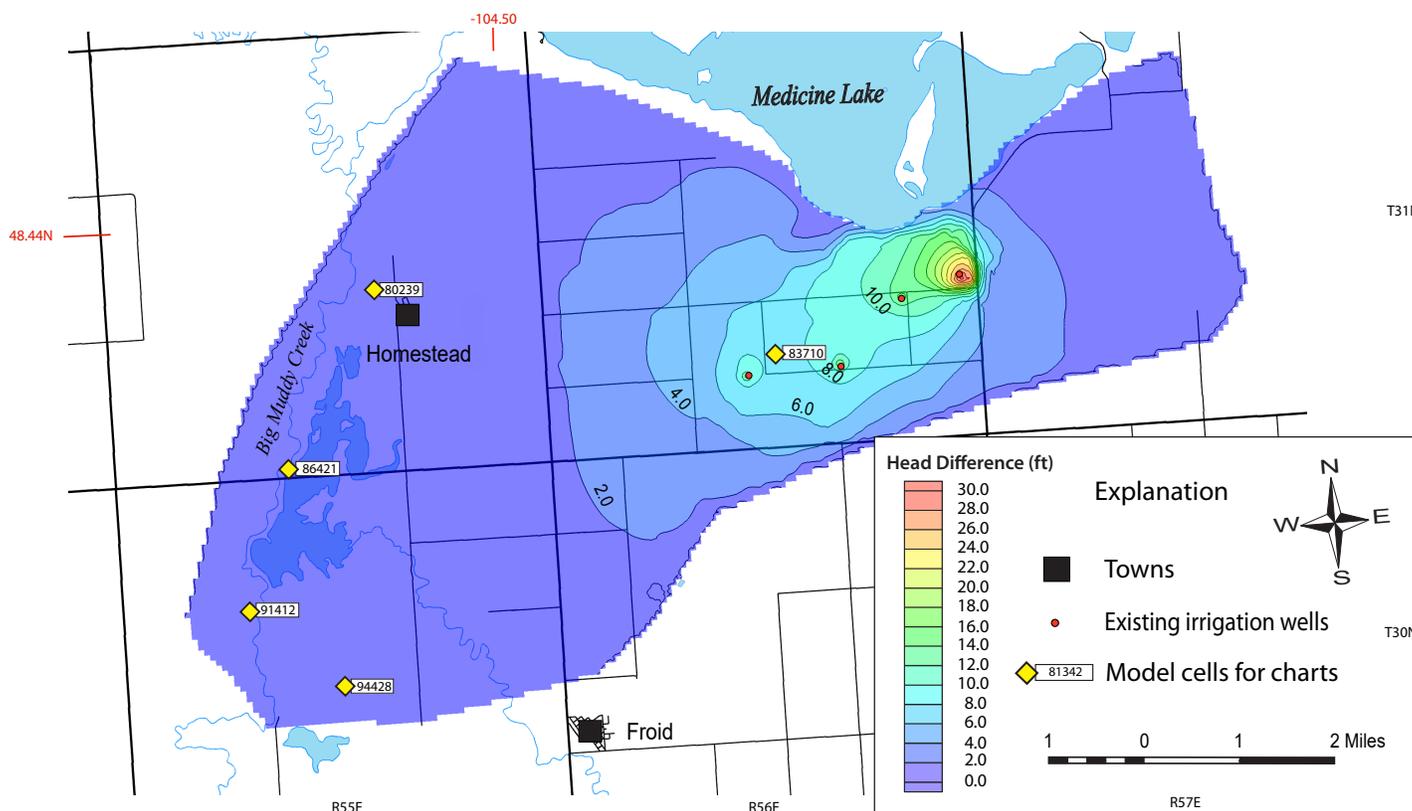


Figure 28. The difference between “no use” and 2015 volume scenarios shows well interference and the increased drawdown due to hydrogeologic boundaries at the eastmost irrigation well. This model output is for the point in time with greatest drawdown, mid-July.

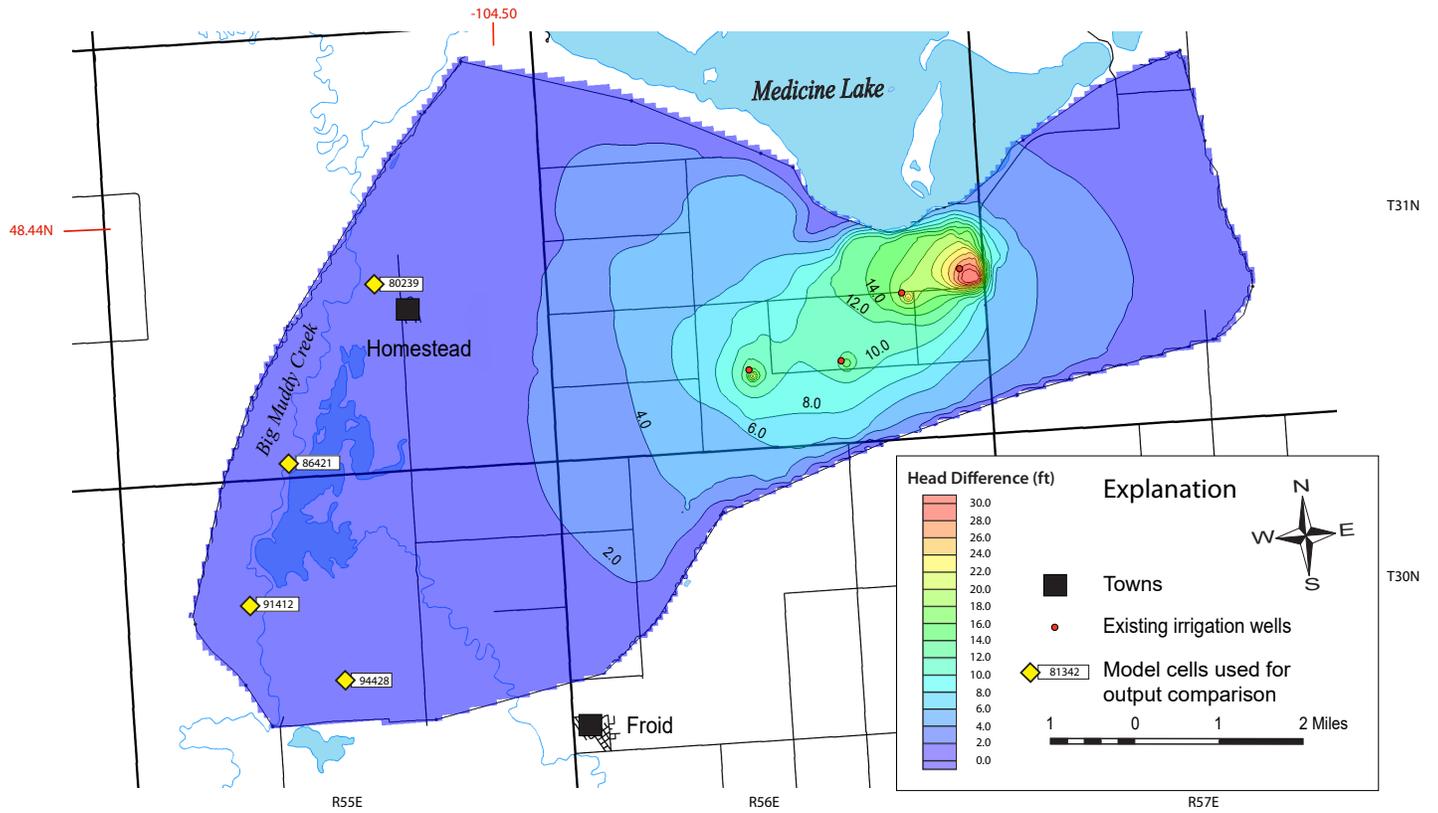


Figure 29. The head differences between the “no-use” scenario and a scenario where the four existing irrigation wells pump maximum allocation volumes show increased well interference. The model predicts an increased zone of influence with a maximum drawdown of 0.45 ft at cell 80239. This model output is for the point in time with greatest drawdown, mid-July.

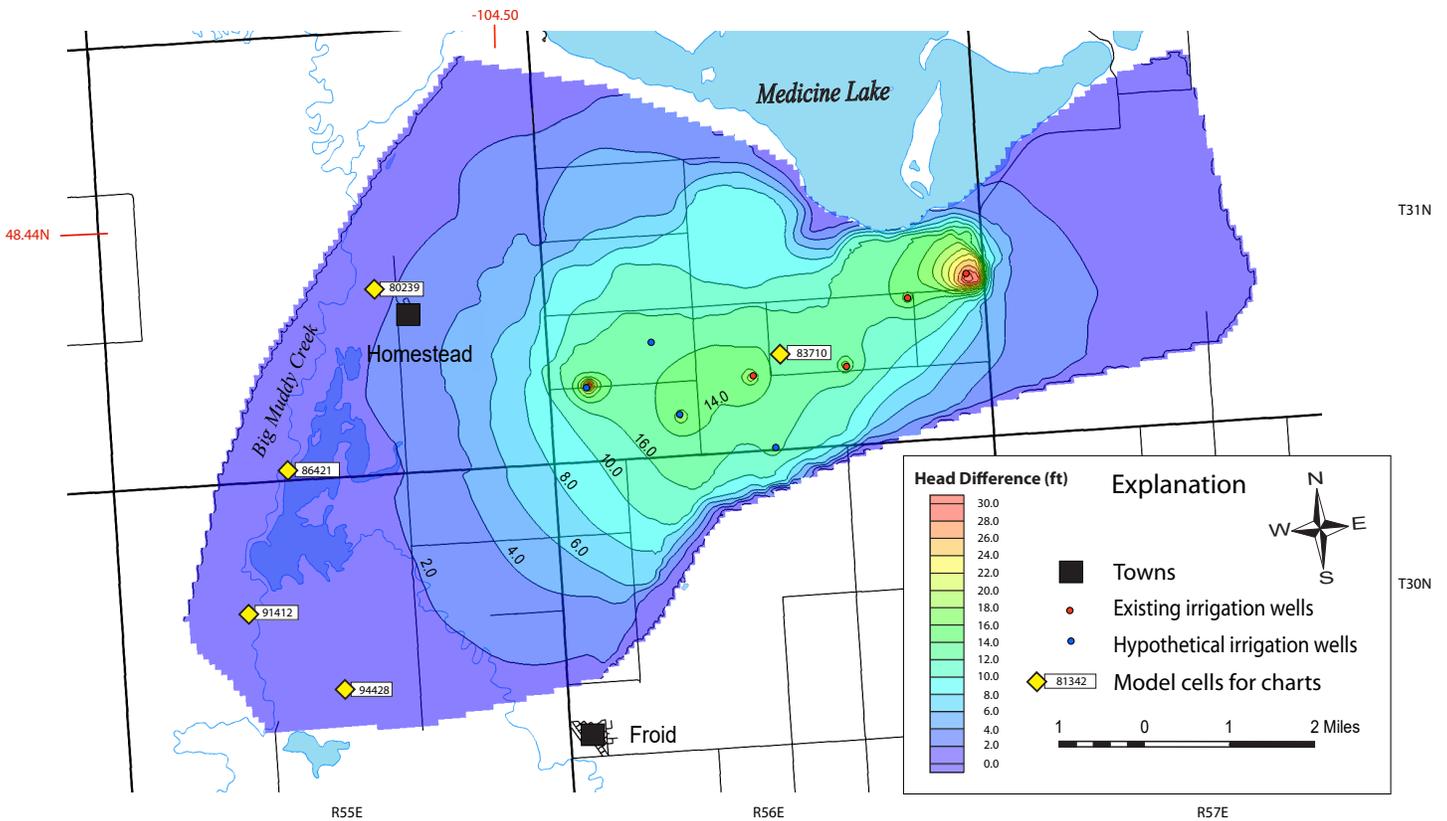


Figure 30. The head difference between a “no pumping” simulation and a simulation with eight irrigation wells pumping at 2015 volume predicts 2.1 ft of drawdown near the Big Muddy Creek wetlands at cell 80239. This model output is for the point in time with greatest drawdown, mid-July.

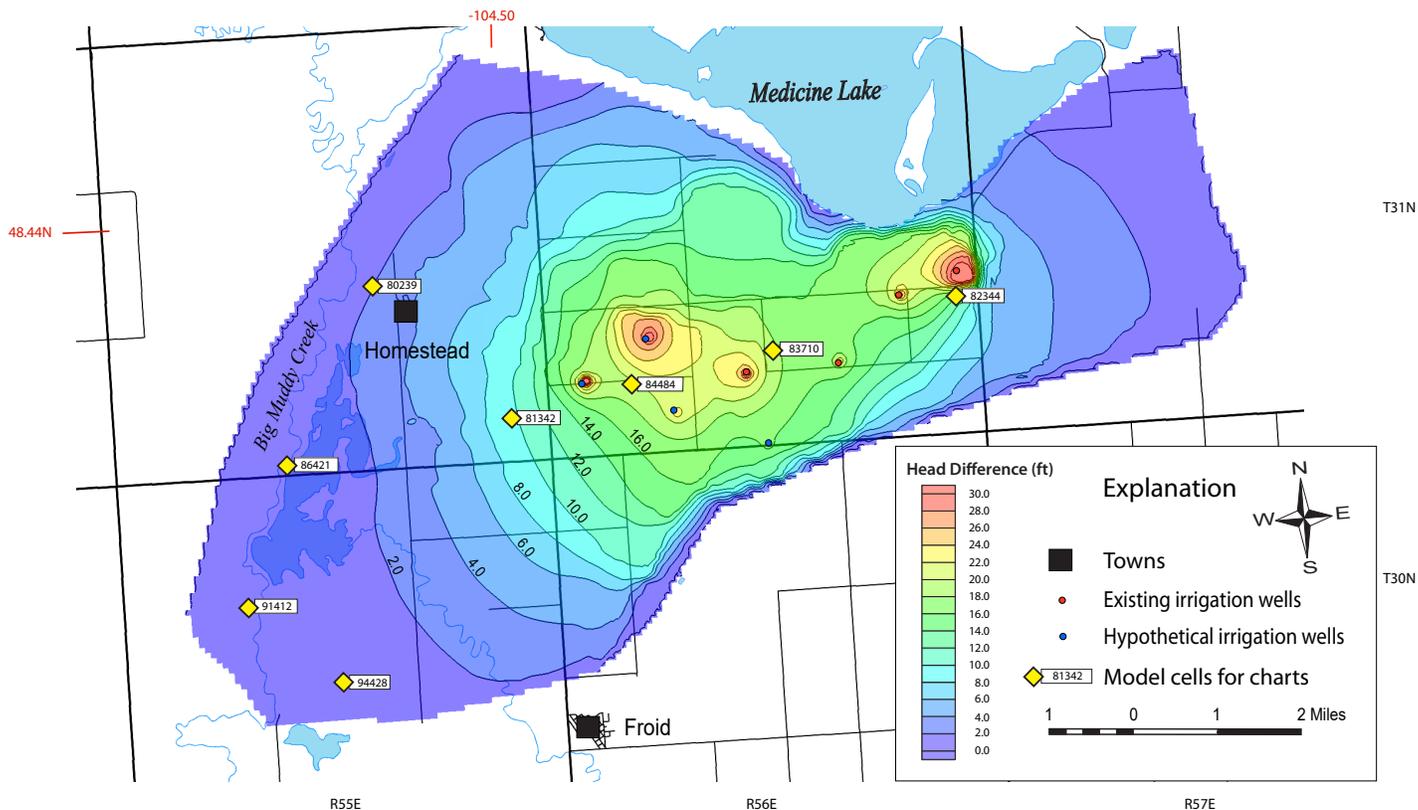


Figure 31. The head difference between a “no pumping” simulation and a simulation with all eight irrigation wells pumping with full allocation volumes predicts 2 ft of drawdown extending to the Big Muddy Creek wetlands near Homestead (cell 80239). This model output is for the point in time with greatest drawdown, mid-July.

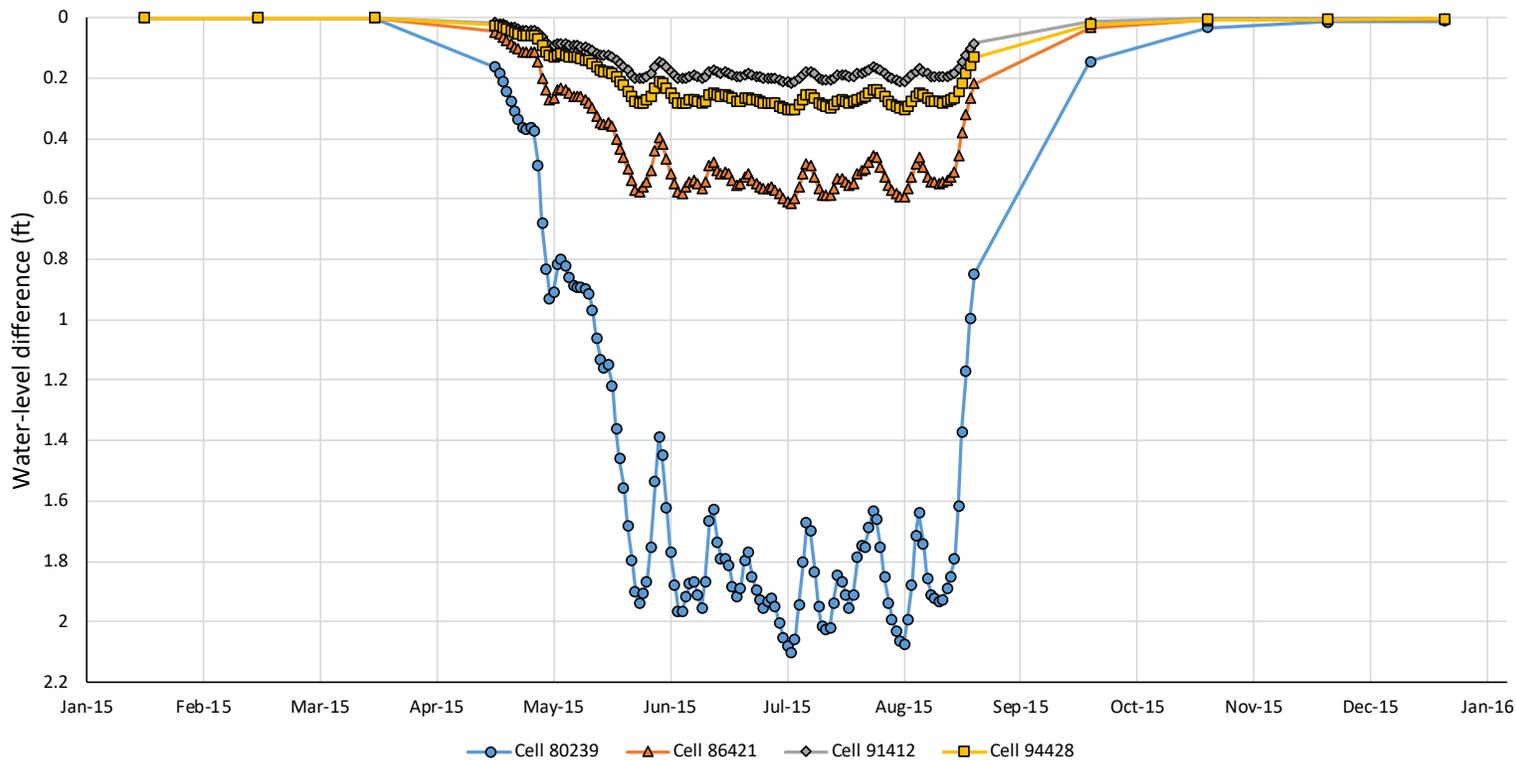


Figure 32. The water levels at four model cells along the Big Muddy show water-level declines of up to 2 ft with eight wells pumping the maximum allocation volume (2,585 acre-ft/yr total).

scenario showed minor seasonal water-level fluctuations resulting from ET and recharge. The time-series heads simulated for the other scenarios were subtracted from the “no use” heads. This removed fluctuations resulting from ET and recharge, thereby isolating the head changes resulting from pumping.

*Scenario 3: Four Well 2015 Full Allocation Volume Pumping (1,500 acre-ft)*

The third scenario simulated the four existing irrigation wells pumping at the full permitted allocation volume, increasing the cumulative aquifer drawdown. The water use during this scenario was 210 percent greater than the 2015 volume. To extract this volume during the model run, the four irrigation wells pumped nearly continuously. Water-use records show that area irrigators have not used the full allocation volume, even during dry years. The scenario predicts maximum drawdown of 0.45 ft at cell 80239 near Homestead (fig. 29).

*Scenario 4: Eight Wells Pumping at 2015 Volume (1,104 acre-ft)*

The fourth scenario simulated irrigation development by doubling the number of wells in the focus area. Four hypothetical wells were placed in an area with ample farmland and high-quality groundwater (locations shown in fig. 30), and thus suitable for irrigation expansion. The pumping schedule resembled that measured in the 2015 irrigation season. The hypothetical well pumping rates (set at 650, 700, 750, and 800 gpm) were similar to those of the existing irrigation wells (table 6). The total volume pumped from the four new wells was approximately 620 acre-ft. Pumping from the additional irrigation wells expanded the area of drawdown (fig. 30), primarily to the west, where the hypothetical wells were located. This scenario predicts approximately 1.4 ft of water-level change in the aquifer near cell 80239, west of Homestead.

*Scenario 5: Eight Wells Pumping the Full Allocation Volume (2,585 acre-ft)*

The fifth scenario simulated eight wells pumping the full allocation for a total volume of 2,585 acre-ft during the irrigation season. Pumping from the four existing wells equaled their permitted allocation of approximately 1,500 acre-ft. The four hypothetical wells pumped approximately 270 acre-ft each, for a total

withdrawal of 1,080 acre-ft. Pumping was nearly continuous during the simulation to extract this volume.

This scenario produced the greatest drawdown of the four pumping scenarios tested (figs. 31, 32). The drawdown at four selected cells along the Big Muddy (cell locations in fig. 31) showed 2.1 ft at cell 80239 near Homestead, with less drawdown at the other cells to the southwest (fig. 32).

## SUMMARY AND CONCLUSIONS

Information compiled and collected during this project improved the understanding of the Clear Lake aquifer hydrogeologic system. Drilling activities added to the monitoring network, which provided data for aquifer mapping. Estimates of irrigation water use were improved by measuring pumping rates with an ultrasonic flowmeter and refining estimates of pumping duration at each irrigation well. The increased accuracy of these measurements likely improved the numeric model calibration and related conclusions about the effects of irrigation water use.

Data collected in the South Medicine Lake focus area resulted in a detailed conceptual model for this segment of the Clear Lake aquifer. The cross sections constructed from borehole analysis defined boundaries of the buried valley aquifer. These boundaries include longitudinal hydraulic barriers formed by bedrock ridges that restrict lateral flow in the aquifer. These barriers appear to eliminate direct hydraulic connection between the aquifer and Medicine Lake.

Findings of the hydrogeologic investigation and the numerical model development allowed for refinement of the conceptual model. Flow in the South Medicine Lake focus area starts at a groundwater divide beneath the sand hills southeast of Medicine Lake. Recharge to this area includes infiltration of precipitation and snowmelt in the sand hills, possible groundwater flow from unmapped gravel deposits, and flow from the Fort Union aquifers. From the divide, groundwater predominately flows southwest and discharges in the Big Muddy Creek Valley. Discharge from the buried valley aquifer is through seepage to wetlands along Big Muddy Creek and through the groundwater system to the lower alluvial zone downgradient from the focus area. Groundwater flow is driven by recharge in the sand hills area, discharge as seepage and ET at the wetlands, irrigation pumping, and interaquifer flow to the southwest.

The first objective of this project was to *Determine how much groundwater withdrawal from the Clear Lake aquifer is possible without adversely affecting groundwater and surface water.*

The model suggests that the maximum volume of groundwater withdrawals from the Clear Lake aquifer without detrimental effects is near. The model scenario simulating the 2015 irrigation season did not indicate appreciable drawdown near Homestead. The model run with existing wells pumping the full allocation volume simulates a decrease in heads up to 0.45 ft near Homestead, and other wetland locations showed declines less than 0.1 ft.

The 2014–2015 water-level trends in both groundwater and surface water show stable conditions and that current pumping has not degraded surface-water resources. The current management approach of using continuous monitoring and collaborative review of new water-use applications by the TAC while allowing relatively slow development has been successful at preventing conflict between stakeholders.

The second objective of this project was to *Determine if water development or other land uses have affected water quality in the Clear Lake aquifer.*

Two likely sources of poor-quality water that could affect the aquifer include: (1) flow from saline lakes into the aquifer induced by pumping associated with increased irrigation development, and (2) leaks and spills associated with hydrocarbon extraction. Water-quality monitoring in the Clear Lake aquifer has not identified increases in aquifer salinity induced by pumping near saline lakes. As of the completion of this investigation, leaks and spills of saline water and hydrocarbons associated with energy development have not degraded water in the Clear Lake aquifer to levels unusable for irrigation. Low-level concentrations of organic compounds likely associated with hydrocarbon extraction have been identified in the Clear Lake aquifer, but are unlikely to affect irrigation supplies owing to the low concentration levels detected (Meredith and Kuzara, 2018).

Areas of poor water quality described in this report result from naturally occurring aquifer conditions and may limit irrigation development in the Clear Lake aquifer. Where coal or sandstone aquifers in the Fort Union bedrock are hydraulically connected to the alluvial aquifer, discharge from the bedrock mixes with

the lower alluvial zone and degrades water quality. Most of the lower alluvial zones of the Clear Lake aquifer have marginal to poor water quality for irrigation use. The best water quality in the lower alluvial zone is located within 6 mi of the sand hills southeast of Medicine Lake (fig. 21B). Outside of this area, development of groundwater for irrigation may encounter elevated SAR.

The upper outwash zone of the Clear Lake aquifer has lower SAR values and better irrigation potential (fig. 21A). Most of these areas have glacial outwash at the surface and are likely recharged by direct infiltration of rainwater and snowmelt. Water quality in several isolated areas are mapped as marginal to poor water quality. Where permeable outwash zones are layered in the subsurface, water quality varies with depth. This is an important consideration for future development, and the limited available data indicate improved water quality in the shallower zones.

## RECOMMENDATIONS

Maintaining the existing network of monitoring wells is critical to track water-level fluctuations throughout the aquifer. Long-term records provide a broader view of aquifer conditions and the data necessary to manage water resources during periods of drought and climate change, as recharge, ET, and water use changes. The continuous water-level records provide comprehensive data needed to evaluate new irrigation development.

Information gained from new irrigation development will include results from new aquifer tests. As new test data become available, the spatial distribution and range of aquifer parameters will be refined. These refinements could be used to update the existing groundwater models.

The model calibration could be improved by including the water-level data collected in 2016 and 2017. There was little water use in 2016 owing to the higher than normal precipitation, which was followed by a year of high water use in the extreme drought of 2017. A 3-yr transient calibration would extend the predictive power of the current model, and would help define the role of local recharge given the extreme changes between 2016 and 2017.

The connections between the Clear Lake aquifer and the Big Muddy Creek Valley wetlands are poorly

understood, and there is little monitoring of water levels or quality. Modeling the connections between the aquifer and surface-water features will require additional investigation.

Investigating the vertical variability of aquifer yield and water quality may identify additional productive zones of high-quality water within the aquifer. In the past, wells were drilled into the deepest, most productive aquifer zone, but better water quality is often encountered in shallower layers. These shallow zones may show development potential when isolated from deeper poor-water-quality zones.

New technology advances have shown that aerial electromagnetic surveys (AEM) are better and less expensive at defining buried valley aquifers than exploration drilling. Recent work in the Canadian prairies, North Dakota, and Nebraska have demonstrated the value of this geophysical tool. An AEM survey could potentially identify additional productive areas of the Clear Lake aquifer.

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**APPENDIX A**  
**CLEAR LAKE AQUIFER GROUNDWATER**  
**MONITORING NETWORK**

GWIC ID	Latitude	Longitude	Geomethod	Datum	Ground Surface Altitude (ft amsl)	Twn	Rng	Sec	Quarter Section	Total Depth (ft)	Date Completed m/d/yr	Aquifer Code*	First Date Measured (m/d/yr)	Last Date Measured (m/d/yr)	No. of Readings
3605	48.38959	-104.54132	MAP	NAD83	1945.94	30N	55E	3	AAAA	185	10/12/1984	112ALVM	11/30/1984	8/4/2016	51066
3606	48.36071	-104.54148	NAV-GPS	NAD83	1973.65	30N	55E	10	DDDD	147	10/11/1984	112OTSH	11/30/1984	8/4/2016	243
3607	48.36020	-104.58270	UNKNOWN	NAD27	1942	30N	55E	16	BBBA	172	11/20/1983	112ALVM	8/19/1985	4/21/1998	48
3613	48.38880	-104.25861	NAV-GPS	WGS84	2230	30N	57E	1	BBB	162	NA	125FRUN	7/26/2016	7/26/2016	1
3677	48.38955	-104.45278	NAV-GPS	WGS84	2107.9	31N	56E	33	CCCD	350	9/22/1984	112ALVM	10/11/1984	10/1/2016	54597
3678	48.47396	-104.25133	NAV-GPS	WGS84	1991.63	31N	57E	1	BACA	125	NA	112ALVM	7/19/1983	10/17/2016	203
3679	48.46017	-104.31014	NAV-GPS	NAD83	1969.69	31N	57E	9	ABAD	220	6/22/1983	112ALVM	7/18/1983	10/15/2016	125
3680	48.40629	-104.30985	NAV-GPS	WGS84	2105	31N	57E	28	ABDA	230	NA	112ALVM	7/26/2016	7/26/2016	1
3682	48.39012	-104.29468	NAV-GPS	WGS84	2209	31N	57E	34	CDDC	90	1/1/1909	112OTSH	1/9/1999	3/19/1999	155
3762	48.49138	-104.50619	NAV-GPS	NAD83	1951.06	32N	55E	25	DCCD	200	6/27/1987	112ALVM	1/1/1900	8/4/2016	6195
3763	48.49138	-104.50619	NAV-GPS	NAD83	1951.15	32N	55E	25	DCCD	200	6/27/1987	112OTSH	5/9/1995	8/4/2016	21984
3764	48.49138	-104.50619	NAV-GPS	NAD83	1951.15	32N	55E	25	DCCD	200	6/27/1987	112ALVM	5/9/1995	8/4/2016	6640
3765	48.50580	-104.47440	MAP	NAD27	1985	32N	56E	20	CCCC	240	4/2/1975	112ALVM	6/4/2014	6/12/2014	150
3766	48.49082	-104.45425	NAV-GPS	NAD83	1986.76	32N	56E	29	DDDD	233	9/19/1984	112ALVM	10/9/1984	10/15/2016	81093
3767	48.49082	-104.45426	NAV-GPS	NAD83	1986.54	32N	56E	29	DDDD	118	9/19/1984	112OTSH	10/11/1984	10/15/2016	61478
3769	48.49788	-104.28109	NAV-GPS	NAD83	1950.49	32N	57E	26	CBBB	270	7/17/1984	112ALVM	7/18/1984	10/17/2016	40517
3770	48.49775	-104.28107	NAV-GPS	NAD83	1950.19	32N	57E	26	CBBB	140	7/17/1984	112OTSH	8/10/1984	10/17/2016	21282
3771	48.48874	-104.28746	NAV-GPS	NAD83	1947.99	32N	57E	34	ABDC	100	6/24/1988	112OTSH	8/15/1995	10/17/2016	41
3772	48.55496	-104.18183	NAV-GPS	NAD83	1976.85	32N	58E	4	DBBD	318	7/19/1984	125FRUN	7/26/1984	10/9/2016	218
3773	48.55502	-104.18186	NAV-GPS	NAD83	1977	32N	58E	4	DBBD	143	7/25/1984	112OTSH	7/26/1984	10/9/2016	42115
3774	48.53424	-104.17277	NAV-GPS	NAD83	2006.16	32N	58E	10	CCCC	210	8/28/1984	112ALVM	11/30/1984	10/19/2016	10349
3776	48.52110	-104.10940	MAP	NAD27	2040	32N	58E	13	DDDB	295	4/22/1961	112ALVM	1/14/2000	2/14/2000	66
3777	48.53376	-104.20560	NAV-GPS	NAD83	1959.23	32N	58E	17	ABBB	80	8/23/1986	112OTSH	6/5/1995	10/19/2016	14471
3779	48.49806	-104.17297	NAV-GPS	NAD83	2012.8	32N	58E	27	BCCC	270	9/18/1984	112ALVM	11/30/1984	10/19/2016	6268
3854	48.56920	-104.19780	NAV-GPS	NAD27	2001	33N	57E	36	CAAA	140	7/30/1985	112OTSH	7/30/1985	5/19/2001	74
3858	48.61384	-104.14441	NAV-GPS	NAD83	1994.04	33N	58E	17	ADDD	130	7/30/1984	112OTSH	8/9/1984	9/19/2016	56588
3861	48.60401	-104.10028	NAV-GPS	NAD83	2011.09	33N	58E	23	BBCB	137	8/29/1985	112OTSH	9/11/1985	9/19/2016	764
3862	48.59823	-104.09945	NAV-GPS	NAD83	2010.47	33N	58E	23	CBBC	150	8/1/1984	112OTSH	11/6/1982	9/19/2016	3585
3866	48.60633	-104.04819	NAV-GPS	NAD83	2012.2	33N	58E	24	AAAA	238	8/21/1984	112OTSH	11/30/1984	9/19/2016	241
3867	48.59160	-104.04800	MAP	NAD27	1990.9	33N	58E	25	AAAA	336	8/24/1984	112ALVM	6/9/1985	10/8/1987	20
3868	48.59207	-104.04802	NAV-GPS	NAD83	1989.4	33N	58E	25	AAAA	150	8/24/1984	112OTSH	11/30/1984	8/11/2016	41843
3869	48.58408	-104.14289	NAV-GPS	NAD83	1990.69	33N	58E	28	CBBA	130	7/30/1984	112OTSH	5/21/1985	9/17/2016	227
3870	48.58656	-104.14710	NAV-GPS	NAD83	1992.08	33N	58E	29	ACDA	100	9/17/1980	112OTSH	11/6/1982	10/20/2016	56
3871	48.58553	-104.14540	NAV-GPS	NAD83	1993.13	33N	58E	29	ADDB	98	7/25/1984	112OTSH	7/26/1984	10/20/2016	461
3872	48.58555	-104.14546	NAV-GPS	NAD83	1992.84	33N	58E	29	ADDB	330	7/30/1984	125FRUN	8/8/1984	10/20/2016	261
3873	48.58856	-104.17115	NAV-GPS	NAD83	1994.44	33N	58E	30	ACAA	150	9/18/1980	112OTSH	5/9/1995	10/20/2016	44
3944	48.71518	-104.08545	NAV-GPS	NAD83	2057.39	34N	58E	11	DBAB	130	8/7/1984	112OTSH	8/8/1984	4/13/2016	16499
3945	48.70438	-104.07265	MAP	NAD83	2065.62	34N	58E	13	BBDD	153	11/12/1980	112OTSH	4/1/1981	4/20/2016	58
3946	48.69723	-104.07247	MAP	NAD83	2092.48	34N	58E	13	CDBB	154	NA	112OTSH	4/1/1981	4/20/2016	58
3947	48.68554	-104.14395	NAV-GPS	NAD83	2041.33	34N	58E	20	DAAA	110	8/10/1984	112OTSH	8/11/1984	9/19/2016	5054
3948	48.69160	-104.10529	MAP	NAD83	2044.11	34N	58E	22	AACC	104	11/5/1980	112OTSH	4/1/1981	4/20/2016	59
3949	48.69338	-104.09972	MAP	NAD83	2049.61	34N	58E	23	BBBB	90	11/9/1980	112OTSH	4/1/1981	4/14/2016	58
4048	48.82194	-104.04952	NAV-GPS	NAD83	2076.2	35N	58E	1	AADA	45	NA	112OTSH	5/9/1995	4/14/2016	42
42303	48.36305	-104.61023	TRS-SEC	NAD83	1926	30N	55E	7	DDB	95	8/5/1980	112ALVM	5/8/2012	5/9/2012	89
43113	48.42142	-104.37297	NAV-GPS	NAD83	2013.4	31N	56E	24	DDBB	160	3/5/1990	112ALVM	5/9/1995	10/13/2016	69622

GWIC ID	Latitude	Longitude	Geomethod	Datum	Ground Surface Altitude (ft amsl)	Twn	Rng	Sec	Quarter Section	Total Depth (ft)	Date Completed m/d/yr	Aquifer Code*	First Date Measured (m/d/yr)	Last Date Measured (m/d/yr)	No. of Readings
43116	48.40409	-104.43808	NAV-GPS	NAD83	2088.99	31N	56E	28	DDCC	253	3/28/1990	112ALVM	5/9/1995	10/12/2016	65451
43124	48.46831	-104.25584	TRS-SEC	NAD83	2015	31N	57E	1	CBA	88	11/11/1961	112ALVM	8/18/2004	8/18/2004	1
43144	48.41494	-104.32505	NAV-GPS	WGS84	2099	31N	57E	29	ADAA	170	4/24/1958	112SNGR	7/27/2016	7/27/2016	1
43145	48.15773	-104.35419	NAV-GPS	WGS84	2032	31N	57E	30	ABDB	210	3/4/1988	112ALVM	10/21/2015	10/13/2016	7
44469	48.53786	-104.16732	NAV-GPS	NAD83	2012.16	32N	58E	10	CCAA	131	9/17/1984	112OTSH	6/4/1996	10/19/2016	29231
44482	48.53385	-104.20577	NAV-GPS	NAD83	1961.69	32N	58E	17	ABBB	146	8/21/1986	112OTSH	6/13/1995	10/19/2016	200
44484	48.53386	-104.20577	NAV-GPS	NAD83	1961.58	32N	58E	17	ABBB	146	8/21/1986	112OTSH	6/13/1995	10/19/2016	198
44497	48.48055	-104.23389	NAV-GPS	NAD83	1978.3	32N	58E	31	CBDC	270	6/20/1988	112ALVM	5/7/1996	10/17/2016	191
44498	48.48240	-104.22572	NAV-GPS	NAD83	1976.34	32N	58E	31	DBBD	271	8/27/1988	112ALVM	10/19/1995	10/17/2016	42
45291	48.56800	-104.20220	NAV-GPS	NAD27	2000	33N	57E	36	CACA	80	3/7/1990	112OTSH	5/11/1995	9/26/2001	55
45292	48.56797	-104.19496	NAV-GPS	NAD83	1986.4	33N	57E	36	DBDB	100	3/8/1990	112ALVM	5/11/1995	10/19/2016	17892
45325	48.58122	-104.18204	MAP	NAD83	1989.23	33N	58E	30	CBDD	120	8/1/1988	112OTSH	5/9/1995	10/20/2016	44
45336	48.60170	-104.09913	TRS-SEC	NAD83	1982	33N	58E	23	BCBD	147	8/30/1985	112ALVM	9/11/1985	9/11/1985	1
45977	48.70631	-104.04834	NAV-GPS	NAD83	2093.25	34N	58E	13	AAAD	180	8/8/1984	112ALVM	8/11/1984	9/19/2016	28932
45978	48.70610	-104.04837	NAV-GPS	NAD83	2094.76	34N	58E	13	AAAD	105	8/9/1984	112OTSH	10/29/1984	9/19/2016	230
45980	48.70431	-104.08356	MAP	NAD83	2060.24	34N	58E	14	ACAA	126	10/2/1980	112OTSH	5/9/1995	4/14/2016	44
45981	48.69721	-104.09453	MAP	NAD83	2056.53	34N	58E	14	CBDD	150	5/6/1977	112OTSH	11/5/1982	5/1/2016	65
45982	48.69735	-104.09486	NAV-GPS	NAD83	2054.9	34N	58E	14	CBDD	160	12/1/1976	112OTSH	5/9/1995	10/22/2015	60499
45983	48.69711	-104.08363	MAP	NAD83	2060.51	34N	58E	14	DBDD	95	6/26/1980	112OTSH	5/9/1995	4/14/2016	43
46003	48.66799	-104.15805	MAP	NAD83	2034.9	34N	58E	29	CDBA	81	10/23/1980	112OTSH	11/6/1982	4/20/2016	60
46960	48.81598	-104.05489	TRS-SEC	NAD83	2071	35N	58E	1	DAB	45	10/18/1982	112OTSH	10/20/2005	4/14/2016	13
46965	48.81575	-104.05469	MAP	NAD83	2068.8	35N	58E	1	DBDD	40	10/18/1982	112OTSH	4/8/1983	4/14/2016	59
121117	48.42134	-104.37299	NAV-GPS	NAD83	2007	31N	56E	24	DDBB	155	8/15/1990	112ALVM	10/19/1995	10/13/2016	42
121781	48.46745	-104.26332	NAV-GPS	NAD83	2026.41	31N	57E	2	DAC	130	9/7/1990	112ALVM	7/11/1996	10/17/2016	176
135910	48.49137	-104.50671	TRS-SEC	NAD83	1952.55	32N	55E	25	DCCD	220	12/10/1992	112ALVM	5/9/1995	10/19/1995	2
148578	48.60575	-104.10702	NAV-GPS	NAD83	1960.1	33N	58E	22	ABAA	6	6/1/1989	112OTSH	8/8/1990	9/19/2016	68591
149003	48.57771	-104.19310	MAP	NAD83	2022.1	33N	57E	25	DCCC	120	2/14/1995	112OTSH	5/9/1995	10/20/2016	36
149610	48.78025	-104.04866	NAV-GPS	NAD83	2093.1	35N	58E	24	AAAA	92.5	5/12/1982	112OTSH	6/8/1982	9/19/2016	232
150614	48.48351	-104.29082	NAV-GPS	NAD83	1951.73	32N	57E	34	DBBB	68	NA	112OTSH	5/9/1995	10/17/2016	42
150615	48.48481	-104.26836	NAV-GPS	NAD83	1961.4	32N	57E	35	ACCBA	260	NA	112ALVM	8/15/1995	10/17/2016	162
150769	48.55256	-104.22214	LORAN-C	NAD83	2002	32N	58E	6	DBDD	100	NA	112OTSH	5/9/1995	10/16/2007	25
150940	48.42164	-104.38857	NAV-GPS	NAD83	2013.84	31N	56E	24	CBB	150	2/4/1995	112ALVM	5/11/1995	10/12/2016	92800
152253	48.57369	-104.17279	NAV-GPS	NAD83	1967	33N	58E	31	ACAB	5	6/1/1995	112OTSH	6/20/1995	10/19/2016	34270
152254	48.40416	-104.43792	NAV-GPS	NAD83	2089.49	31N	56E	28	DDCC	244	7/27/1990	112ALVM	5/9/1995	10/12/2016	43
153314	48.59577	-104.13847	MAP	NAD83	1997.97	33N	58E	21	CCAA	110	NA	112OTSH	6/9/1982	10/20/2016	48
154292	48.80007	-104.04857	NAV-GPS	NAD83	2093.8	35N	58E	12	DAD	180	10/13/1981	112OTSH	2/2/1996	9/19/2016	186
154904	48.52152	-104.50057	TRS-SEC	NAD83	1965	32N	55E	13	DD	840	6/1/1995	211FHHC	6/18/2013	6/18/2013	1
155430	48.56689	-104.16035	NAV-GPS	NAD83	1987.62	33N	58E	32	CACC	91	NA	112OTSH	8/16/1982	10/19/2016	66
155431	48.66799	-104.15801	MAP	NAD83	2034.9	34N	58E	29	CDBA	81	NA	112OTSH	5/9/1995	9/19/2016	40
155916	48.69728	-104.10583	NAV-GPS	NAD83	2054.6	34N	58E	15	DACC	160	6/19/1996	112OTSH	6/19/1996	4/23/2016	55209
155919	48.70433	-104.10619	NAV-GPS	NAD83	2052.69	34N	58E	15	ABDD	160	6/19/1996	112OTSH	6/19/1996	4/23/2016	36008
155920	48.48337	-104.29126	NAV-GPS	NAD83	1947	32N	57E	34	DBBB	69	6/17/1996	112OTSH	7/17/1996	10/17/2016	52314
155921	48.58843	-104.18214	NAV-GPS	NAD83	2018.5	33N	58E	30	BACC	115	6/17/1996	112OTSH	6/4/1997	10/20/2016	163
155929	48.41896	-104.38884	NAV-GPS	NAD83	2029.74	31N	56E	24	CCCC	180	6/7/1996	112ALVM	6/4/1996	7/5/2016	33709
155930	48.41736	-104.39052	NAV-GPS	NAD83	2025.9	31N	56E	26	ABDD	180	4/5/1996	112ALVM	10/25/1996	10/13/2016	37

GWIC ID	Latitude	Longitude	Geomethod	Datum	Ground Surface Altitude (ft amsl)	TwN	Rng	Sec	Quarter Section	Total Depth (ft)	Date Completed m/d/yr	Aquifer Code*	First Date Measured (m/d/yr)	Last Date Measured (m/d/yr)	No. of Readings
155931	48.48496	-104.27037	NAV-GPS	NAD83	1961	32N	57E	35	BCDD	260	5/11/1996	112ALVM	6/4/1996	10/17/2016	32662
155933	48.48512	-104.27018	NAV-GPS	NAD83	1961	32N	57E	35	BCDD	260	5/31/1996	112ALVM	5/31/1996	10/17/2016	40
157817	48.50183	-104.29233	NAV-GPS	NAD83	1943.67	32N	57E	27	ABAA	7	NA	112OTSH	5/15/1996	10/15/2016	47542
159786	48.48339	-104.30252	NAV-GPS	NAD83	1959.03	32N	57E	34	CBBB	160	9/24/1996	112OTSH	6/3/1997	10/15/2016	2959
159791	48.45504	-104.31903	TRS-SEC	NAD83		31N	57E	9	BDCC	160	9/26/1996	112OTSH	7/22/1997	4/7/1998	10
159795	48.53778	-104.16746	NAV-GPS	NAD83	2012.42	32N	58E	10	CCAA	150	5/17/1986	112OTSH	6/4/1996	11/1/2012	29
160086	48.58499	-104.17939	MAP	NAD83	1990.31	33N	58E	30	BDDD	132	11/29/1996	112OTSH	11/7/1996	10/20/2016	41
161780	48.48532	-104.30012	NAV-GPS	NAD83	1940.09	32N	57E	34	BCBD	8.6	NA	112OTSH	5/30/1997	9/17/2016	63811
161781	48.50532	-104.26018	NAV-GPS	NAD83	1949.83	32N	57E	23	DDDA	8.6	5/28/1997	112OTSH	5/28/1997	10/17/2016	48541
161782	48.57331	-104.13838	NAV-GPS	NAD83	1957	33N	58E	33	BDBC	10	5/28/1997	112OTSH	6/3/1997	10/20/2016	52011
161783	48.56249	-104.18696	NAV-GPS	NAD83	1960.48	32N	58E	4	BAAB	5	5/28/1997	112OTSH	5/28/1997	10/19/2016	43747
161784	48.72023	-104.08409	NAV-GPS	NAD83	2040.7	34N	58E	11	ABDA	8.3	5/29/1997	112OTSH	5/29/1997	10/22/2015	46584
161785	48.73691	-104.08469	NAV-GPS	NAD83	2045	34N	58E	2	ABAA	8.8	5/29/1997	112OTSH	5/29/1997	10/22/2015	45916
161786	48.79906	-104.05065	NAV-GPS	NAD83	2052.4	35N	58E	12	DADC	9.2	5/29/1997	112OTSH	6/17/1997	8/20/2016	49450
161787	48.43178	-104.38128	NAV-GPS	NAD83	1951.38	31N	56E	24	BBDA	10	5/30/1997	112OTSH	5/30/1997	10/21/2015	38561
161788	48.43128	-104.38802	NAV-GPS	NAD83	1946.78	31N	56E	24	BBCB	10	5/30/1997	112OTSH	6/20/1997	10/13/2016	164
161792	48.57803	-104.18350	TRS-SEC	NAD83	1982	33N	58E	30	CCDD	118	4/7/1997	112OTSH	4/29/1997	10/20/2016	38
161801	48.57757	-104.18137	NAV-GPS	NAD83	1986.51	33N	58E	30	CDD	120	4/5/1997	112OTSH	5/8/1997	10/19/2016	15213
161802	48.74050	-104.06180	NAV-GPS	NAD83	2065.72	35N	58E	36	DBB	90	11/12/1996	112OTSH	10/10/2003	9/19/2016	2513
168129	48.71896	-104.07255	NAV-GPS	NAD83	2066.63	34N	58E	12	BDDD	140	4/27/1998	112OTSH	7/24/1998	9/19/2016	3055
168130	48.42194	-104.40010	NAV-GPS	NAD83	1995.6	31N	56E	23	CDAA	160	2/5/1998	112ALVM	4/7/1998	10/12/2016	91365
168131	48.41981	-104.41144	NAV-GPS	NAD83	1994.52	31N	56E	22	DDDA	160	2/6/1998	112ALVM	4/7/1998	10/12/2016	65505
168132	48.42030	-104.50760	NAV-GPS	NAD27	1978	31N	55E	24	DCBC	240	NA	112ALVM	4/2/1998	4/23/2004	23
168866	48.45911	-104.31505	NAV-GPS	NAD83	1970.62	31N	57E	9	BC	240	4/10/1998	112ALVM	8/4/1998	10/15/2016	37
168867	48.45591	-104.31508	NAV-GPS	NAD83	1974.63	31N	57E	9	BB	240	4/11/1998	112ALVM	5/6/1998	10/13/2016	50982
169212	48.44777	-104.45417	NAV-GPS	NAD83	2027.4	31N	56E	9	CCCC	108	4/19/1999	112ALVM	5/17/1999	10/12/2016	158
169213	48.43284	-104.43302	NAV-GPS	NAD83	2057.55	31N	56E	21	AAAA	130	4/20/1999	112ALVM	5/17/1999	10/12/2016	53071
169214	48.42557	-104.45436	NAV-GPS	NAD83	2091.14	31N	56E	20	DAAA	160	4/20/1999	112ALVM	5/17/1999	10/12/2016	27543
169215	48.41124	-104.47590	NAV-GPS	NAD83	2058.69	31N	56E	29	BCCC	190	4/21/1999	112ALVM	5/17/1999	10/1/2016	54734
169217	48.40406	-104.41092	NAV-GPS	NAD83	2058.11	31N	56E	27	DDDD	227	4/22/1999	112ALVM	5/17/1999	10/12/2016	120401
169218	48.41118	-104.41090	NAV-GPS	NAD83	2054.23	31N	56E	27	DAAA	224	4/22/1999	112ALVM	4/27/1999	10/12/2016	92363
169219	48.43462	-104.36716	NAV-GPS	NAD83	1973.54	31N	57E	18	CCCB	80	4/23/1999	112OTSH	5/17/1999	8/13/2014	83
169220	48.44792	-104.34620	NAV-GPS	NAD83	1972.22	31N	57E	8	CCCC	135	4/23/1999	112OTSH	6/17/1999	10/22/2013	76
169372	48.72187	-104.08309	NAV-GPS	NAD83	2050.6	34N	58E	11	AABB	220	3/17/1999	112OTSH	5/18/1999	7/23/2016	8808
169373	48.72200	-104.08313	NAV-GPS	NAD83	2048.3	34N	58E	11	AABB	30	3/19/1999	112OTSH	4/20/1999	7/23/2016	1230
169374	48.73665	-104.08366	NAV-GPS	NAD83	2050	34N	58E	2	ABAA	200	3/19/1999	112OTSH	4/20/1999	7/23/2016	1155
171418	48.69772	-104.10579	NAV-GPS	NAD83	2055.14	34N	58E	15	DBDD	130	4/27/1999	112OTSH	10/24/2005	4/20/2016	21
175651	48.40411	-104.43792	NAV-GPS	NAD83	2089.88	31N	56E	28	DDC	246	8/26/1999	112ALVM	4/17/2001	4/21/2016	30
184125	48.37863	-104.53640	NAV-GPS	NAD83	1956.31	30N	55E	2	CAA	184	2/11/1998	112OTSH	8/1/2000	7/5/2016	8898
193352	48.71881	-104.07310	NAV-GPS	NAD83	2069	34N	58E	12	B	108	10/11/2001	112OTSH	10/20/2005	4/13/2016	24
193370	48.70441	-104.10550	MAP	NAD83	2053.74	34N	58E	15	A	142	10/5/2001	112OTSH	4/27/2006	4/20/2016	22
206410	48.40052	-104.52432	NAV-GPS	NAD83	1954	31N	55E	35	AACA	180	4/26/1998	112ALVM	10/26/2004	10/31/2012	15
206533	48.78048	-104.06338	NAV-GPS	NAD83	2070.3	35N	58E	13	CCDC	80	9/15/2003	112OTSH	9/17/2003	9/20/2016	96
206537	48.78051	-104.07771	NAV-GPS	NAD83	2066.5	35N	58E	13	CCCC	100	9/16/2003	112OTSH	9/16/2003	8/20/2016	43041
206538	48.78050	-104.08172	NAV-GPS	NAD83	2071.8	35N	58E	14	DDCC	125	9/16/2003	112OTSH	9/17/2003	9/19/2016	92

GWIC ID	Latitude	Longitude	Geomethod	Datum	Ground Surface Altitude (ft amsl)	Twn	Rng	Sec	Quarter Section	Total Depth (ft)	Date Completed m/d/yr	Aquifer Code*	First Date Measured (m/d/yr)	Last Date Measured (m/d/yr)	No. of Readings
206539	48.75147	-104.04999	NAV-GPS	NAD83	2052	35N	58E	25	DDDD	180	9/16/2003	112OTSH	9/17/2003	9/19/2016	99
206541	48.83900	-104.06100	NAV-GPS	NAD27	2060	36N	58E	25	DCCC	133	9/15/2003	112OTSH	9/15/2003	9/15/2003	1
206546	48.75152	-104.07788	NAV-GPS	NAD83	2064.3	35N	58E	25	CCCC	120	9/18/2003	112OTSH	9/18/2003	9/20/2016	13905
206547	48.54852	-104.15179	NAV-GPS	NAD83	1984.14	32N	58E	10	AAAA	200	9/18/2003	112OTSH	9/18/2003	10/19/2016	9843
206548	48.55929	-104.13953	NAV-GPS	NAD83	1993.68	32N	58E	2	ABCC	180	9/19/2003	112OTSH	9/19/2003	10/19/2016	8594
212128	48.40471	-104.41003	NAV-GPS	NAD83	2056.85	31N	56E	27	DDDD	240	3/27/2004	112ALVM	4/8/2004	10/12/2016	381
215189	48.75153	-104.08830	NAV-GPS	NAD83	2052.8	35N	58E	26	DCCC	175	11/7/2004	112ALVM	12/3/2004	9/19/2016	82
215220	48.69356	-104.12128	NAV-GPS	NAD27	2047.05	34N	58E	15	CCCC	140	11/7/2004	112OTSH	12/3/2004	9/20/2016	85
215223	48.67910	-104.09940	NAV-GPS	WGS84	2061.7	34N	58E	23	CCCC	140	11/8/2004	112OTSH	12/3/2004	9/20/2016	81
221597	48.62120	-104.12170	NAV-GPS	NAD83	2009.8	33N	58E	10	CCCC	128	9/12/2005	112OTSH	9/18/2005	9/20/2016	73
221602	48.62140	-104.15120	NAV-GPS	NAD83	2008.49	33N	58E	8	DDCC	108	9/13/2005	112OTSH	9/18/2005	9/20/2016	49395
221606	48.62121	-104.16607	NAV-GPS	NAD83	2055.08	33N	58E	8	DDDD	88	9/13/2005	112ALVM	9/18/2005	9/19/2016	73
221614	48.49810	-104.22950	NAV-GPS	NAD83	1979.1	32N	58E	30	BDDC	288	9/14/2005	112ALVM	9/18/2005	10/17/2016	37262
221630	48.49818	-104.22954	NAV-GPS	NAD83	1979.1	32N	58E	30	BDDC	168	9/14/2005	112OTSH	9/18/2005	10/17/2016	44799
221634	48.39805	-104.41073	NAV-GPS	NAD83	2100.03	31N	56E	35	BCCB	278	9/15/2005	112ALVM	9/20/2005	10/12/2016	72
221637	48.39378	-104.41078	NAV-GPS	NAD83	2093.13	31N	56E	35	CBBC	288	9/16/2005	112ALVM	9/19/2005	10/12/2016	61602
221646	48.41830	-104.36772	NAV-GPS	NAD83	2029.1	31N	57E	30	BBBB	68	9/20/2005	112OTSH	9/20/2005	10/13/2016	22045
221647	48.40400	-104.35750	NAV-GPS	NAD83	2044.13	31N	57E	30	CDDD	195	9/19/2005	112ALVM	9/20/2005	10/13/2016	55020
221649	48.49065	-104.41029	NAV-GPS	NAD83	1983.41	32N	56E	35	BBBB	168	9/22/2005	112OTSH	11/10/2005	10/15/2016	50878
221651	48.49065	-104.41037	NAV-GPS	NAD83	1983.6	32N	56E	35	BBBB	248	9/21/2005	112ALVM	9/22/2005	10/15/2016	48345
221652	48.41830	-104.35762	NAV-GPS	NAD83	2022.41	31N	57E	30	BAAA	180	9/21/2005	112ALVM	11/8/2005	7/26/2016	64046
222551	48.69308	-104.09996	TRS-SEC	NAD83	2049.81	34N	58E	23	BBBB	95	5/11/2005	112OTSH	4/25/2006	4/14/2016	22
223319	48.74060	-104.06160	NAV-GPS	NAD83	2067.69	35N	58E	36	DDBB	180	12/5/2005	112OTSH	4/25/2006	4/13/2016	19
223321	48.81505	-104.05506	NAV-GPS	WGS84	2068.6	35N	58E	1	ACCC	40	12/12/2005	112OTSH	4/25/2006	4/14/2016	18
235004	48.75516	-104.07262	NAV-GPS	NAD83	2085.18	35N	58E	25	C	170	3/13/2007	112OTSH	5/22/2009	8/20/2016	31194
242569	48.54974	-104.21745	NAV-GPS	NAD83	1982.48	32N	58E	5	DDDD	100	4/10/2008	112OTSH	10/29/2008	10/19/2016	18
247227	48.69718	-104.07322	MAP	NAD83	2096.5	34N	58E	13	CDBB	150	NA	112OTSH	7/23/2008	4/20/2016	393
250391	48.75877	-104.07110	NAV-GPS	NAD83	2084.43	35N	58E	25	CAB	150	4/1/2009	112OTSH	5/22/2009	10/30/2012	8
250399	48.71494	-104.08201	NAV-GPS	NAD83	2056.16	34N	58E	11	DAB	140	4/9/2009	112OTSH	4/27/2009	10/22/2015	10
250466	48.75513	-104.07211	NAV-GPS	NAD83	2083.06	35N	58E	25	CCA	160	3/26/2009	112OTSH	5/22/2009	10/22/2015	13
252391	48.75873	-104.06383	TRS-SEC	NAD83	2083.73	35N	58E	25	DBB	130	4/9/2009	112OTSH	5/22/2009	8/20/2016	57740
252392	48.71063	-104.08316	NAV-GPS	NAD83	2058.12	34N	58E	11	D	100	9/30/2009	112OTSH	11/4/2009	8/20/2016	80149
263859	48.49147	-104.50597	NAV-GPS	NAD83	1952.3	32N	55E	25	DCCD	72	NA	112OTSH	5/9/1995	8/4/2016	53
267045	48.36127	-104.60488	NAV-GPS	NAD83	1930	30N	55E	8	CCBD	100	NA	112ALVM	5/8/2012	5/9/2012	93
268466	48.36119	-104.60484	TRS-SEC	NAD83	1930	30N	55E	8	CCC	105	5/1/2012	112ALVM	5/8/2012	5/9/2012	103
273937	48.60614	-104.18503	NAV-GPS	NAD83	2019.99	33N	58E	19	BBBA	136	6/4/2013	112OTSH	6/4/2013	10/20/2016	15900
280602	48.40408	-104.36800	NAV-GPS	WGS84	2047.04	31N	56E	25	DDDD	203	10/15/2014	112ALVM	10/15/2014	10/13/2016	10100
280612	48.41113	-104.36776	NAV-GPS	WGS84	2043	31N	56E	25	DAAA	220	10/16/2014	112ALVM	10/16/2014	10/13/2016	14854
280618	48.42514	-104.38897	NAV-GPS	WGS84	1984.82	31N	56E	24	CBBB	140	10/17/2014	112ALVM	11/24/2014	10/12/2016	10592
280621	48.42556	-104.42151	MAP	WGS84	2001	31N	56E	22	CBBB	59	10/18/2014	112ALVM	10/18/2014	10/12/2016	15650
280641	48.48351	-104.41044	NAV-GPS	WGS84	1959.55	32N	56E	35	BCCC	140	10/20/2014	112OTSH	10/20/2014	10/15/2016	6462
280643	48.48396	-104.35423	NAV-GPS	WGS84	1957.7	32N	57E	31	ACCD	248	10/21/2013	112OTSH	10/20/2014	9/19/2016	10260
280645	48.50608	-104.31126	NAV-GPS	WGS84	1961.32	32N	57E	21	DCCD	120	10/22/2014	112OTSH	10/22/2014	10/15/2016	479
280650	48.52396	-104.26361	NAV-GPS	WGS84	1973.71	32N	57E	14	DACD	90	10/22/2014	112OTSH	10/22/2014	9/20/2016	7899
280651	48.55559	-104.13442	NAV-GPS	WGS84	1991.65	32N	58E	2	DABB	105	10/23/2014	112OTSH	10/23/2014	3/29/2016	7592

GWIC ID	Latitude	Longitude	Geomethod	Datum	Ground Surface Altitude (ft amsl)				Quarter Section	Total Depth (ft)	Date Completed m/d/yr	Aquifer Code*	First Date Measured (m/d/yr)	Last Date Measured (m/d/yr)	No. of Readings
					TwN	Rng	Sec								
280652	48.49110	-104.43201	NAV-GPS	WGS84	1977.88	32N	56E	28	DDDD	140	10/23/2014	112OTSH	11/23/2014	10/15/2016	10082
283920	48.49167	-104.51806	NAV-GPS	WGS84		32N	55E	25	CCC	64	7/16/2014	112OTSH	8/18/2015	9/14/2015	2
703881	48.55179	-104.19691	NAV-GPS	NAD83	1973.22	32N	58E	5	DDAB	112	9/13/1980	112OTSH	5/9/1995	10/19/2016	42
703926	48.70161	-104.08961	MAP	NAD83	2054.52	34N	58E	14	BDDD	90	1/1/1974	112OTSH	5/9/1995	4/20/2016	43
703929	48.66817	-104.15251	MAP	NAD83	2040	34N	58E	29	DCBA	122	9/10/1981	112OTSH	5/9/1995	4/20/2016	42
703974	48.81774	-104.05506	NAV-GPS	NAD83	2071.1	35N	58E	1	ACDA	32	NA	112OTSH	5/10/1995	9/19/2016	184
703976	48.81778	-104.04941	MAP	NAD83	2080.8	35N	58E	1	ADDA	48	NA	112OTSH	5/9/1995	4/14/2016	40
703979	48.76604	-104.04861	NAV-GPS	NAD83	2064.9	35N	58E	24	DDDD	73	5/25/1982	112OTSH	6/8/1982	9/19/2016	231

*AQUIFER CODE	DESCRIPTION
112OTSH	PLEISTOCENE OUTWASH
112ALVM	PLEISTOCENE ALLUVIUM
125FRUN	FORT UNION FM
211FHHC	FOX HILLS–HELL CREEK FM

Note. The aquifer codes were assigned based on well depth, lithologies encountered, and geologic setting at each well.

**APPENDIX B**  
**WATER-QUALITY DATA**

GWIC ID	Sample Number	Site Type	Depth (ft)	Sample Date	Water Temp °C	Lab pH	Lab SC (µS/cm)	Ca	Mg	Na	K	Fe	Mn	SiO <sub>2</sub>	HCO <sub>3</sub>	SO <sub>4</sub>	Cl	NO <sub>3</sub> -N	F	TDS	SAR unit-less	Contam. Index unit less
<b>Water quality of the alluvial aquifer</b>																						
3605	1984Q1146	WELL	185	10/15/1984	8.7	7.95	1,991	102	60	306	5.7	0.2	0.07	23.1	941	396	5	0.02	0.2	1,361	6.0	0.002
3605	206981	WELL	185	6/13/2014	9.2	8.07	1,994	101	60	320	5.5	0.9	0.06	21.3	931	437	6	<0.010	0.62	1,410	6.2	0.003
3607	1985Q0824	WELL	172	7/8/1985	7.75	2,431	2,431	106	63	437	6.1	1.0	0.09	27.3	1069	569	9	0.11	0.7	1,746	8.3	0.004
3607	1996Q0267	WELL	172	8/14/1995	12.6	7.94	2,260	111	64	436	5.8	3.2	0.04	26.4	1074	550	10	0.05	0.2	1,736	8.1	0.004
3608	1985Q0437	WELL	160	5/27/1985	9.2	7.45	2,040	180	65	272	6.9	5.4	0.54	30.9	776	606	8	0.6	0.3	1,538	4.6	0.004
3609	1984Q0036	WELL	85	4/2/1984	7.38	2,020	2,020	153	64	250	6.9	4.6	0.51	27.4	748	571	7	0.67	0.2	1,454	4.3	0.004
3671	1985Q0565	WELL	222	6/12/1985	7.68	1,850	1,850	49	31	386	5.7	2.3	0.14	20.5	1096	170	10	0.2	0.8	1,216	10.6	0.005
3672	1983Q0696	WELL	213	7/20/1983	8.03	1,858	1,858	41	25	401	5.0	2.7	0.12	17.9	1004	248	20	0.22	1.2	1,257	12.1	0.011
3677	1985Q0374	WELL	350	5/3/1985	7.76	2,003	2,003	137	69	281	7.5	0.8	0.10	30.8	847	522	5	0.06	0.6	1,472	4.9	0.003
3677	206980	WELL	350	6/13/2014	9.3	7.63	2,560	204	96	321	8.2	0.7	0.20	26.8	920	666	7	<0.050	0.67	1,784	4.6	0.003
3678	1984Q0790	WELL	125	8/12/1984	9.0	7.7	939	116	42	40	5.5	3.4	0.21	30.6	556	87	1	1.95	0.3	602	0.8	0.001
3679	1983Q0668	WELL	220	7/18/1983	8	8	845	87	40	62	4.3	2.6	0.29	21.4	615	2	2	0.03	0.6	525	1.4	0.002
3680	1981Q0366	WELL	230	5/13/1981	7.66	1,581	1,581	65	32	230	6.1	3.5	0.06	29.1	701	257	4	0.1	0.65	974	5.8	0.002
3762	1987Q0487	WELL	200	6/30/1987	8.01	1,809	1,809	90	44	284	4.5	2.0	0.06	27.3	982	160	33	0.08	0.5	1,129	6.1	0.018
3762	206983	WELL	200	6/12/2014	8.4	8.04	1,774	92	43	293	4.7	0.1	0.05	23.3	989	173	46	<0.010	0.56	1,163	6.3	0.026
3764	1987Q0488	WELL	200	7/1/1987	7.89	2,153	2,153	103	59	321	5.4	2.3	0.11	25.9	1014	348	27	0.06	0.4	1,390	6.3	0.012
3764	206983	WELL	200	6/12/2014	8.8	7.83	2,012	111	67	304	5.9	1.4	0.16	23.9	1026	360	34	<0.010	0.59	1,413	5.6	0.017
3765	1979Q0389	WELL	240	8/21/1978	8.13	1,796	1,796	13	13	436	3.6	0.4	0.82	10.4	1124	0	76	<1	2.7	1,109	20.6	0.042
3766	1984Q1143	WELL	233	10/14/1984	8.06	1,729	1,729	49	30	355	4.1	0.1	0.04	19.0	1192	9	23	0.02	1.2	1,077	9.9	0.013
3766	2001Q0014	WELL	233	6/28/2000	9.0	7.87	1,780	47	31	324	4.3	3.2	0.12	18.4	1187	<25	26	<5	0.92	1,038	9.0	0.015
3766	206953	WELL	233	6/12/2014	9.1	8.37	1,687	46	30	362	4.4	0.1	0.09	17.2	1200	10	30	<0.010	0.86	1,114	10.2	0.018
3766	210030	WELL	233	8/31/2015	9.8	7.64	1,754	48	30	356	4.5	3.7	0.11	19.1	1230	10	32	<0.010	0.99	1,112	9.9	0.018
3769	1984Q0792	WELL	270	8/10/1984	7.81	1,762	1,762	32	17	397	3.7	0.8	0.04	15.9	1088	89	30	0.05	0.3	1,122	14.2	0.017
3769	2001Q0015	WELL	270	6/27/2000	8.51	2,250	2,250	19	11	454	3.5	0.8	0.02	10.9	1247	43	55	<5	1.45	1,213	20.5	0.024
3769	206945	WELL	270	6/12/2014	9.7	8.14	1,856	59	32	378	4.8	<0.04	0.06	18.3	1027	237	24	<0.010	0.85	1,268	9.8	0.013
3776	1979Q0392	WELL	295	8/26/1978	10.0	7.83	2,114	63	40	442	4.2	2.8	0.05	28.3	1293	172	18	1.4	0.8	1,409	10.7	0.008
3778	1985Q0731	WELL	222	6/21/1985	9.5	8.15	1,735	64	39	355	3.1	0.1	0.08	29.4	1315	8	7	0.15	0.7	1,155	8.6	0.004
3779	1985Q1038	WELL	270	9/12/1985	7.51	1,330	1,330	82	53	445	4.4	7.1	0.10	28.3	1269	299	42	0.65	0.3	1,587	9.4	0.031
3779	2001Q0016	WELL	270	6/28/2000	7.99	2,510	2,510	80	5	403	5.1	7.3	0.11	26.7	855	478	8	<5	0.76	1,435	11.8	0.003
3779	206906	WELL	270	6/11/2014	8.8	8.27	2,346	82	57	442	5.2	0.2	0.09	24.6	1116	524	8	0.06	0.58	1,712	9.2	0.003
3867	1984Q1142	WELL	336	10/13/1984	8.05	2,999	2,999	62	40	678	6.6	0.1	0.51	23.0	1682	359	31	0.03	0.2	2,030	16.4	0.010
43116	1996Q0266	WELL	253	8/14/1995	7.79	1,330	1,330	132	55	110	6.2	4.2	0.16	28.6	664	250	4	<0.05	0.5	918	2.0	0.003
44498	209590	WELL	271	6/29/2015	11.3	7.84	1,427	109	44	153	5.7	3.7	0.22	27.4	725	208	4	0.04	0.55	913	3.1	0.003
121117	1996Q0259	WELL	155	8/14/1995	9.1	8.21	1,102	121	44	84	5.2	4.6	0.25	29.7	587	150	4	<0.05	0.3	732	1.7	0.003
135910	1996Q0263	WELL	220	8/14/1995	11.3	7.71	1,744	85	49	312	5.2	3.6	0.06	26.0	1031	213	33	<0.05	<.1	1,234	6.7	0.019
135910	207271	WELL	220	6/23/2014	10.8	7.4	1,823	92	51	307	5.1	0.5	0.05	25.9	1076	206	42	0.05	0.61	1,260	6.4	0.023

GWIC ID	Sample Number	Site Type	Depth (ft)	Sample Date	Water Temp °C	Lab pH	Lab SC (µS/cm)	Ca	Mg	Na	K	Fe	Mn	SiO <sub>2</sub>	HCO <sub>3</sub>	SO <sub>4</sub>	Cl	NO <sub>3</sub> -N	F	TDS	SAR unit-less	Contam. Index unit-less
150615	1996Q0270	WELL	260	8/15/1995	10.9	8.16	950	90	39	79	4.9	4.0	0.09	25.7	572	100	3	<0.05	0.5	629	1.8	0.003
150940	206965	WELL	150	6/12/2014	9.1	7.92	1,131	111	44	93	5.0	2.4	0.15	28.2	610	204	3	<0.010	0.45	790	1.9	0.002
155929	206966	WELL	180	6/12/2014	11.5	7.86	1,111	112	47	72	6.0	<0.038	0.17	27.8	589	175	3	<0.010	0.37	733	1.4	0.003
155930	209589	WELL	180	6/29/2015	9.4	7.63	1,292	121	52	96	5.7	5.8	0.16	29.6	648	216	4	<0.010	0.47	850	1.8	0.003
169130	206967	WELL	160	6/12/2014	9.5	7.91	1,076	105	41	77	5.0	0.4	0.18	27.4	571	178	3	<0.010	0.47	718	1.6	0.002
169131	206964	WELL	160	6/12/2014	8.9	8.04	1,297	82	39	174	5.3	0.2	0.05	25.7	682	225	3	<0.010	0.59	890	4.0	0.002
168132	1996Q5000	WELL	240	2/11/1996	8.57	8.07	1,434	79	20	223	8.0	0.0	0.10	14.9	598	323	15	0.5	<0.5	970	5.8	0.010
168132	2004Q0323	WELL	240	12/12/2003	8.3	8.03	1,853	28	17	383	3.7	1.3	0.10	14.9	987	234	7	<0.5	<0.5	1,175	14.1	0.004
168866	1998Q5000	WELL	240	4/13/1998	8.46	8.46	867	60	33	74	5.0	0.1	0.19	23.4	513	0	4	0	0	431	1.9	0.005
168867	2000Q0467	WELL	240	11/5/1999	7.55	7.55	920	62	33	93	4.7	1.6	0.19	23.4	532	80	1	<5	0.46	560	2.4	0.001
168867	206969	WELL	240	6/13/2014	10.0	8.05	914	62	33	101	4.0	0.1	0.16	21.0	581	64	1	<0.010	0.61	573	2.6	0.001
169212	2000Q0468	WELL	108	11/4/1999	9.4	7.17	1,171	82	40	150	5.7	0.2	0.55	19.8	668	195	5	<5	0.75	829	3.4	0.005
169212	206901	WELL	108	6/10/2014	8.9	7.6	1,591	125	61	177	6.6	0.3	0.77	19.9	784	342	11	0.05	0.54	1,131	3.2	0.007
169213	2000Q0473	WELL	130	11/4/1999	9.2	7.12	1,240	108	53	92	8.9	0.1	0.67	18.6	605	213	11	<5	0.68	805	1.8	0.009
169213	206902	WELL	130	6/10/2014	10.0	7.45	1,211	112	57	87	6.5	<0.038	0.28	21.7	537	285	11	0.16	0.51	846	1.7	0.009
169214	2000Q0461	WELL	160	11/4/1999	9.5	7.37	1,923	108	21	267	11.3	<0.05	0.61	18.0	732	406	6	<5	0.64	1,200	6.1	0.003
169214	206903	WELL	160	6/10/2014	9.7	7.54	1,678	153	50	176	15.9	6.1	2.83	17.3	828	341	4	0.22	0.4	1,174	3.2	0.002
169215	2000Q0469	WELL	190	11/4/1999	10.0	7.54	1,346	92	47	159	7.8	1.4	0.09	26.7	689	253	4	<5	0.69	932	3.4	0.003
169217	2000Q0462	WELL	227	11/2/1999	9.4	7.14	1,835	199	79	105	6.7	2.8	0.50	27.6	720	539	5	<5	0.54	1,321	1.6	0.003
169217	206977	WELL	227	6/13/2014	9.6	7.79	1,581	164	62	122	5.4	2.1	0.34	26.5	677	379	4	<0.010	0.41	1,099	2.1	0.003
169218	2000Q0460	WELL	224	11/4/1999	10.1	7.45	1,228	113	47	95	6.0	4.1	0.15	30.0	611	227	3	<5	0.64	827	1.9	0.003
169218	206968	WELL	224	6/12/2014	9.4	7.85	1,292	126	53	98	5.4	1.6	0.14	28.5	621	274	3	<0.010	0.48	895	1.9	0.002
175651	206986	WELL	246	5/31/2014	12.1	7.72	1,707	168	71	156	6.7	1.5	0.18	27.3	786	443	5	0.05	0.5	1,267	2.5	0.003
186196	1998Q5005	WELL	190	4/27/1998	7.98	7.98	984	12	33	138	4.0	0.2	0.18	27.3	580	0	4	0	0	477	4.7	0.004
206410	1998Q5011	WELL	180	4/27/1998	8.08	8.08	1,899	28	60	273	7.0	0.1	0.25	33.2	903	279	7	0	0	1,099	6.7	0.004
212128	209385	WELL	240	6/1/2015	9.9	7.45	1,553	147	60	126	6.2	3.1	0.25	33.2	646	318	4	<0.010	0.47	1,015	2.2	0.002
215189	2005Q5004	WELL	175	3/22/2005	8.31	2,688	48	18	532	6.0	0.9	0.9	0.89	22.8	915	634	10	0	0	1,700	16.6	0.004
221606	2006Q0687	WELL	88	12/28/2005	7.6	7.37	1,440	155	78	103	7.0	0.2	0.09	22.8	635	385	3	1.21	<0.5	1,065	1.7	0.002
221614	2006Q0653	WELL	288	12/29/2005	9.3	7.78	2,450	86	53	480	5.9	5.6	0.09	23.8	985	583	11	0.698	0.8	1,735	10.1	0.004
221614	206985	WELL	288	6/11/2014	9.9	8.09	2,537	82	50	503	5.2	<0.075	0.07	21.6	1152	575	9	<0.010	0.69	1,814	10.8	0.003
221634	2006Q0688	WELL	278	1/17/2006	7.4	8.11	2,120	74	38	443	6.8	0.0	0.10	26.1	846	526	9	<0.5	1.14	1,540	10.5	0.004
221634	206978	WELL	278	6/13/2014	11.0	8.08	2,325	75	45	456	6.1	<0.075	0.05	25.6	868	637	8	<0.010	1.1	1,682	10.3	0.004
221637	2006Q0691	WELL	288	1/13/2006	7.7	8.77	1,939	8	3	531	3.0	0.1	0.02	26.2	1193	29	19	<0.5	<0.05	1,292	40.0	0.010
221637	206979	WELL	288	6/13/2014	10.2	8.73	2,120	3	2	582	2.3	0.1	<0.010	24.6	1376	33	19	0.05	3.36	1,401	62.0	0.009
221647	2006Q0689	WELL	195	1/9/2006	9.0	7.38	1,523	149	83	122	7.1	6.6	0.71	28.2	718	326	5	<0.5	0.59	1,083	2.0	0.004
221647	206975	WELL	195	6/13/2014	9.3	7.96	1,633	148	82	115	6.5	1.1	0.32	24.6	775	372	5	<0.010	0.44	1,137	1.9	0.003
221651	2006Q0690	WELL	248	12/30/2005	8.5	8.02	2,270	38	25	558	5.3	1.5	0.06	20.4	1546	51	50	<0.5	<0.05	1,510	17.2	0.022

GWIC ID	Sample Number	Site Type	Depth (ft)	Sample Date	Water Temp °C	Lab pH	Lab SC (µS/cm)	Ca	Mg	Na	K	Fe	Mn	SiO <sub>2</sub>	HCO <sub>3</sub>	SO <sub>4</sub>	Cl	NO <sub>3</sub> -N	F	TDS	SAR unit-less	Contam. Index unit-less
221651	206946	WELL	248	6/12/2014	8.8	8.38	2,300	33	23	560	5.2	0.3	0.05	18.9	1594	38	58	0.04	1.73	1,557	18.2	0.025
221651	222085	WELL	248	9/19/2016	10.5	7.67	2,475	34	24	559	5.0	2.0	0.04	21.4	1683	46	60	<0.010	1.71	1,582	18.0	0.024
221652	206971	WELL	180	6/13/2014	9.5	7.77	1,005	141	42	24	3.9	0.8	1.03	24.8	514	140	3	<0.010	0.12	634	0.5	0.003
252906	1028E+10	WELL	130	9/30/2014			1,000	122	57	27	4.9	0.5	0.50			117	2		0.18	332	0.5	0.002
267045	2012Q0503	WELL	100	10/10/2011		8.1	1,988	92	19	414	5.0	0.4				107	7			1,877	10.3	0.004
280602	208671	WELL	203	11/24/2014	8.0	7.8	2,106	193	129	165	11.0	0.8	0.07	24.3	924	664	7	<0.010	0.4	1,649	2.3	0.003
280618	208674	WELL	140	11/24/2014	8.5	7.75	1,056	104	41	97	5.0	3.6	0.18	27.1	610	147	2	<0.010	0.49	727	2.0	0.002
280618	222090	WELL	140	9/19/2016	10.4	7.08	1,087	103	41	89	5.5	4.7	0.16	30.8	618	148	3	<0.010	0.48	730	1.9	0.002
280621	210405	WELL	59	11/20/2015	8.5	7.27	2,029	165	84	173	9.0	0.1	1.09	27.5	779	510	20	4.82	0.75	1,378	2.7	0.010
280621	222091	WELL	59	9/19/2016	11.0	7.25	1,949	171	90	176	10.0	<0.038	0.98	28.7	776	530	21	2.6	0.49	1,414	2.7	0.011
<b>Water quality of the glacial outwash aquifer</b>																						
3606	1984Q1145	WELL	147	10/15/1984	7.7	7.79	2,488	95	45	464	6.2	0.2	0.75	18.7	871	680	7	0.05	0.2	1,746	9.8	0.003
3606	206982	WELL	147	6/13/2014	9.1	8.4	2,632	35	19	597	4.3	<0.075	0.09	13.8	972	572	7	<0.050	0.73	1,750	20.2	0.003
3611	1982Q0428	WELL	47	6/10/1982		7.32	2,069	185	133	144	7.5	1.2	0.10	13.9	573	863	4	2.5	0.52	1,628	2.0	0.002
3614	1981Q0365	WELL	48	5/13/1981		8.13	1,292	105	58	69	4.5	0.0	0.02	12.6	422	184	18	33.9	0.2	694	1.3	0.014
3673	1983Q0844	WELL	50	8/8/1983		7.85	1,301	48	28	236	3.4	0.3	0.06	12.4	700	158	5	0.02	0.5	836	6.7	0.004
3676	1982Q0950	WELL	40	8/27/1982		7.83	1,249	184	61	26	7.4	0.0	1.04	26.9	534	254	20	1.57	0.3	845	0.4	0.016
3682	1982Q0951	WELL	90	8/30/1982		7.72	1,256	166	73	26	5.2	0.0	0.04	21.2	503	241	20	9.44	0.2	809	0.4	0.016
3683	1981Q0369	WELL	140	5/13/1981		7.67	1,829	65	53	253	7.9	1.3	0.04	9.5	711	343	4	0.03	0.24	1,088	5.6	0.002
3761	1982Q0430	WELL	40	6/11/1982	6.5	7.43	2,842	150	92	417	9.9	0.3	1.82	19.2	788	865	135	0.65	0.58	2,080	6.6	0.048
3763	1987Q0486	WELL	200	7/11/1987		7.78	1,973	84	47	324	5.1	2.3	0.06	27.5	991	239	29	0.06	0.6	1,247	7.0	0.015
3763	206962	WELL	200	6/12/2014	8.6	7.91	1,871	89	53	313	5.5	0.1	0.06	24.7	1,018	253	43	<0.010	0.57	1,285	6.5	0.023
3767	1984Q1144	WELL	118	10/14/1984	8.1	7.92	2,003	142	77	263	6.3	0.2	0.09	26.5	1,080	291	40	0.03	0.2	1,378	4.4	0.020
3767	2001Q0017	WELL	118	6/28/2000	9.0	8.26	2,220	154	77	241	6.2	6.6	0.12	27.4	1,120	310	37	<5	<5	1,411	4.0	0.017
3767	206957	WELL	118	6/12/2014	8.5	7.76	2,087	150	80	268	6.5	0.3	0.11	25.8	1,067	353	41	<0.010	0.26	1,451	4.4	0.020
3767	209930	WELL	118	8/3/2015	21.1	7.16	2,166	148	82	289	6.4	3.7	3.36	30.1	1,141	351	42	<0.010	0.33	1,516	4.7	0.019
3770	1984Q0793	WELL	140	8/11/1984	8.5	7.8	1,600	108	51	213	5.0	5.5	0.20	26.0	779	304	3	0.03	0.1	1,099	4.2	0.002
3770	2001Q0020	WELL	140	8/26/2000		7.64	1,679	107	50	203	5.1	5.3	0.21	25.6	775	304	6	<5	0.69	1,089	4.1	0.003
3770	206900	WELL	140	6/9/2014	9.1	7.64	1,659	110	55	224	5.9	4.8	0.31	25.6	794	367	8	<0.010	0.46	1,192	4.3	0.005
3771	1988Q1401	WELL	100	8/21/1988		8.3	1,058	94	51	85	4.1	3.2	0.24	29.1	682	73	2	0.21	0.4	676	1.8	0.002
3771	1996Q0268	WELL	100	8/15/1995	9.2	7.93	1,224	101	49	136	4.9	2.5	0.28	25.4	662	200	4	<0.05	0.1	849	2.8	0.003
3773	1984Q0795	WELL	143	8/10/1984		7.34	1,307	128	62	99	4.8	3.1	0.25	25.4	646	265	2	0.04	0.1	907	1.8	0.002
3773	2001Q0019	WELL	143	6/27/2000	12.0	7.52	1,380	126	60	93	5.9	0.1	0.23	23.7	614	265	3	0.641	<5	878	1.7	0.002
3773	222103	WELL	143	9/21/2016	8.4	7.17	1,325	126	62	92	4.9	3.0	0.23	25.2	612	274	3	<0.010	0.19	892	1.7	0.002
3774	1986Q0957	WELL	210	6/10/1986		7.59	1,522	171	73	61	5.1	2.8	0.41	30.6	422	497	7	0.09	0.3	1,056	1.0	0.005
3774	206905	WELL	210	6/11/2014	9.2	7.67	2,162	293	123	91	6.9	4.7	0.85	25.9	438	1,147	19	0.05	0.19	1,928	1.1	0.009
3777	1986Q0958	WELL	80	8/23/1986		7.92	1,501	118	57	176	5.1	0.4	0.11	30.0	768	267	11	0.08	0.4	1,043	3.3	0.007

GWIC ID	Sample Number	Site Type	Depth (ft)	Sample Date	Water Temp °C	Lab pH	Lab SC (µS/cm)	Ca	Mg	Na	K	Fe	Mn	SiO <sub>2</sub>	HCO <sub>3</sub>	SO <sub>4</sub>	Cl	NO <sub>3</sub> -N	F	TDS	SAR unit-less	Contam. Index unit less
3777	222100	WELL	80	9/20/2016	8.9	7.4	1,927	145	69	215	6.1	5.4	0.13	27.8	805	483	16	0.05	0.25	1,364	3.7	0.008
3848	1982Q0949	WELL	30	8/25/1982		8.23	509	75	26	2	1.1	<.002	0.00	12.5	302	9	3	7.34	0.1	285	0.1	0.005
3852	1982Q0948	WELL	90	8/24/1982		7.33	2,205	177	139	179	8.3	11.8	0.23	18.0	833	726	7	0.062	0.4	1,679	2.4	0.003
3853	1985Q0564	WELL	101	6/9/1985	10.6	7.82	1,558	183	89	59	5.8	3.3	0.85	26.2	662	405	3	0.17	0.1	1,101	0.9	0.002
3854	1985Q1035	WELL	140	7/30/1985		7.55	1,622	173	83	67	5.8	0.1	0.93	28.5	681	346	3	0.27	0.2	1,044	1.0	0.002
3855	1982Q0946	WELL	128	8/17/1982		7.73	1,031	125	59	35	5.1	0.5	0.41	20.0	550	84	3	0.11	0.2	603	0.6	0.003
3856	1982Q0953	WELL	30	8/13/1982		7.92	1,191	64	62	124	11.8	0.4	0.53	28.2	534	201	23	0.2	0.5	779	2.6	0.020
3857	1982Q0952	WELL	40	8/19/1982		8.19	679	72	28	44	4.2	2.7	0.30	25.5	349	88	2	0.11	0.1	439	1.1	0.003
3858	1984Q0799	WELL	130	8/9/1984		7.77	660	68	26	42	3.8	2.5	0.31	26.7	347	89	2	0.09	0.1	431	1.1	0.003
3858	210025	WELL	130	6/27/2000	10.1	7.84	808	80	30	45	4.2	2.8	0.34	25.5	337	95	38	<.5	<.5	487	1.1	0.047
3858	1982Q0432	WELL	130	8/30/2015	8.4	7.36	892	89	35	51	4.8	3.4	0.35	26.5	352	121	82	<.010	0.21	586	1.2	0.092
3859	1996Q0264	WELL	120	6/9/1982		7.35	864	74	36	69	5.1	2.1	0.29	25.6	401	131	32	0.08	0.26	573	1.6	0.037
3859	209588	WELL	120	8/17/1995	9.7	7.69	1,092	93	40	72	5.3	3.2	0.37	25.6	380	100	110	<.05	<.1	636	1.6	0.101
3861	1985Q1039	WELL	120	6/15/2015	8.2	7.71	1,079	87	43	84	5.9	3.1	0.35	24.9	402	135	92	<.010	0.18	669	1.9	0.086
3862	1984Q0800	WELL	137	9/4/1985		7.71	1,227	63	33	200	4.4	2.9	0.13	27.1	682	168	9	0.06	0.5	847	5.0	0.007
3862	1984Q0800	WELL	150	8/9/1984		7.89	1,426	53	27	271	4.1	1.5	0.07	25.2	842	143	10	0.1	0.2	950	7.6	0.007
3866	1985Q1036	WELL	238	9/10/1985		7.44	2,023	65	31	426	6.3	2.3	0.05	29.2	1,281	174	1	0.08	0.1	1,366	10.9	0.001
3868	1985Q1037	WELL	150	9/9/1985		7.51	1,780	84	35	322	5.0	3.8	0.10	29.5	971	120	110	0.07	0.3	1,189	7.4	0.062
3869	1984Q0796	WELL	130	8/8/1984		7.52	1,161	110	50	51	4.9	5.2	0.43	26.8	317	86	181	0.07	0.2	672	1.0	0.156
3869	210028	WELL	130	6/12/2000	9.8	7.57	880	71	32	58	4.2	3.3	0.27	26.0	336	86	56	<.5	<.5	501	1.4	0.064
3870	1982Q0434	WELL	100	8/31/2015	9.6	7.25	831	76	36	53	4.4	3.5	0.27	27.5	339	107	76	<.010	0.25	550	1.3	0.091
3870	1996Q0271	WELL	100	6/13/1982	8.5	7.35	1,077	104	53	68	4.5	3.4	0.34	25.8	540	191	4	0.49	0.24	720	1.4	0.004
3871	1984Q0797	WELL	100	8/17/1995	9.1	7.84	650	67	28	36	3.5	2.6	0.33	26.4	330	100	7	<.05	0.1	433	0.9	0.010
3871	2001Q0021	WELL	98	8/8/1984		7.72	634	62	27	39	3.4	2.6	0.31	26.9	336	80	2	0.13	0.1	408	1.0	0.003
3871	210026	WELL	98	6/28/2000	9.0	7.77	699	65	28	35	3.4	2.7	0.31	25.5	323	76	10	<.5	<.5	405	0.9	0.015
3873	1984Q0540	WELL	98	8/30/2015	8.4	7.44	711	73	34	36	3.8	3.2	0.34	26.8	331	107	33	<.010	0.23	480	0.9	0.046
3873	1996Q0262	WELL	150	6/12/1984	8.0	7.73	1,055	104	54	67	4.5	0.1	0.34	26.6	538	186	5	0.29	0.2	713	1.3	0.004
3873	204784	WELL	150	8/17/1995	11.1	7.49	1,081	107	56	68	4.8	3.2	0.35	25.0	562	200	4	<.05	<.1	745	1.3	0.004
3873	1984Q0801	WELL	150	8/22/2013	10.6	7.26	1,168	111	60	67	5.4	1.7	0.35	24.2	560	224	4	<.010	0.26	772	1.3	0.004
3944	1984Q0801	WELL	130	8/8/1984		7.61	1,542	107	48	209	5.6	5.1	0.17	26.8	677	367	5	0.12	0.2	1,107	4.2	0.003
3944	2008Q05024	WELL	130	5/7/2008		6.98	7,102	56	55	251	5.8	0.0			598	521	77	1.2		1,261	5.7	0.011
3945	1982Q0436	WELL	153	6/13/1982		7.2	971	90	39	79	3.4	3.8	0.29	27.7	442	199	5	0.02	0.24	664	1.8	0.005
3945	207615	WELL	153	7/15/2014	9.5	7.49	1,094	102	45	90	3.5	3.2	0.36	26.2	455	290	9	<.010	0.22	793	1.9	0.008
3946	1982Q0435	WELL	154	6/13/1982		7.2	1,211	121	48	96	4.5	0.9	0.54	23.7	503	294	5	0.26	0.21	843	1.9	0.004
3946	207760	WELL	154	7/30/2014	8.4	7.41	1,220	99	44	87	4.0	0.6	0.50	23.4	328	286	7	<.010	0.17	712	1.8	0.005
3947	1984Q0802	WELL	110	8/11/1984		7.61	1,279	104	48	130	5.4	4.8	0.47	28.2	576	273	3	0.13	0.2	879	2.7	0.002
3947	2001Q0013	WELL	110	6/27/2000	10.0	7.43	1,466	124	55	132	5.6	5.8	0.62	26.0	565	348	5	<.5	0.37	981	2.5	0.004

GWIC ID	Sample Number	Site Type	Depth (ft)	Sample Date	Water Temp °C	Lab pH	Lab SC (µS/cm)	Ca	Mg	Na	K	Fe	Mn	SiO <sub>2</sub>	HCO <sub>3</sub>	SO <sub>4</sub>	Cl	NO <sub>3</sub> -N	F	TDS	SAR unit-less	Contam. Index unit less
3947	210023	WELL	110	8/30/2015	8.4	7.12	1,430	117	57	147	6.1	5.3	0.56	26.9	595	359	7	<0.010	0.32	1,018	2.8	0.005
3948	1982Q0437	WELL	104	6/13/1982	7.71	968	80	35	92	3.6	2.8	0.28	0.28	27.0	439	192	4	0.16	0.26	653	2.2	0.004
3948	204782	WELL	104	8/21/2013	9.3	7.38	1,063	94	43	87	4.4	2.1	0.39	24.1	397	283	14	<0.010	0.18	747	1.9	0.013
3949	1982Q0438	WELL	90	6/13/1982	7.62	830	69	31	72	3.2	2.4	0.25	0.25	26.8	398	145	3	0.21	0.26	547	1.8	0.004
3949	1986Q0269	WELL	90	8/17/1995	10.5	7.86	888	88	39	57	3.0	2.2	0.34	25.9	355	200	7	<0.05	0.1	597	1.3	0.007
4048	1983Q0666	WELL	45	7/14/1983	8.0	7.97	764	66	24	62	3.7	0.3	0.46	20.3	249	117	57	0.01	0.08	473	1.7	0.074
4051	1973Q0403	WELL	140	7/10/1973	7.39	2,930	181	142	390	4.4	0.0	2.30	19.9	956	1,022	26	0.54	0.1	2,259	5.3	0.009	
4052	1984Q0541	WELL	140	6/12/1984	8.5	7.53	1,501	135	55	154	5.7	2.7	0.20	28.4	601	387	8	0.04	0.2	1,072	2.8	0.005
4149	1973Q0402	WELL	140	7/11/1973	7.44	33,100	1,663	857	857	5,000	32.0	0.1	1.42	9.2	272	413	12,770	<0.023	<.1	20,879	24.8	0.386
44438	1982Q0431	WELL	82	6/11/1982	8.2	7.24	1,798	128	63	224	5.8	5.6	0.07	27.2	997	202	41	0.04	0.47	1,188	4.1	0.023
44466	222021	WELL	91	8/24/2016	8.4	7.3	1,939	145	66	265	6.0	4.8	0.06	25.7	899	444	53	0.08	0.25	1,452	4.6	0.027
44469	1984Q5003	WELL	131	9/18/1984	7.92	1,636	122	24	250	7.0	1.0	1.0	1.037	1,037	93	1	8	0	0.25	1,017	5.4	0.001
44473	222020	WELL	100	8/24/2016	9.6	7.31	2,308	1	0	605	0.9	0.1	<0.010	22.1	854	528	66	<0.010	0.25	1,642	196.4	0.028
45291	1990Q5000	WELL	80	3/8/1990	NA	7.89	1,079	168	56	91	8.0	0.0	NA	NA	622	346	8	0	NA	983	1.6	0.007
45292	1990Q5001	WELL	100	3/8/1990	NA	7.92	1,150	196	56	125	8.0	0.1	NA	NA	732	355	14	0	NA	1,115	2.0	0.012
45315	1982Q0433	WELL	152	6/9/1982	8.8	7.58	1,833	40	20	409	4.4	1.0	0.04	19.7	1,060	151	49	0.11	1.06	1,217	13.2	0.027
45315	1984Q0791	WELL	152	8/9/1984	8.5	7.78	1,742	38	19	394	4.0	1.1	0.04	19.8	1,005	146	29	1.63	0.2	1,148	13.1	0.017
45325	1989Q1374	WELL	120	10/13/1989	8.5	7.95	1,565	96	73	122	5.8	<.004	<.002	24.0	617	318	5	<.007	0.65	948	2.3	0.003
45325	1996Q0260	WELL	120	8/15/1995	10.0	7.59	1,374	125	64	107	5.2	2.1	0.35	23.2	619	300	6	<.05	0.2	937	1.9	0.004
45325	207618	WELL	120	7/16/2014	9.6	7.46	1,442	143	74	104	4.9	2.5	0.39	22.5	645	386	7	<.010	0.2	1,063	1.8	0.005
45327	1996Q0272	WELL	91	8/14/1995	9.9	7.83	848	83	35	64	4.8	1.8	0.33	26.2	436	100	7	0.1	0.2	537	1.5	0.008
45980	207612	WELL	126	7/9/2014	10.2	7.35	1,434	121	57	149	4.9	3.9	0.30	24.7	577	398	12	0.9	0.21	1,056	2.8	0.008
45981	204781	WELL	150	8/21/2013	9.6	7.08	1,498	122	54	144	5.3	4.7	0.19	25.7	469	418	40	0.07	0.18	1,045	2.7	0.027
45983	207611	WELL	95	7/9/2014	12.0	6.74	1,113	117	51	69	3.7	3.1	0.53	24.5	385	341	14	0.36	0.17	813	1.3	0.012
46003	204783	WELL	81	8/21/2013	19.5	7.13	1,408	126	57	115	6.8	2.8	0.68	22.8	467	393	23	0.07	0.14	977	2.1	0.016
46965	1983Q0665	WELL	40	7/14/1983	7.8	1,842	127	46	46	191	5.8	1.6	0.66	23.5	282	105	403	2.58	0.08	1,047	3.7	0.219
47739	1985Q0832	WELL	95	7/23/1985	8.9	7.52	1,375	115	48	139	6.1	1.6	0.41	26.1	418	356	40	5.87	0.3	944	2.7	0.029
125758	1990Q5002	WELL	80	6/24/1990	NA	7.78	1,384	108	29	260	7.0	0.0	NA	NA	842	286	20	0.6	NA	1,126	5.7	0.014
142105	207274	WELL	100	6/24/2014	9.0	7.02	1,490	175	82	69	5.8	3.2	0.70	24.0	703	382	4	0.31	0.21	1,092	1.1	0.002
149610	1982Q5050	WELL	92.5	6/8/1982	7.9	574	62	23	26	4.2	0.9	0.54	23.0	296	65	4	0.22	0.3	355	0.7	0.006	
150614	1996Q0265	WELL	68	8/17/1995	9.5	7.39	1,893	251	107	113	7.4	10.6	0.25	26.6	586	800	12	<.05	0.3	1,616	1.5	0.006
150614	207272	WELL	68	6/23/2014	9.1	7.24	1,411	139	59	123	6.0	5.7	0.15	26.1	679	320	6	<.010	0.46	1,019	2.2	0.004
150769	1996Q0273	WELL	100	8/15/1995	11.0	7.85	1,854	214	129	108	7.2	4.6	0.62	23.2	595	750	8	0.1	0.3	1,538	1.4	0.004
159786	1997Q5000	WELL	160	9/24/1996	8.11	1,899	75	60	302	1.0	0.4	0.4	0.21	732	474	12	0.1	0.57	772	2.4	0.003	
159786	2000Q0466	WELL	160	11/5/1999	9.5	7.72	1,190	91	46	112	4.8	3.9	0.21	26.4	662	157	4	<.5	0.47	447	1.4	0.002
159789	2000Q0470	WELL	180	11/5/1999	8.9	7.82	750	71	28	53	4.2	2.4	0.14	20.3	455	44	1	<.5	0.47	447	1.4	0.002
159791	1997Q5001	WELL	160	10/7/1996	8.5	902	123	1	101	5.0	0.2	0.2	0.14	488	119	6	0.1	0.47	601	2.5	0.007	

GWIC ID	Sample Number	Site Type	Depth (ft)	Sample Date	Water Temp °C	Lab pH	Lab SC (µS/cm)	Ca	Mg	Na	K	Fe	Mn	SiO <sub>2</sub>	HCO <sub>3</sub>	SO <sub>4</sub>	Cl	NO <sub>3</sub> -N	F	TDS	SAR unit-less	Contam. Index unit less	
159795	1986Q5015	WELL	150	5/20/1986		7.47	2,002	87	37	340	5.0	0.2			1,233	120	10	0		1,209	7.7	0.005	
159795	1986Q5014	WELL	150	5/28/1986		7.9	1,848	87	43	345	5.0	0.1			1,245	120	10	0		1,223	7.6	0.005	
159795	2010Q5098	WELL	150	12/2/2009		7.86	1,883	60	44	257	6.0	0.1			988	145	5	0		1,004	6.1	0.003	
159795	2010Q5099	WELL	150	12/22/2009		7.88	1,854	64	49	289	7.0	0.3			1,159	103	0	0		1,083	6.6	0.000	
159795	207273	WELL	150	6/24/2014	9.5	7.51	1,846	78	41	359	4.6	1.8	0.12		24.6	129	10	0.06	0.79	1,276	8.2	0.006	
160086	207763	WELL	132	7/29/2014	10.3	7.47	1,082	94	49	67	3.6	1.7	0.22		15.7	107	2	0.57	0.1	438	1.4	0.002	
161782	2006Q0295	WELL	10	9/16/2005	14.5	7.69	682	72	30	34	4.8				408	44	5			391	0.8	0.007	
161785	2006Q0309	WELL	8.8	9/15/2005	10.1	7.22	1,615	129	59	206	7.1				698	414	15			1,174	3.8	0.009	
161792	1997Q5003	WELL	118	NA		7.69	1,434	121	74	107	6.0	0.1			622	304	55	0.1		973	1.9	0.038	
161802	2003Q5036	WELL	90	4/28/2003		7.08	2,140	71	48	333	5.4	0.2			568	659	16	0.1		1,426	7.5	0.007	
161802	2004Q0324	WELL	90	12/12/2003	7.7	7.61	1,667	73	37	257	5.3	3.4	0.58		634	404	13	<0.5	<0.5	1,130	6.1	0.008	
161802	1997Q5004	WELL	90	NA		8.01	1,766	36	29	376	8.0	0.0			598	446	26	0.3		1,216	11.3	0.015	
161804	1997Q5006	WELL	180	NA		8.13	2,434	46	50	481	110	0.6			610	741	28	0.2		1,658	11.7	0.012	
168129	1998Q5006	WELL	130	NA		8.16	2,060	44	42	416	9.0	0.3			525	592	38	0.2		1,400	10.8	0.018	
168129	2008Q5025	WELL	140	5/4/1998		8.15	1,065	69	29	123	6.0	0.2			476	223	7	0		691	3.1	0.007	
169219	2000Q0465	WELL	140	5/2/2008		7.2	9,752	102	32	383	4.4	0.2			445	383	26	0.2		899	2.9	0.003	
169220	2000Q0464	WELL	80	11/4/1999	8.3	7.55	950	104	38	48	4.2	6.5	0.80		642	24	2	<5	0.26	572	1.0	0.002	
169372	2000Q0472	WELL	135	11/4/1999	8.7	7.76	790	73	33	48	4.8	2.2	0.68		492	47	1	<5	0.43	475	1.2	0.001	
169373	2000Q0471	WELL	220	11/2/1999	7.2	7.23	1,338	114	49	143	5.7	4.5	0.17		572	368	13	<5	<5	1,005	2.8	0.010	
169374	2000Q0463	WELL	30	11/2/1999	7.4	7.78	745	51	31	58	2.7	<0.1	0.18		315	94	3	15.89	<5	413	1.6	0.003	
186197	1998Q5007	WELL	200	11/7/1999	6.9	7.44	1,804	125	59	244	7.4	6.4	0.14		747	467	11	<5	<5	1,311	4.5	0.006	
193352	204320	WELL	90	5/4/1998		8.33	1,605	36	36	288	7.0	0.4			586	446	16	0.8		1,121	8.1	0.010	
193370	207761	WELL	108	7/18/2013	10.5	7.41	1,170	98	42	108	4.2	2.4	0.21		478	273	8	0.2	0.19	797	2.3	0.007	
206533	2004Q0326	WELL	142	7/30/2014	9.1	7.36	1,325	102	49	109	4.8	4.7	0.22		413	287	6	<0.010	0.17	791	2.2	0.005	
206533	222082	WELL	80	12/9/2003	7.6	7.67	1,964	73	39	329	8.4	3.2	0.34		675	500	10	0.568	<0.5	1,320	7.7	0.005	
206537	2004Q0321	WELL	80	9/20/2016	8.6	7.34	2,178	81	45	380	9.2	4.7	0.25		790	530	13	<0.010	0.16	1,481	8.4	0.006	
206539	2004Q0320	WELL	100	12/9/2003	7.4	7.3	1,987	132	60	250	7.4	7.0	0.11		666	578	10	<0.5	<0.5	1,399	4.5	0.005	
206546	2004Q0322	WELL	180	12/12/2003	6.9	7.71	2,850	118	60	487	6.7	5.1	0.55		892	862	34	<0.5	<1.0	2,058	9.1	0.012	
206546	2005Q5003	WELL	120	3/22/2005		7.88	2,082	179	61	257	6.8	7.1	0.16		686	556	9	<0.5	<0.5	1,399	4.6	0.004	
206546	222093	WELL	120	9/20/2016	9.7	7.09	2,134	136	67	287	7.3	6.7	0.08		732	553	8	0		1,395	4.6	0.004	
206547	2004Q0325	WELL	200	12/9/2003	6.8	7.72	1,994	109	39	312	4.9	4.9	0.10		806	593	10	<0.010	0.19	1,535	5.0	0.005	
206547	206904	WELL	200	6/11/2014	8.8	7.97	2,013	104	40	376	5.4	0.0	0.09		1,108	211	43	<0.5	<0.5	1,294	6.5	0.021	
206548	2004Q0327	WELL	200	12/12/2003	7.8	7.81	2,160	84	49	369	7.3	5.0	0.15		1,026	243	50	<0.010	0.36	1,349	7.9	0.025	
206548	2006Q0297	WELL	180	9/11/2005	9.3	7.58	2,150	79	46	404	7.2	4.7	0.11		1,209	291	52	<0.5	<0.5	1,478	7.9	0.024	
206548	206899	WELL	180	6/9/2014	8.2	7.37	2,279	86	49	438	7.7	4.7	0.11		1,198	305	60	0.06	0.44	1,406	8.9	0.022	
215220	222094	WELL	140	9/20/2016	9.2	7.38	1,610	73	40	252	6.1	1.8	0.12		840	258	4	<0.010	0.16	1,563	9.3	0.026	
																					1,075	5.9	0.003

GWIC ID	Sample Number	Site Type	Depth (ft)	Sample Date	Water Temp °C	Lab pH	Lab SC (µS/cm)	Ca	Mg	Na	K	Fe	Mn	SiO <sub>2</sub>	HCO <sub>3</sub>	SO <sub>4</sub>	Cl	NO <sub>3</sub> -N	F	TDS	SAR unit-less	Contam. Index unit-less
215223	222095	WELL	140	9/20/2016	8.8	7.38	1,361	99	41	158	5.1	0.6	0.20	25.8	587	315	9	<-0.010	0.2	943	3.4	0.006
21597	2006Q0652	WELL	128	12/27/2005	8.4	7.67	1,157	77	42	169	5.9	2.5	0.31	23.2	613	222	15	<-0.5	<-0.50	859	3.9	0.013
221597	222096	WELL	128	9/20/2016	9.7	7.41	1,163	82	39	135	4.9	3.5	0.16	26.0	676	136	5	<-0.010	0.35	765	3.1	0.005
221602	2006Q0685	WELL	108	12/28/2005	7.7	7.53	694	68	27	51	4.2	1.6	0.27	24.1	349	99	6	1.27	0.15	454	1.3	0.009
221602	222097	WELL	108	9/20/2016	8.8	7.51	741	76	29	51	4.1	1.6	0.25	24.3	377	112	11	<-0.010	0.14	495	1.3	0.015
221630	2006Q0654	WELL	168	12/29/2005	8.5	7.62	1,063	88	45	101	5.1	5.5	0.19	25.0	659	94	4	<-0.5	<-0.05	691	2.2	0.004
221630	206984	WELL	168	6/11/2014	9.6	7.82	1,123	99	49	94	4.4	1.6	0.18	23.3	692	98	3	<-0.010	0.55	714	1.9	0.002
221646	2006Q0650	WELL	68	1/5/2006	8.6	7.52	811	91	30	44	3.9	0.9	0.95	26.2	515	73	4	<-0.5	0.22	528	1.0	0.005
221646	206970	WELL	68	6/13/2014	8.6	8.1	765	92	29	35	3.4	<-0.015	0.75	26.0	476	66	4	<-0.010	0.17	491	0.8	0.005
221649	2006Q0686	WELL	168	12/30/2005	8.1	7.59	2,170	115	73	374	6.8	4.6	0.27	26.5	1,053	461	15	<-0.5	0.5	1,597	6.7	0.007
221649	206947	WELL	168	6/12/2014	8.3	7.92	2,493	132	88	390	6.4	2.3	0.11	24.5	1,127	636	16	0.04	0.33	1,850	6.4	0.006
221649	222086	WELL	168	9/19/2016	11.1	7.22	2,649	142	92	385	6.7	5.6	0.12	27.6	1,261	603	16	<-0.050	0.53	1,902	6.2	0.006
222551	207759	WELL	95	7/30/2014	8.4	7.13	1,192	105	48	81	3.2	3.2	0.44	25.3	307	356	20	<-0.010	0.14	791	1.6	0.016
223319	2008Q0623	WELL	180	5/2/2008	7.65	7.65	5,724	44	44	406	5.8	0.2	0.10	25.9	610	732	51	1.2	0.32	1,608	10.3	0.009
223319	207614	WELL	180	7/15/2014	10.5	7.63	1,594	62	34	269	4.8	2.5	0.10	25.9	646	399	17	<-0.010	0.32	1,133	6.8	0.011
242569	207616	WELL	100	7/16/2014	10.5	7.3	1,878	222	114	76	5.9	5.0	0.47	18.7	516	768	30	0.6	0.25	1,495	1.0	0.016
248885	2009Q0601	WELL	150	12/16/2008	8.22	7.95	2,935	64	36	435	6.0	0.7	0.18	25.7	1,135	255	6	0	0	1,362	10.8	0.002
250391	207613	WELL	150	7/15/2014	9.7	7.43	1,255	108	49	126	5.1	3.5	0.18	25.7	521	343	6	<-0.010	0.12	922	2.5	0.005
250396	204318	WELL	200	7/16/2013	10.9	7.6	1,546	112	51	155	5.3	1.6	0.19	27.0	593	350	13	0.16	0.42	1,008	3.0	0.009
262871	209975	WELL	115	8/19/2015	7.3	7.31	2,579	89	45	485	6.8	2.1	3.53	26.3	950	664	14	<-0.010	0.25	1,804	10.5	0.005
262872	209976	WELL	50	8/19/2015	8.9	7.29	3,110	126	62	560	9.2	3.0	2.95	27.6	946	961	30	<-0.050	0.35	2,248	10.2	0.010
273937	207042	WELL	136	6/11/2014	7.8	8	1,268	106	58	102	5.0	2.1	0.35	24.6	627	252	3	<-0.010	0.23	862	2.0	0.003
273937	210092	WELL	136	9/13/2015	8.0	7.06	1,268	107	58	107	4.8	5.0	0.36	25.1	616	255	4	<-0.010	0.27	869	2.1	0.003
280641	208669	WELL	140	1/23/2014	7.8	7.77	1,725	86	50	296	5.6	0.3	0.08	22.6	958	285	12	<-0.010	0.47	1,230	6.3	0.007
280641	222087	WELL	140	9/19/2016	10.3	7.26	1,804	86	49	279	5.9	3.0	0.06	26.8	978	289	12	<-0.010	0.45	1,213	6.0	0.007
280643	208668	WELL	248	11/23/2014	8.5	8.15	1,986	38	24	464	4.6	0.2	0.05	20.0	1,269	115	41	<-0.010	1.22	1,343	14.5	0.021
280643	222088	WELL	248	9/19/2016	11.0	7.71	2,062	38	25	449	4.8	1.7	0.04	22.1	1,306	104	44	<-0.010	1.2	1,333	13.9	0.021
280645	208667	WELL	120	11/23/2014	7.1	7.83	2,511	149	97	366	5.3	1.3	0.42	14.3	697	975	11	<-0.010	0.34	1,962	5.7	0.005
280645	222098	WELL	120	9/20/2016	8.8	7.36	2,681	152	98	357	6.1	3.9	0.34	15.7	715	1,002	12	<-0.050	0.39	1,999	5.6	0.004
280650	210404	WELL	90	1/20/2015	8.1	7.27	2,489	171	108	240	5.8	<-0.038	1.32	21.9	756	747	22	6.69	0.66	1,698	3.5	0.009
280650	222099	WELL	90	9/20/2016	8.5	7.35	2,429	176	110	251	5.9	<-0.075	1.37	21.9	754	804	25	7.56	0.35	1,773	3.7	0.010
280651	208666	WELL	105	11/23/2014	7.5	7.54	1,718	133	64	219	6.3	0.6	0.13	22.0	871	337	29	<-0.010	0.34	1,240	3.9	0.017
280652	208672	WELL	140	11/23/2014	7.6	7.71	3,325	298	153	393	8.6	2.8	0.40	23.7	1,466	850	21	<-0.050	0.38	2,473	4.6	0.006
280652	222084	WELL	140	9/19/2016	10.7	7	3,287	286	139	378	8.3	13.6	0.32	28.3	1,526	904	22	<-0.050	0.4	2,530	4.6	0.007
283920	209968	WELL	64	8/18/2015	9.6	7.3	1,027	108	33	80	3.1	4.0	4.13	21.8	578	126	7	<-0.010	0.44	671	1.7	0.007
283920	210098	WELL	64	9/14/2015	9.3	7.37	1,014	108	32	80	2.7	4.6	1.63	22.8	580	116	7	<-0.010	0.43	661	1.7	0.006
703881	207617	WELL	112	7/16/2014	9.8	7.5	1,450	141	71	108	4.6	3.7	0.23	22.9	677	368	4	<-0.010	0.19	1,057	1.9	0.003

GWIC ID	Sample Number	Site Type	Depth (ft)	Sample Date	Water Temp °C	Lab pH	Lab SC (µS/cm)	Ca	Mg	Na	K	Fe	Mn	SiO <sub>2</sub>	HCO <sub>3</sub>	SO <sub>4</sub>	Cl	NO <sub>3</sub> -N	F	TDS	SAR unit-less	Contam. Index unit-less
703926	204319	WELL	90	7/16/2013	9.7	7.52	1,550	128	57	143	5.3	2.5	0.62	24.1	485	420	49	1.45	0.18	1,069	2.6	0.031
703929	207762	WELL	122	7/29/2014	8.2	7.59	986	78	32	63	4.1	3.6	0.25	18.5	98	128	6	38.09	0.1	420	1.5	0.006
703974	204317	WELL	32	7/16/2013	9.2	7.46	691	71	26	43	3.2	0.4	0.59	22.4	302	131	7	0.38	0.14	453	1.1	0.010
<b>Water quality of the Fort Union Formation aquifer.</b>																						
3610	1979Q0512	WELL	186	9/20/1978	8.53	2,090	3	2	550	2.1	0.1	0.1	0.01	7.1	1,245	127	1	0.4	5.3	1,330	64.2	0.000
3612	1981Q1189	WELL	195	7/31/1981	8.18	2,889	4	3	736	2.5	0.2	0.2	0.01	6.8	959	690	4	0.09	2	1,922	65.3	0.001
3613	1981Q0367	WELL	162	5/13/1981	7.53	2,530	139	97	326	10.1	2.1	2.1	0.25	13.5	864	798	7	<0.2	0.31	1,819	5.2	0.003
3615	1981Q0368	WELL	87	5/13/1981	7.64	3,335	333	191	152	8.4	0.2	0.2	0.25	10.7	619	1,604	40	3.8	0.26	2,648	1.6	0.012
3681	1979Q0379	WELL	83	8/24/1978	9.3	1,993	33	35	396	5.9	1.1	0.04	0.04	15.0	749	397	6	1.6	2.7	1,284	11.5	0.003
3772	1984Q0794	WELL	318	8/10/1984	8.11	2,186	3	1	579	1.7	0.1	0.1	0.02	10.0	1,410	3	87	0.04	0.1	1,379	73.2	0.040
3772	222102	WELL	318	9/21/2016	9.3	8.29	2,314	2	1	582	2.0	<0.075	<0.010	6.5	1,431	1	104	<0.010	2.99	1,406	84.4	0.045
3850	1982Q0954	WELL	110	8/15/1982	7.86	1,565	157	113	73	6.4	0.3	0.03	0.03	19.0	634	511	5	0.42	0.2	1,201	1.1	0.003
3872	1984Q0798	WELL	330	8/9/1984	8.21	1,940	3	1	502	2.2	0.4	0.01	0.01	11.7	1,174	75	40	0.38	0.1	1,213	67.9	0.021
3872	210027	WELL	330	8/30/2015	10.4	8.51	1,998	2	1	499	2.2	<0.038	<0.005	6.4	1,214	76	50	<0.010	4.18	1,266	72.9	0.025
3943	1983Q0845	WELL	183	8/8/1983	7.82	1,110	7	4	265	3.0	0.3	0.0	0.01	22.3	548	148	3	0.04	0.3	722	19.6	0.003
3950	1979Q0388	WELL	150	8/21/1978	8.5	7.65	1,030	80	38	109	4.1	0.4	0.65	22.1	497	167	5	0.2	0.3	671	2.5	0.005
4050	1976Q1648	WELL	70	3/7/1977	7.35	1,903	164	88	177	6.8	0.1	0.55	0.55	22.6	732	527	7	1.446	0.1	1,356	2.8	0.003
157676	209610	WELL	260	7/8/2015	10.2	8.04	3,839	14	10	875	5.1	0.3	<0.020	8.2	1,382	864	22	<0.050	1.15	2,480	43.5	0.006
206406	200284	WELL	105	6/28/2011	9.5	7.38	2,127	137	81	358	6.6	6.8	0.38	26.1	825	744	8	0.05	0.38	1,775	6.0	0.004
206406	209979	WELL	105	8/20/2015	9.8	7.54	2,557	145	77	344	6.1	3.6	3.53	25.3	853	770	9	0.05	0.53	1,805	5.8	0.004
255044	1.028E+10	WELL	444	9/30/2014		3,360	9	6	1,040	4.2	0.2	0.2	0.01		692	692	22		1.9	1,775	65.1	0.007
<b>Water quality of the Fox Hills-Hell Creek aquifer</b>																						
43095	1979Q0378	PETWELL	1,160	8/24/1978	8.66	3,532	4	1	800	2.6	0.1	<0.1	<0.1	12.8	942	3	709	<.1	0.6	2,036	88.4	0.201
43095	1982Q0429	PETWELL	1,160	6/11/1982	11.0	8.15	3,502	5	1	813	1.9	0.0	<0.001	12.2	1,040	<.1	686	0.43	2.9	2,034	92.8	0.196
43095	209967	PETWELL	1,160	8/18/2015	10.9	8.47	3,699	3	1	865	2.7	0.2	0.28	9.3	994	3	767	<0.050	2.72	2,171	109.7	0.207
154904	204148	WELL	840	6/18/2013	11.9	8.28	3,250	4	1	692	2.4	0.1	<0.010	10.2	1,075	<2,500	578	<0.050	2.16	1,836	83.1	0.178
154904	210102	WELL	840	9/14/2015	16.5	8.21	3,340	4	1	805	1.9	<0.075	<0.010	10.7	1,106	<2,500	594	<0.050	2.29	1,976	94.4	0.178
<b>Water quality of Mesozoic and Paleozoic reservoirs associated with hydrocarbon development</b>																						
895341	1985Q1242	PETWELL	4,949	2/11/1985	8.01	17,000	295	<0.1	5,910	109.0	0.5			1,281	945	8,210	18			16,128	94.8	0.483
895523	1968Q0004	PETWELL	4,590	3/20/1968	7.8	16,000	170	36	5,700	100.0				2,560	68	7,800				15,135	103.6	0.488
895504	1977Q0024	PETWELL	4,990	NA	7.5	36,000	504	72	13,010	127.0				903	3,500	18,300				35,958	143.6	0.508
895581	1986Q5012	PETWELL	10,808	7/7/1986	6.14	237,000	6,290	1,225	83,700	2,270.0	26.0			226	780	141,900	378			236,680	252.9	0.599
895604	1960Q0011	PETWELL	7,976	10/26/1960	6.6	328,000	12,000	1,600	110,000	4,200.0				110	290	200,000				328,144	250.4	0.610
3851	1960Q0061	PETWELL	7,558	NA	6.2	319,000	9,600	1,392	112,800					168	419	194,600				318,894	284.8	0.610
3775	1960Q0060	PETWELL	7,864	NA	5.2	331,000	13,800	2,760	110,800					12	649	202,950				330,965	225.2	0.613
<b>Water quality of surface water</b>																						
3860	1987Q0485	LAKE		6/29/1987	21.8	9.2	5,722	5	232	1,134	61.0	0.0	<0.002	2.5	1,050	2,101	59	0.08	0.2	4,344	15.9	0.010

GWIC ID	Sample Number	Site Type	Depth (ft)	Sample Date	Water Temp °C	Lab pH	Lab SC (µS/cm)	Ca	Mg	Na	K	Fe	Mn	SiO <sub>2</sub>	HCO <sub>3</sub>	SO <sub>4</sub>	Cl	NO <sub>3</sub> -N	F	TDS	SAR unit-less	Contam. Index unit less
3860	1995Q5014	LAKE		9/25/1994		9.21	5,950	5	284	1,360	75.6	0.0	0.01	2.5	<1	2,396	73	<0.01	0.34	5,217	17.2	0.012
3860	1995Q5015	LAKE		11/27/1994		9.17	6,010	6	296	1,290	76.9	0.0	0.01	2.2	<1	2,343	74	<0.01	0.38	4,588	16.0	0.012
3860	1995Q5016	LAKE		12/26/1994		9.2	5,010	6	287	1,300	74.9	0.0	0.01	3.0	<1	2,335	73	0.02	0.32	4,549	16.4	0.015
3860	1995Q5017	LAKE		1/29/1995		9.14	5,930	4	300	1,320	81.6	0.0	0.01	2.7	<1	2,379	73	0.06	0.36	4,561	16.3	0.012
3860	1995Q5018	LAKE		2/27/1995		9.14	5,610	7	258	1,130	89.5	0.1	0.04	3.1	<1	2,177	69	0.19	0.34	3,974	15.0	0.012
3860	1995Q5019	LAKE		4/1/1995		9.23	1,724	3	65	247	20.1	0.0	0.01	5.7	<1	441	16	0.13	0.13	974	6.5	0.009
3860	1995Q5020	LAKE		5/1/1995		9.17	5,620	4	273	1,040	86.8	0.0	0.01	1.6	<1	2,175	69	0.06	0.33	4,646	13.4	0.012
120843	1990Q0326	LAKE		8/29/1990	21.4	9.49	4,727	17	141	1,050	49.7	0.1	0.01	3.9	1,240	1,460	71	0.32	0.34	3,596	18.3	0.015
120844	1990Q0324	LAKE		8/25/1990	22.0	9.41	5,396	10	161	1,190	88.7	0.1	0.01	12.2	1,870	1,270	91	0.25	0.21	3,998	19.8	0.017
120845	1990Q0323	LAKE		8/23/1990	23.3	9.74	11,374	4	322	2,990	272.0	0.0	0.01	5.5	2,420	4,760	128	<0.7	0.03	10,297	35.6	0.011
120847	1990Q0331	LAKE		8/24/1990	19.0	9.29	24,490	3	411	7,360	360.0	0.0	0.00	<1	3,760	12,800	365	0.1	0.11	24,031	77.7	0.015
120848	1990Q0325	LAKE		8/25/1990	21.9	9.52	2,684	12	131	471	33.5	0.0	0.01	13.0	632	864	26	<0.7	0.36	1,985	8.6	0.010
120848	2006Q0322	LAKE		9/16/2005		9.4	1,650	13	83	262	17.5				553	222	55			1,039	5.9	0.033
120850	1990Q0327	LAKE		8/27/1990	19.9	9.79	27,374	9	274	8,840	224.0	0.1	0.06	3.3	2,740	14,800	1,070	<0.7	0.4	27,386	113.4	0.039
120852	1990Q0328	LAKE		8/26/1990	19.6	9.43	7,291	14	118	1,800	35.3	0.1	0.01	3.2	1,120	3,040	80	<0.7	0.13	5,814	34.4	0.011
120854	1990Q0330	LAKE		8/26/1990	20.8	9.38	57,987	7	512	32,700	1,050.0	0.0	0.00	<1	9,300	53,000	4,110	<1.0	0.04	99,630	308.6	0.071
120861	1990Q0346	LAKE		8/24/1990	25.3	8.87	6,525	19	517	1,030	96.6	0.0	<0.02	16.2	1,200	3,080	61	1.08	0.15	5,546	9.6	0.009
120862	1990Q0337	LAKE		8/29/1990	21.5	9.35	5,521	17	89	1,280	59.6	0.2	0.02	5.6	1,464	1,500	97	0.17	0.24	3,967	27.5	0.018
120862	206372	LAKE		9/9/2014	14.3	9	3,488	25	93	689	34.1	<0.075	<0.010	1.4	947	828	61	<0.050	0.3	2,334	14.2	0.018
120863	1990Q0336	LAKE		8/24/1990	26.5	9.78	32,748	2	111	11,240	554.0	0.0	0.01	1.3	7,420	12,800	725	<1.0	0.24	32,398	227.6	0.022
120864	1990Q0338	LAKE		8/24/1990	25.5	9.74	27,769	3	139	9,740	350.0	<0.004	<0.02	7.0	9,690	7,060	977	<1.0	1.1	27,080	176.0	0.035
120865	1990Q0335	LAKE		8/26/1990	16.7	9.05	15,949	8	706	3,780	246.0	0.1	0.01	2.4	1,810	8,640	265	<0.7	0.05	14,843	30.4	0.017
120866	1990Q0334	LAKE		8/24/1990	21.0	9.34	9,409	4	335	2,220	194.0	0.0	0.01	2.5	2,920	2,860	75	<0.7	0.08	7,777	25.9	0.008
120867	1990Q0340	LAKE		8/24/1990	24.7	9.79	174,957	3	194	4,110	320.0	<0.004	0.01	4.4	3,580	4,650	135	<0.7	0.15	12,599	63.0	0.001
120868	1990Q0342	LAKE		8/25/1990	25.2	9.07	3,573	12	191	610	89.5	0.1	0.01	6.8	1,200	947	28	1.19	0.15	2,615	9.3	0.008
120869	1990Q0344	LAKE		8/26/1990	18.4	9.11	12,579	10	272	3,100	181.0	0.1	0.02	3.4	2,650	4,790	256	1.47	0.2	10,307	39.9	0.020
120870	1990Q0333	LAKE		8/26/1990	20.0	9.6	17,952	25	302	7,170	130.0	0.2	0.01	6.9	979	11,000	3,070	0.1	0.08	22,558	86.4	0.171
120871	1990Q0343	LAKE		8/27/1990	23.0	9.95	60,143	4	31	56,000	520.0	0.1	0.01	15.0	15,900	84,600	2,870	<2.0	1.45	162,274	2079.6	0.048
120872	1990Q0345	LAKE		8/29/1990	22.5	9	61,732	376	1,600	47,000	704.0	0.0	0.02	0.3	309	93,300	10,500	<2.0	0.04	153,740	235.7	0.170
120873	1990Q0332	LAKE		8/29/1990	22.4	8.72	28,872	299	1,940	5,780	322.0	0.3	0.02	2.0	212	15,100	4,180	<1.0	0.03	27,775	26.9	0.145
120876	1990Q0347	LAKE		8/24/1990	25.0	10	2,282	6	151	430	23.1	0.0	<0.002	15.2	700	515	12	0.07	0.14	1,802	7.4	0.005
120885	1990Q0365	LAKE		8/30/1990	25.8	10	98,083	2	29	108,000	750.0	0.8	0.01	22.0	71,000	75,700	7,690	<2.0	10.5	282,003	4229.4	0.078
120886	1990Q0366	LAKE		9/3/1990	20.4	9.23	17,952	6	621	4,660	358.0	0.2	0.01	12.5	3,790	8,720	262	0.38	0.26	17,003	40.0	0.015
120887	1990Q0367	LAKE		8/31/1990	18.6	8.94	14,846	25	944	3,410	369.0	0.0	<0.002	33.7	1,601	9,310	258	0.82	0.09	15,379	23.6	0.017
214744	2005Q0045	WETLAND		7/21/2004	26.0	9.63	9,220	5	209	2,304	86.0				1,424	4,100	298			8,053	33.9	0.032
214747	2005Q0047	WETLAND		7/23/2004	21.0	8.01	3,960	136	280	343	43.2				398	457	770			2,225	3.9	0.194
214788	2005Q0185	WETLAND		9/14/2004	16.0	7.23	7,270	267	337	1,063	29.8	0.5	0.11	12.5	304	339	2,687	<0.50	<0.50	4,886	10.2	0.370

GWIC ID	Sample Number	Site Type	Depth (ft)	Sample Date	Water Temp °C	Lab pH	Lab SC (µS/cm)	Ca	Mg	Na	K	Fe	Mn	SiO <sub>2</sub>	HCO <sub>3</sub>	SO <sub>4</sub>	Cl	NO <sub>3</sub> -N	F	TDS	SAR unit-less	Contam. Index unit less
214789	2005Q0186	WETLAND		9/10/2004	16.0	9.27	6,900	9	73	2,094	58.3	0.1	<0.010	1.6	2,317	1,738	279	<2.50	<2.50	5,898	50.9	0.040
214790	2005Q0187	WETLAND		9/14/2004	13.0	8.05	7,510	321	358	1,025	25.7	0.5	0.22	26.1	697	257	2,639	<0.50	<0.50	4,995	9.4	0.351
214791	2005Q0188	WETLAND		9/14/2004	13.0	7.37	15,370	699	752	2,492	67.8	0.9	0.78	15.0	479	763	6,770	0.833	<0.50	11,798	15.6	0.440
214792	2005Q0189	WETLAND		9/14/2004	13.0	7.84	5,210	148	163	723	34.1	0.2	0.38	10.7	531	310	1,287	<0.50	<0.50	2,938	9.8	0.247
214907	2005Q0217	STREAM		9/29/2004	16.2	8.91	2,580	34	93	553	22.4	0.3	0.03	10.1	704	950	34	<0.50	<0.50	2,109	11.1	0.013
214929	2005Q0226	STREAM		9/29/2004	15.5	8.64	2,430	36	80	554	9.2	0.0	0.00	0.4	634	787	21	<0.50	0.36	1,849	11.8	0.009
214929	2009Q0511	STREAM		8/19/2008			3,685	23	87	693	14.0				805	1,100	33			2,457	14.8	0.009
221718	2006Q0298	POND		9/16/2005	20.0	9.17	22,800	28	898	8,307	176.0				1,150	16,887	1,338			28,500	58.9	0.059
262048	200282	LAKE		6/30/2011	22.0	8.7	1,973	25	68	412	22.7	<0.040	<0.020	7.5	514	687	31	<0.050	0.13	1,548	9.7	0.015
262048	206368	LAKE		5/14/2014	11.0	8.77	2,409	22	76	455	22.2	<0.038	<0.005	6.8	604	787	35	0.14	0.21	1,738	10.3	0.014
262243	207632	LAKE		7/18/2014	24.8	9.12	5,275	8	121	1,219	65.2	<0.150	<0.020	8.8	1,300	1,690	95	<0.050	0.25	4,059	23.3	0.018
262246	207635	LAKE		7/18/2014	23.5	9.12	1,598	23	94	245	12.2	0.1	0.02	8.5	472	457	17	<0.010	0.2	1,177	5.1	0.011
262247	207634	LAKE		7/18/2014	23.1	8.79	2,800	40	131	498	17.4	<0.075	<0.010	17.7	770	883	38	<0.010	0.24	2,078	8.6	0.014
262248	207644	LAKE		7/18/2014	25.0	9.13	2,759	28	122	485	16.1	<0.075	0.01	6.9	647	894	37	<0.010	0.24	2,019	8.8	0.014
262249	207633	LAKE		7/18/2014	23.9	8.7	2,400	37	96	433	22.3	<0.038	<0.005	8.8	693	689	34	<0.010	0.23	1,735	8.5	0.014
262250	208368	LAKE		9/8/2014	15.0	8.94	5,137	21	194	1,038	39.4	<0.150	0.04	2.1	1,046	2,008	80	<0.050	0.33	4,049	15.5	0.016
262251	208365	LAKE		9/8/2014	17.5	9.24	2,657	50	116	477	22.0	<0.038	0.01	2.1	461	1,002	41	<0.010	0.3	2,059	8.5	0.015
262252	208361	LAKE		9/8/2014	19.7	8.58	4,571	99	265	720	31.3	<0.150	0.42	6.2	568	2,239	89	<0.050	0.28	3,788	8.6	0.019
262254	208369	LAKE		9/8/2014	17.3	9.44	3,049	32	81	594	20.9	<0.075	<0.010	0.3	445	864	46	<0.050	0.27	2,053	12.7	0.015
262255	208367	LAKE		9/10/2014	10.1	8.69	2,260	34	77	414	20.3	<0.038	<0.005	8.8	604	753	31	0.1	0.26	1,679	9.0	0.014
262256	208362	STREAM		9/8/2014	19.2	8.59	1,632	56	58	269	9.4	<0.038	<0.005	8.6	495	528	13	<0.010	0.24	1,212	6.0	0.008
277423	206960	WETLAND		6/6/2014	14.2	8.03	18,185	337	1,600	4,488	181.3	<0.750	<0.100	1.7	292	16,108	808	<0.100	<0.100	23,666	22.7	0.044
277462	207020	WETLAND		6/6/2014	17.7	9.64	65,461	15	36	28,741	126.4	<0.750	<0.100	16.2	2,955	34,617	2,157	<1.000	4.08	69,939	918.2	0.033
277471	207422	WETLAND		7/8/2014	18.9	8.31	4,602	154	171	790	47.8	0.2	0.14	12.1	780	2,239	104	<0.050	0.29	3,920	10.4	0.023
277475	207330	WETLAND		6/17/2014		7.64	899	67	55	13	54.8	0.1	0.30	13.9	510	24	31	<0.010	0.27	511	0.3	0.034
277481	206956	WETLAND		6/5/2014	17.9	7.49	3,193	443	213	171	33.4	0.6	2.73	18.5	407	1,766	205	<0.010	0.35	3,054	1.7	0.064
277533	206370	WETLAND		5/22/2014	17.5	8.16	1,543	13	34	346	22.7	0.1	0.07	18.2	842	140	48	<0.010	0.08	1,037	11.5	0.031
277540	207421	WETLAND		7/10/2014	24.7	7.7	854	75	51	43	4.0	0.0	0.14	27.5	555	37	6	<0.010	0.26	516	0.9	0.006
277572	206954	WETLAND		6/11/2014	17.7	8.57	10,573	216	658	1,855	87.5	<0.750	<0.100	<1.000	307	4,820	1,705	<0.100	<0.100	9,536	14.2	0.161
277581	207022	WETLAND		6/10/2014	20.9	8.2	3,933	50	234	668	70.0	<0.150	0.08	6.3	869	1,821	67	<0.050	0.28	3,368	8.8	0.017
277595	206951	WETLAND		6/11/2014	16.5	7.57	4,362	117	165	565	19.6	<0.150	0.12	4.8	649	149	1,177	<0.050	0.32	2,518	7.9	0.270
277600	207420	WETLAND		7/10/2014	24.6	9.63	13,983	197	926	3,246	89.3	1.0	<0.100	<1.000	68	10,188	194	<0.100	<0.100	14,962	21.5	0.014
277607	206950	WETLAND		6/1/2014	17.5	8.15	674	73	36	12	35.1	0.0	0.01	6.3	448	10	12	0.05	0.24	407	0.3	0.018
277612	206369	WETLAND		5/28/2014	23.6	8.25	1,205	56	63	150	11.3	0.1	0.01	14.0	632	179	18	<0.010	0.2	809	3.3	0.015
277631	207349	WETLAND		6/19/2014	23.5	8.55	14,399	157	375	3,838	40.6	<0.750	0.25	12.3	963	8,272	582	<0.100	<0.100	13,812	37.9	0.040
278957	208359	POND		9/9/2014	14.2	8.57	327	21	12	28	24.9	0.0	0.02	2.7	176	7	8	<0.010	0.17	204	1.2	0.024
278958	208371	POND		9/9/2014	14.9	9.1	1,774	32	121	219	28.7	<0.038	0.01	0.2	316	682	29	<0.010	0.34	1,348	4.0	0.016

GWIC ID	Sample Number	Site Type	Depth (ft)	Sample Date	Water Temp °C	Lab pH	Lab SC (µS/cm)	Ca	Mg	Na	K	Fe	Mn	SiO <sub>2</sub>	HCO <sub>3</sub>	SO <sub>4</sub>	Cl	NO <sub>3</sub> -N	F	TDS	SAR unit-less	Contam. Index unit less
278960	208363	POND		9/9/2014	13.0	8.96	550	33	36	44	22.2	0.0	0.01	0.5	305	32	7	<0.010	0.2	360	1.3	0.013
278961	208364	LAKE		9/9/2014	15.2	9.38	2,425	19	101	460	24.0	<0.038	<0.005	4.3	471	775	39	<0.010	0.28	1,824	9.3	0.016
278962	208366	POND		9/9/2014	13.2	9.26	2,601	19	81	516	38.5	<0.075	0.01	8.0	626	757	43	<0.010	0.25	1,927	11.5	0.017
279776	208370	LAKE		9/18/2014	23.7	8.95	9,367	17	381	2,005	50.0	<0.150	<0.020	11.9	1,382	2,853	89	<0.050	0.28	6,339	21.7	0.009
890926	1989Q1375	LAKE		10/13/1989	12.4	10.1	4,626	12	656	27,600	724.0	0.1	0.01	0.3	5,490	47,600	4,070	2.94	0.2	86,539	229.8	0.880

Note. Units are mg/L unless noted.

**APPENDIX C**  
**SOUTH MEDICINE LAKE AQUIFER TESTS**

Data loggers measured water levels at 15-min intervals during normal seasonal crop irrigation at the four irrigation well sites in the South Medicine Lake management area. Drawdown was recorded at nearby monitoring wells that are completed in the same aquifer as the irrigation well. The test data were interpreted using the Cooper–Jacob analytical method. Tables summarizing results are shown in the main body of this report (tables 4–7, p. 39–40).

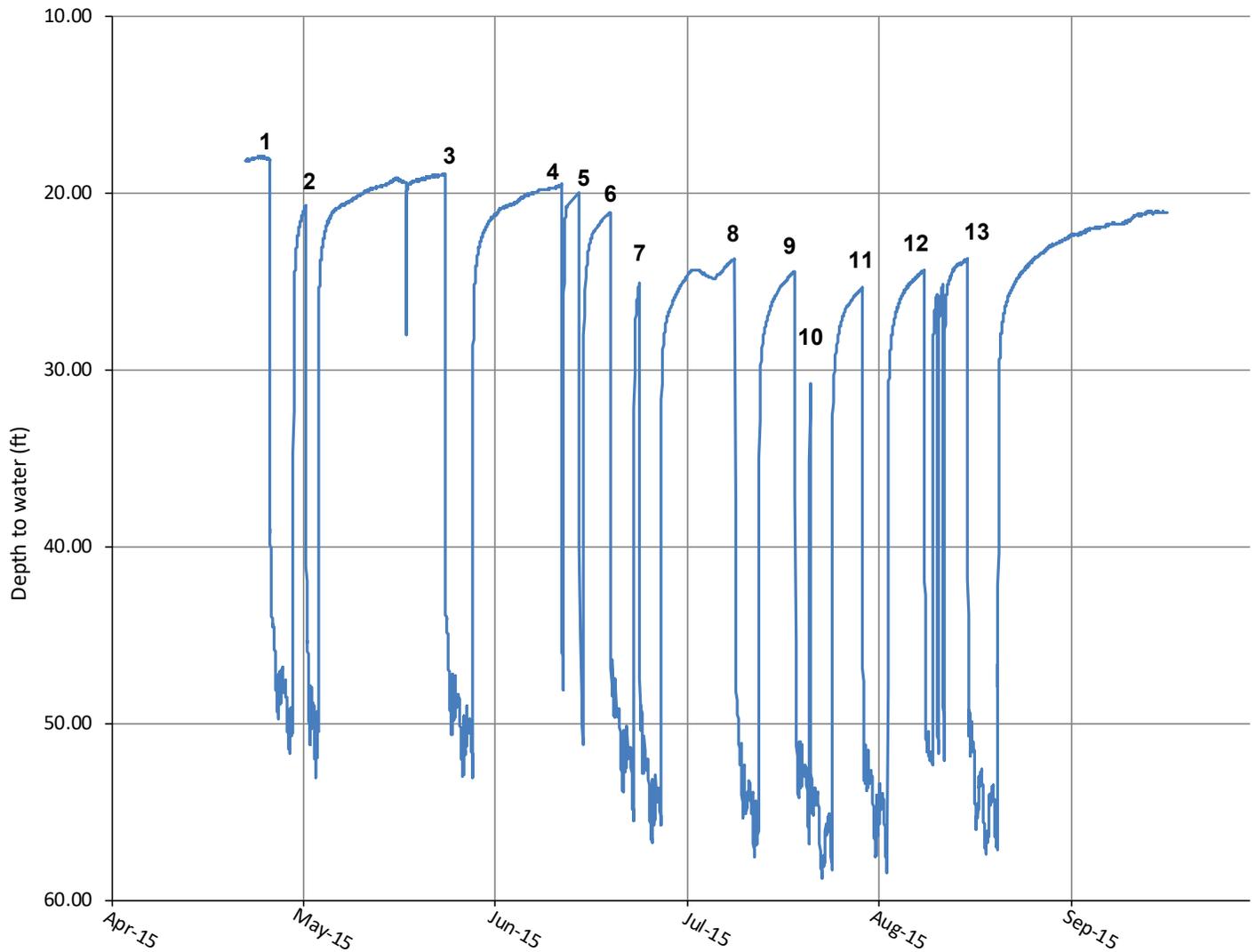


Figure C1-1. Timing of 13 aquifer tests at Nelson 1, pumping irrigation well 121117 and monitoring drawdown at well 43113 located 25 ft away from the irrigation well.

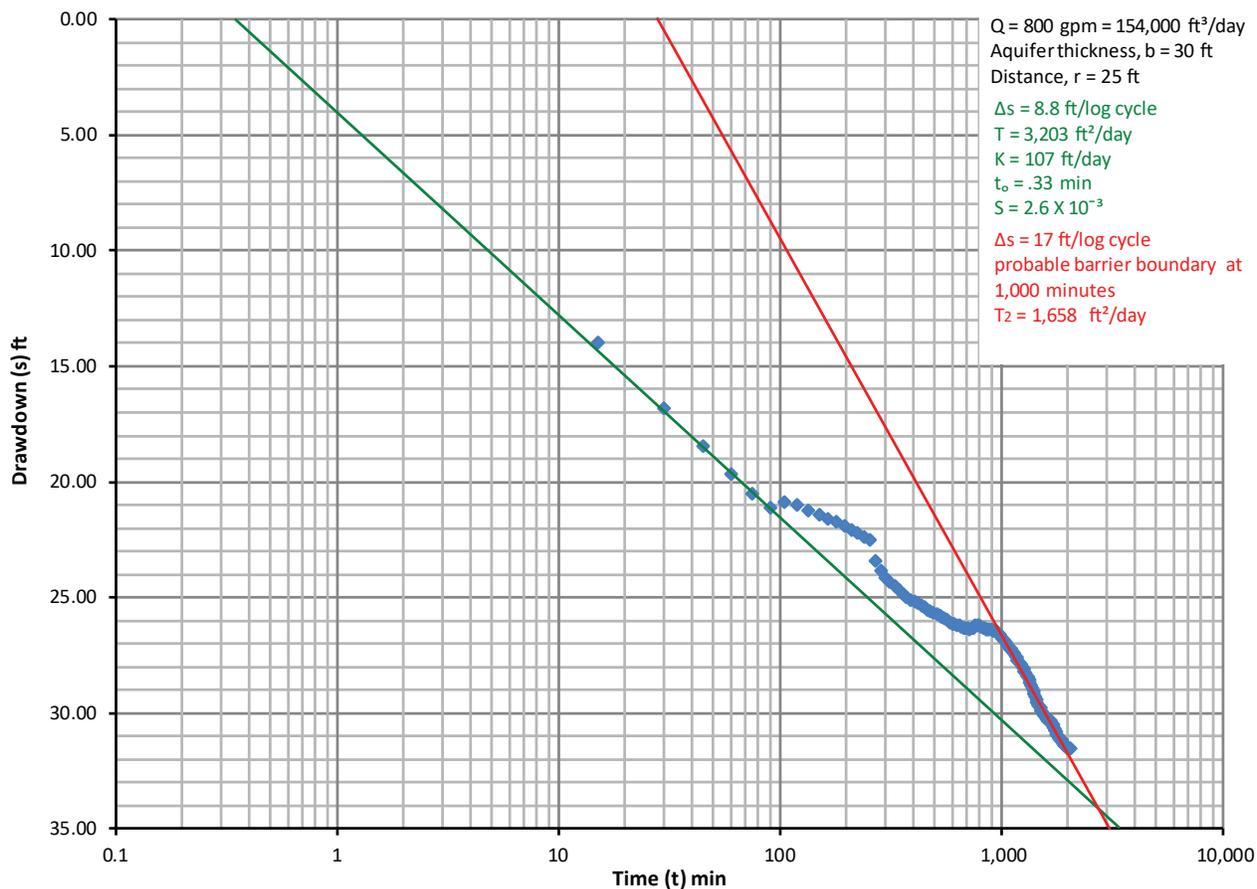


Figure C1-2. Results of Nelson 1 aquifer test #1.

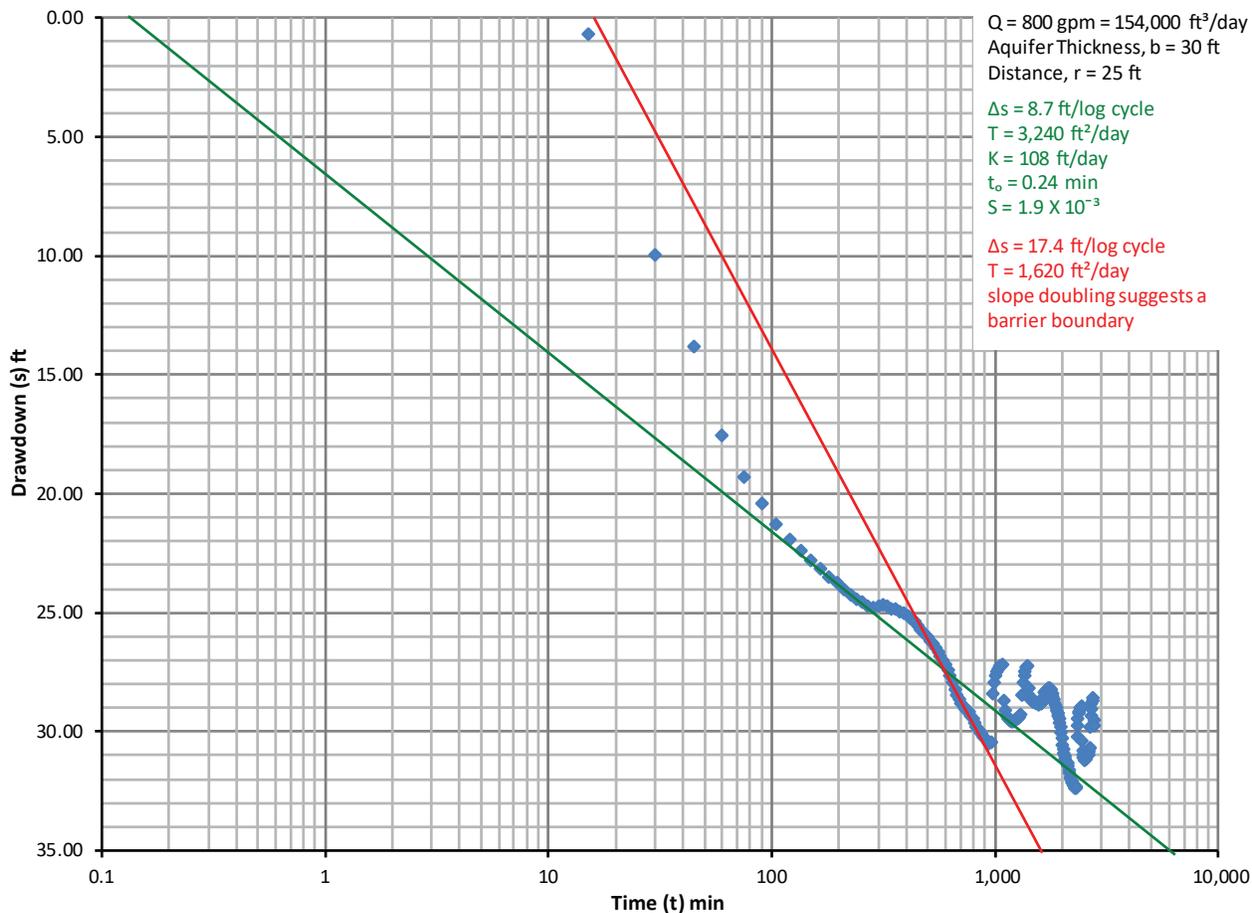


Figure C1-3. Results of Nelson 1 aquifer test #2.

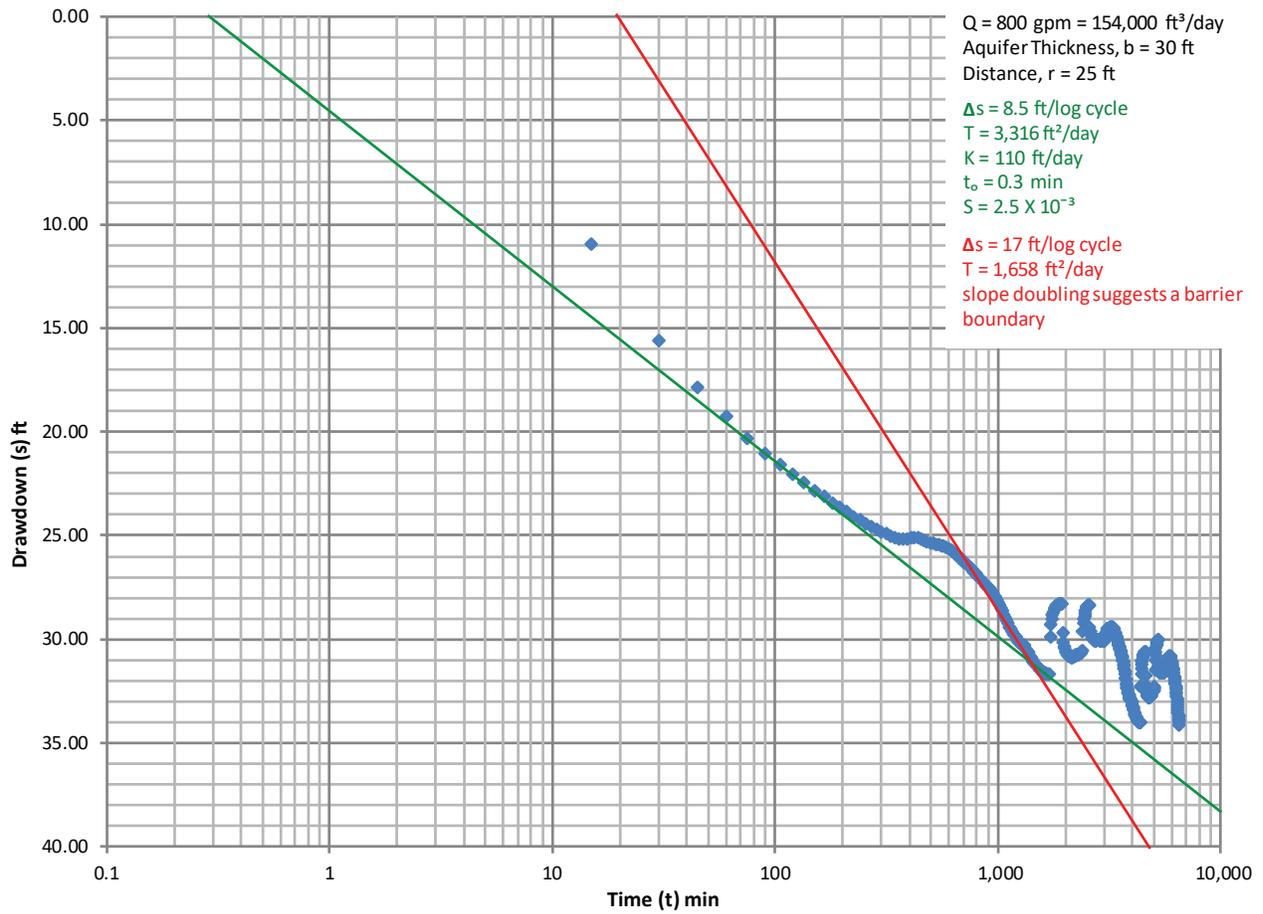


Figure C1-4. Results of Nelson 1 aquifer test #3.

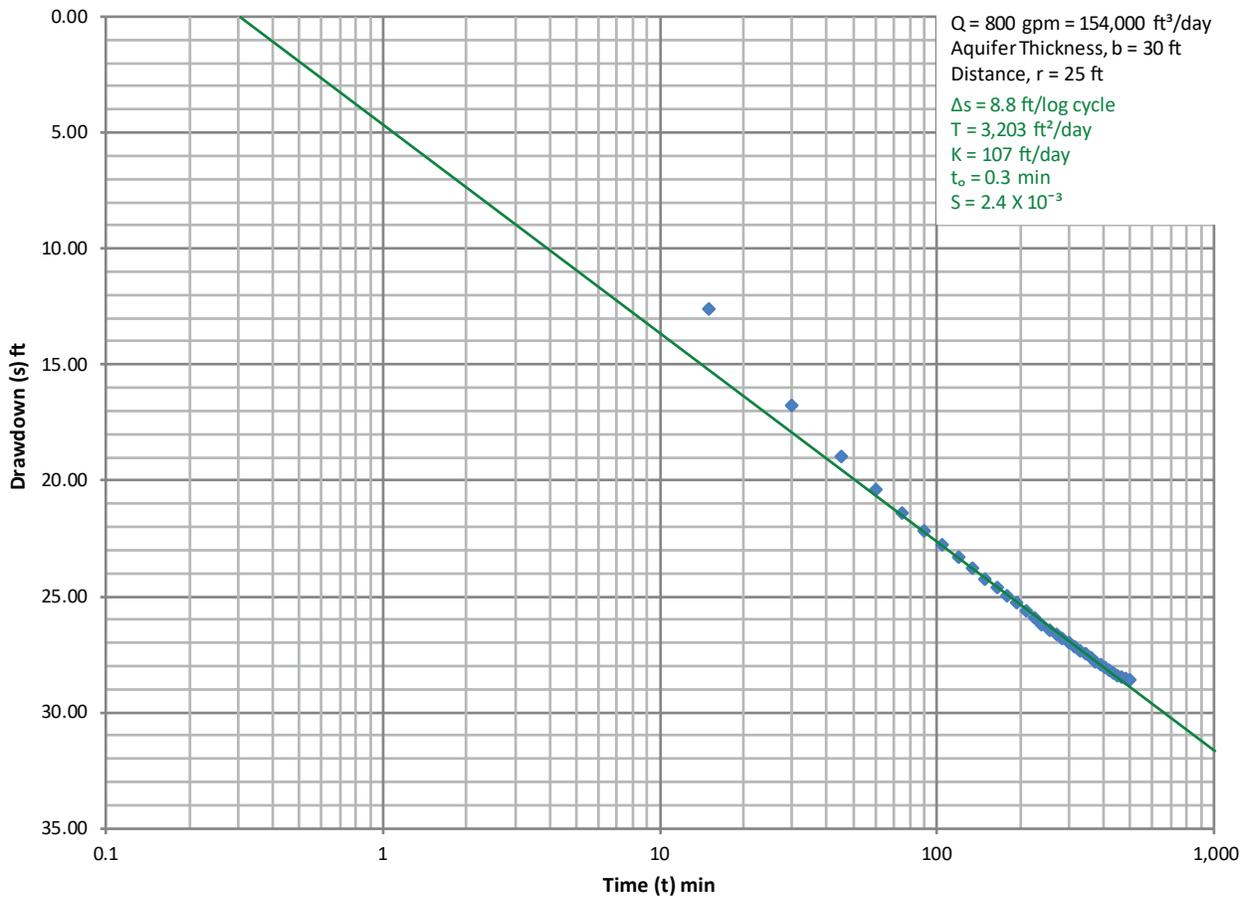


Figure C1-5. Results of Nelson 1 aquifer test #4.

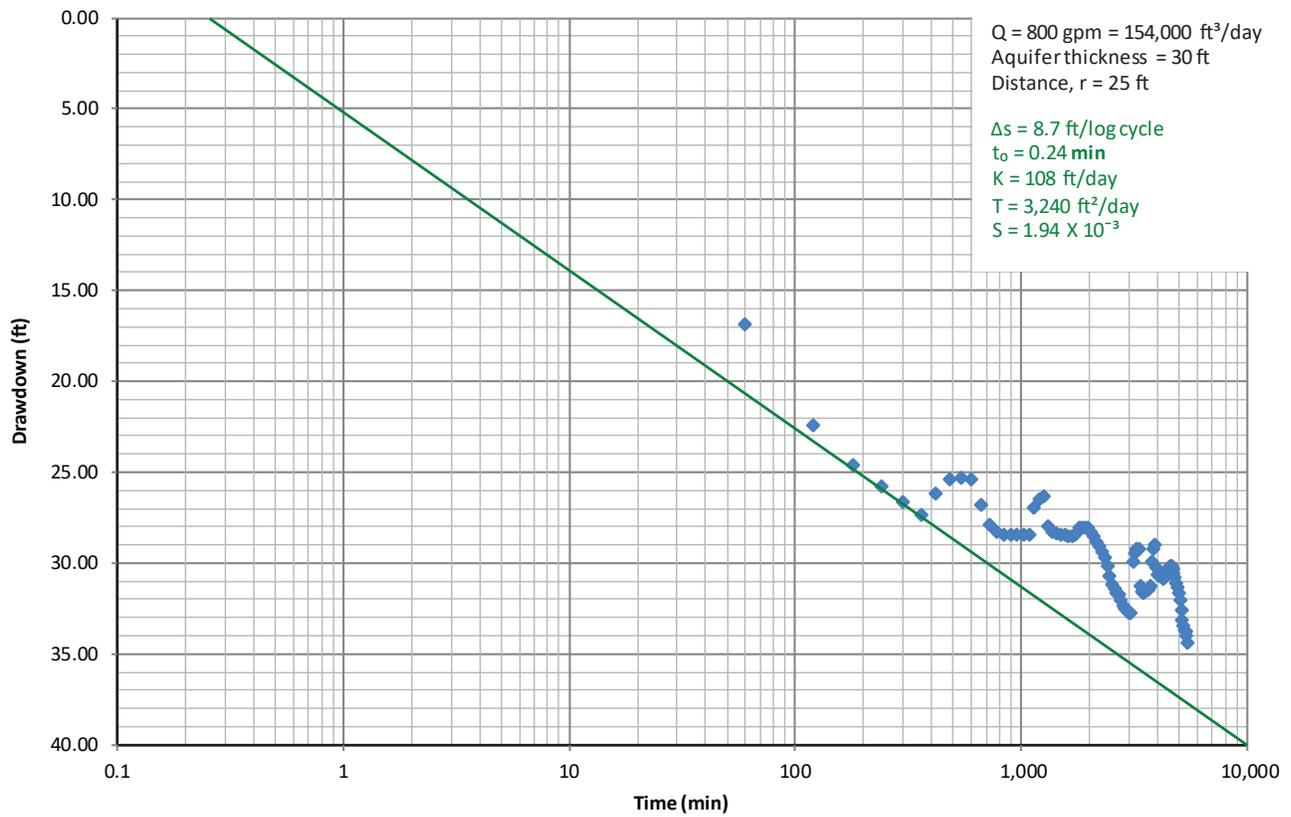


Figure C1-6. Results of Nelson 1 aquifer test #5.

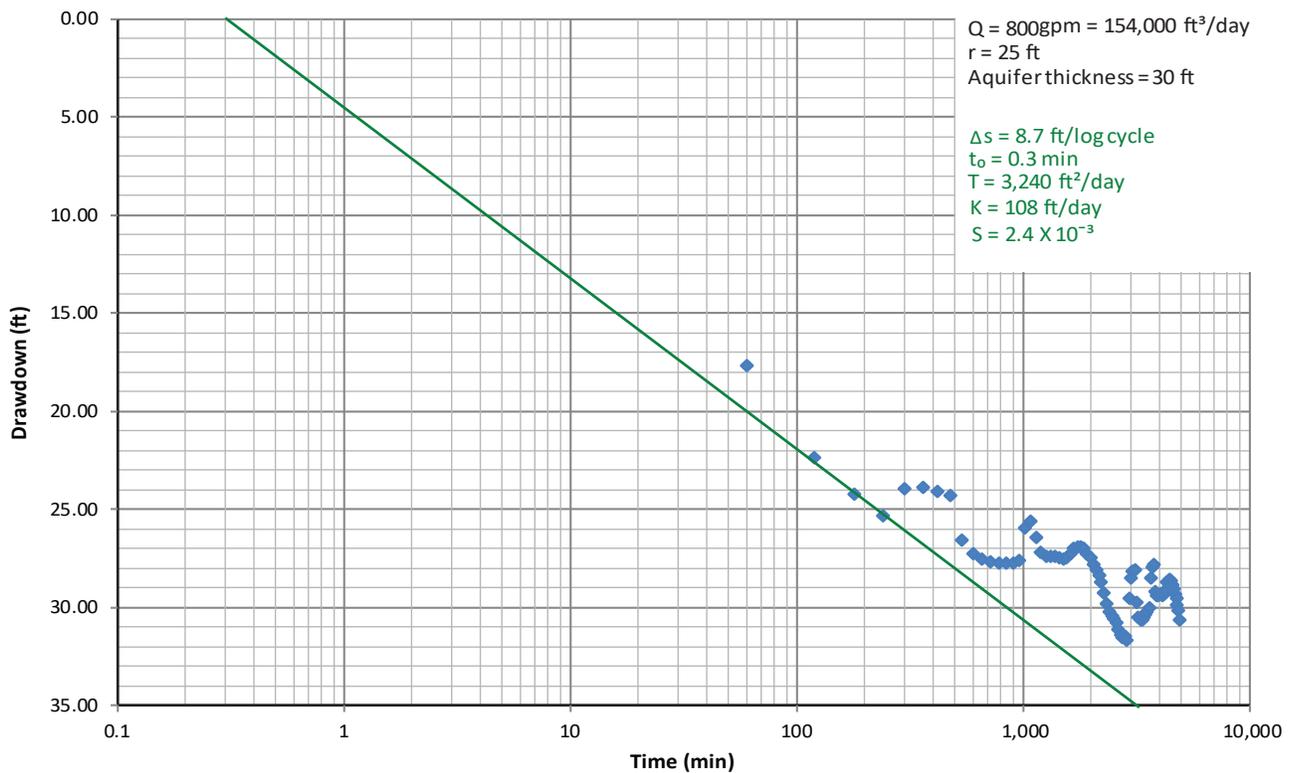


Figure C1-7. Results of Nelson aquifer test #6.

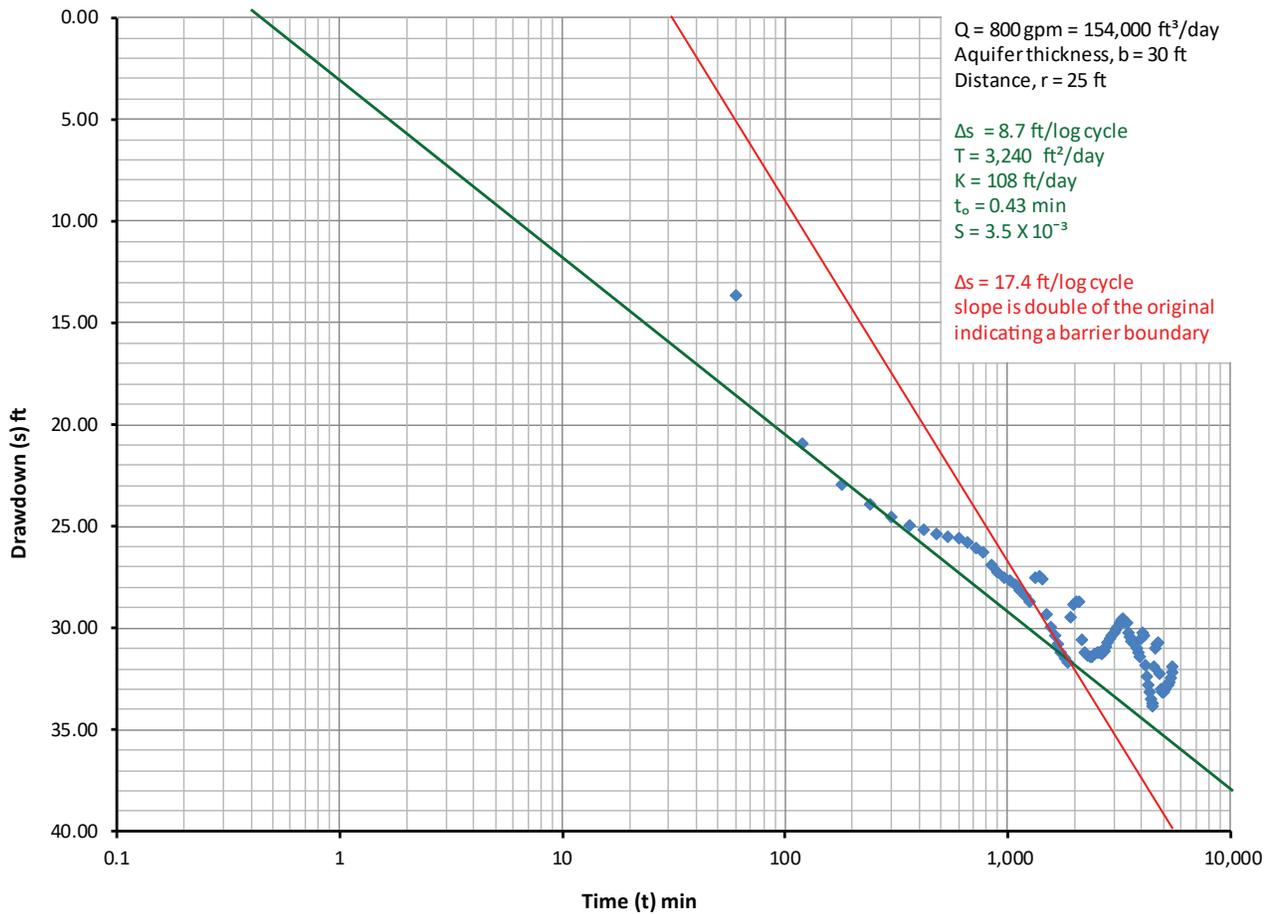


Figure C1-8. Results of Nelson 1 aquifer test #7.

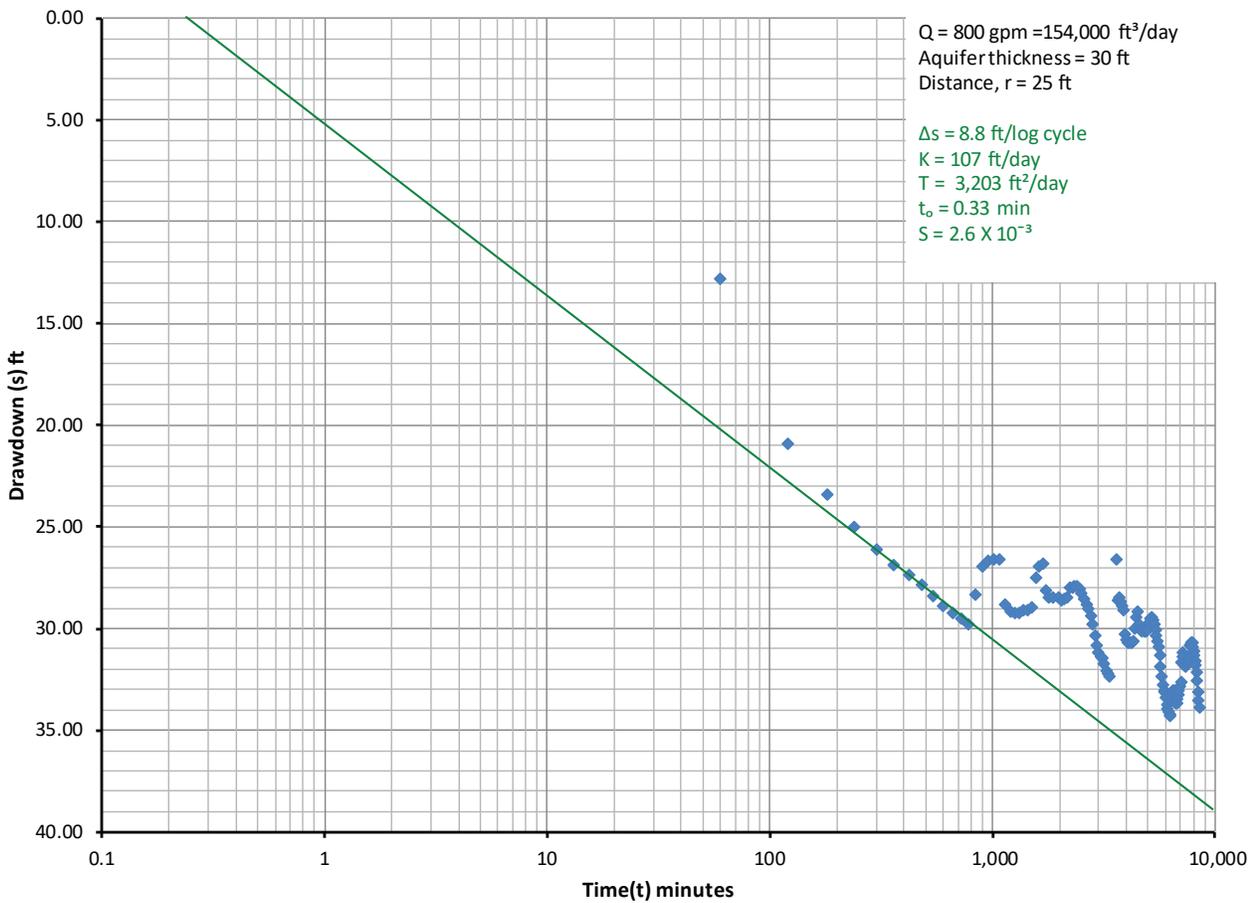


Figure C1-9. Results of Nelson 1 aquifer test #8.

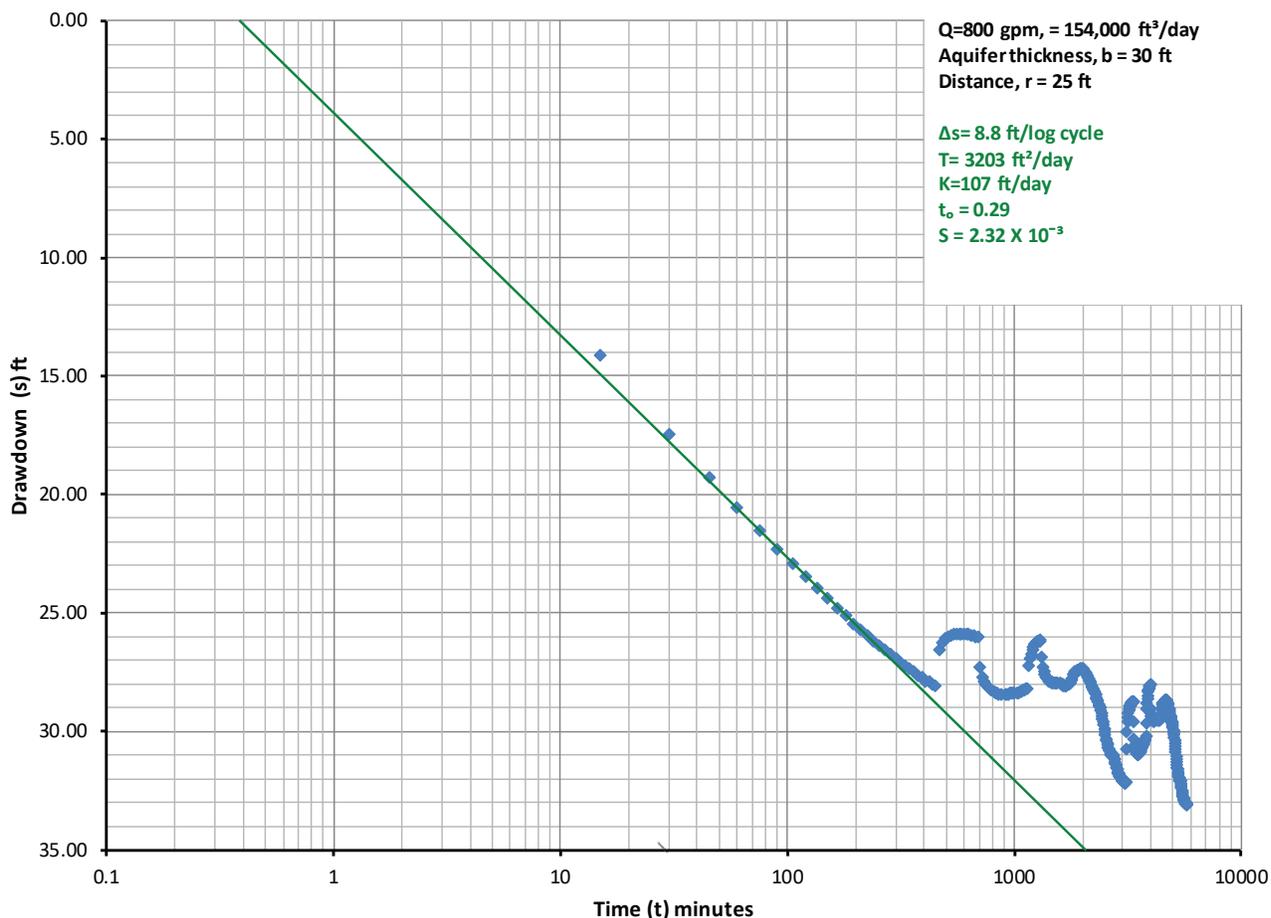


Figure C1-10. Results of Nelson 1 aquifer test #9.

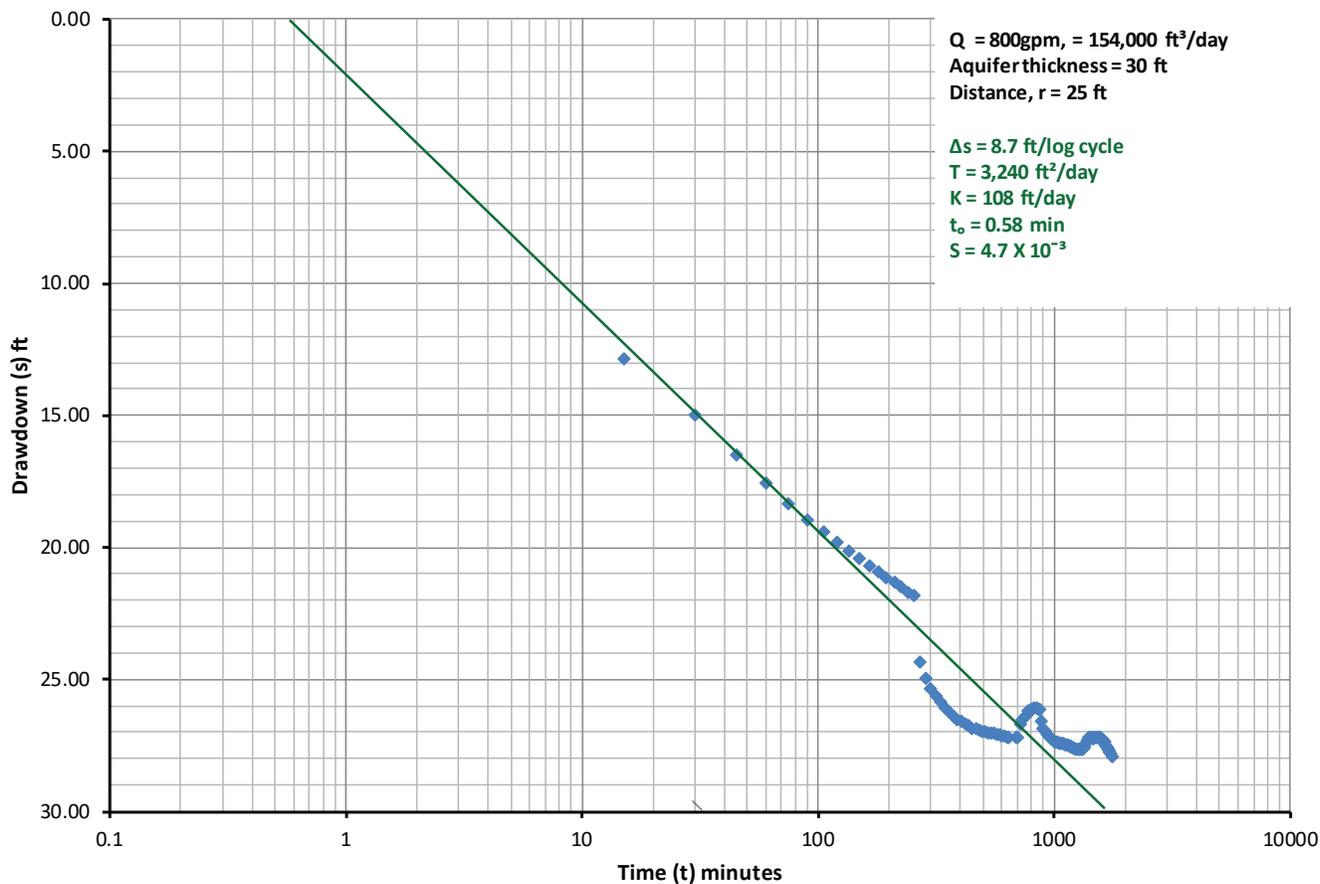


Figure C1-11. Results of Nelson 1 aquifer test #10.

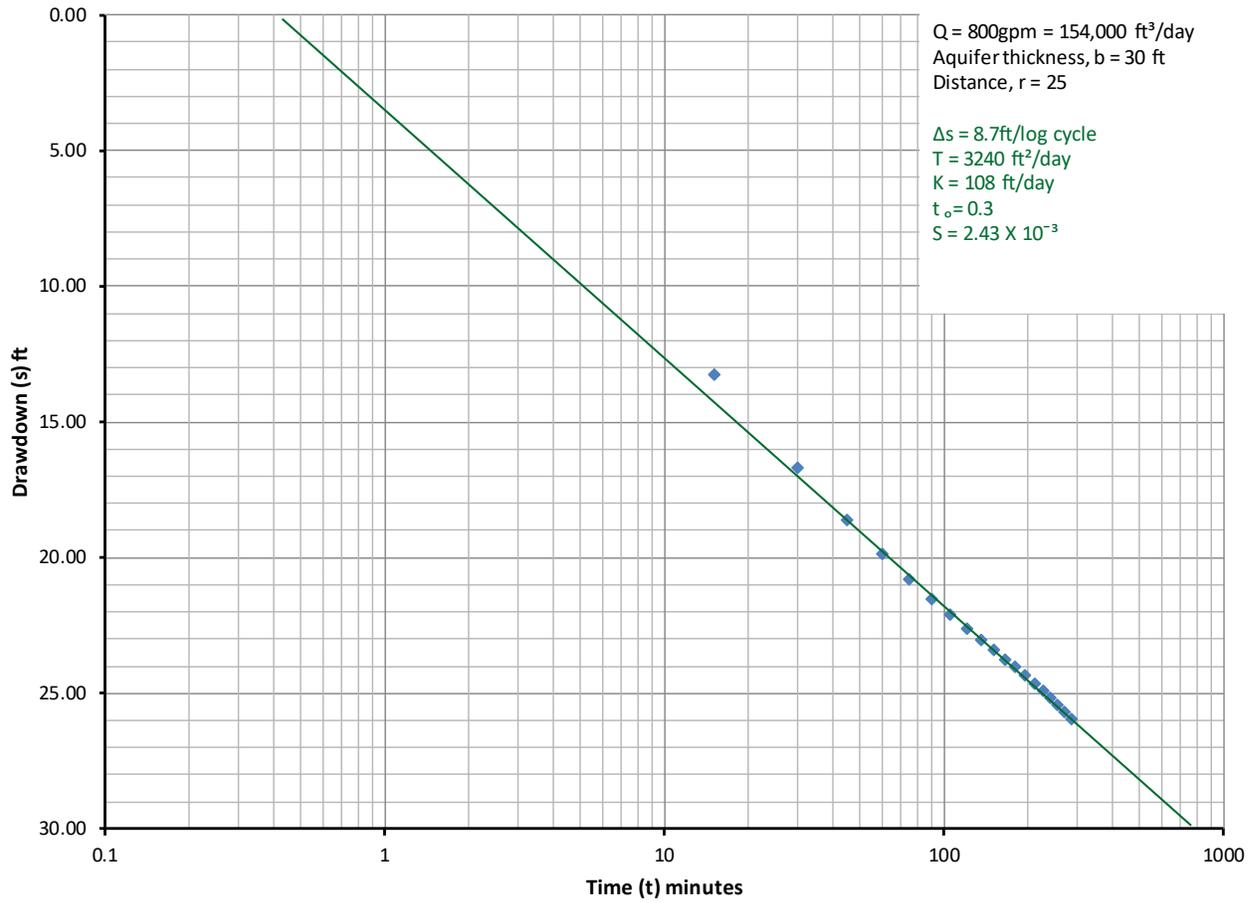


Figure C1-12. Results of Nelson 1 aquifer test #11.

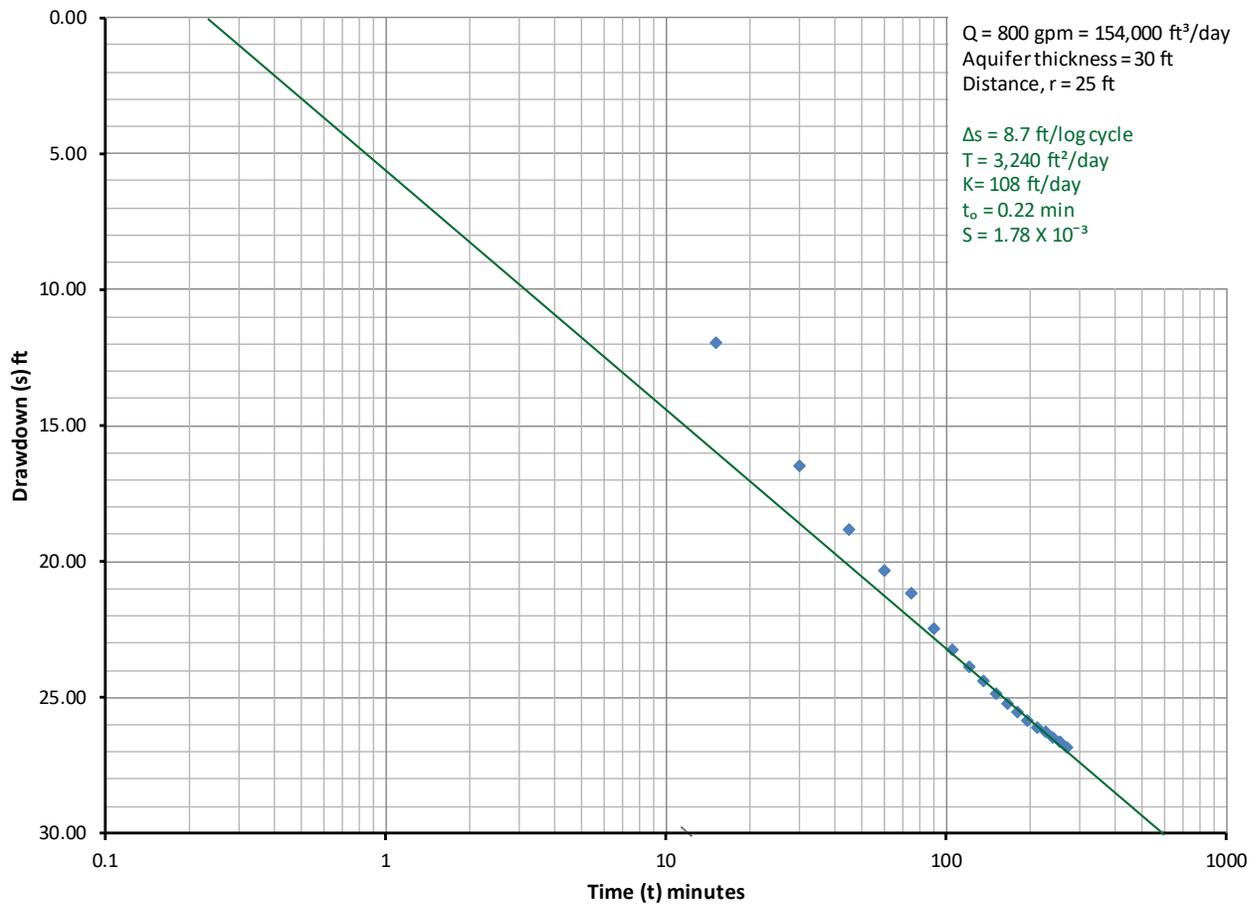


Figure C1-13. Results of Nelson 1 aquifer test #12.

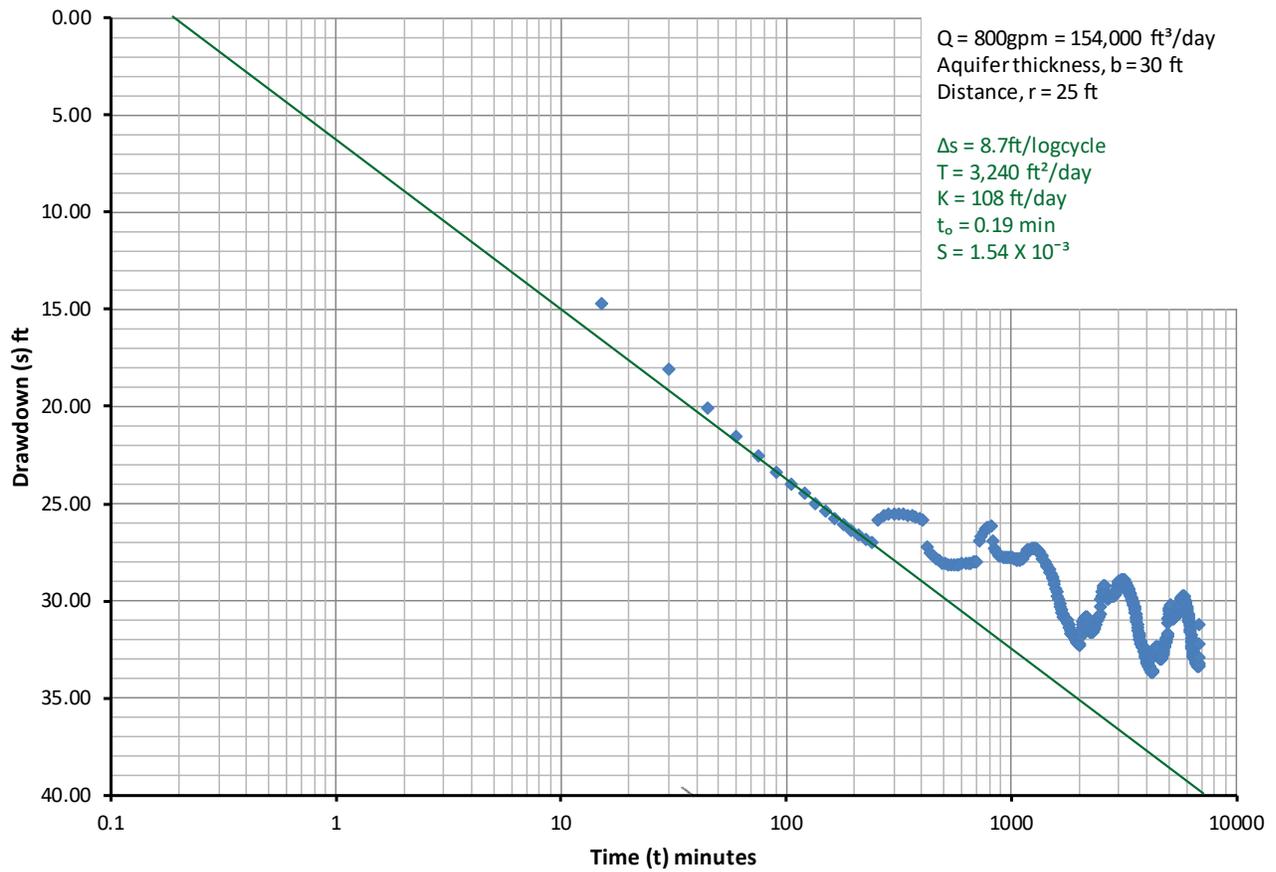


Figure C1-14. Results of Nelson 1 aquifer test #13.

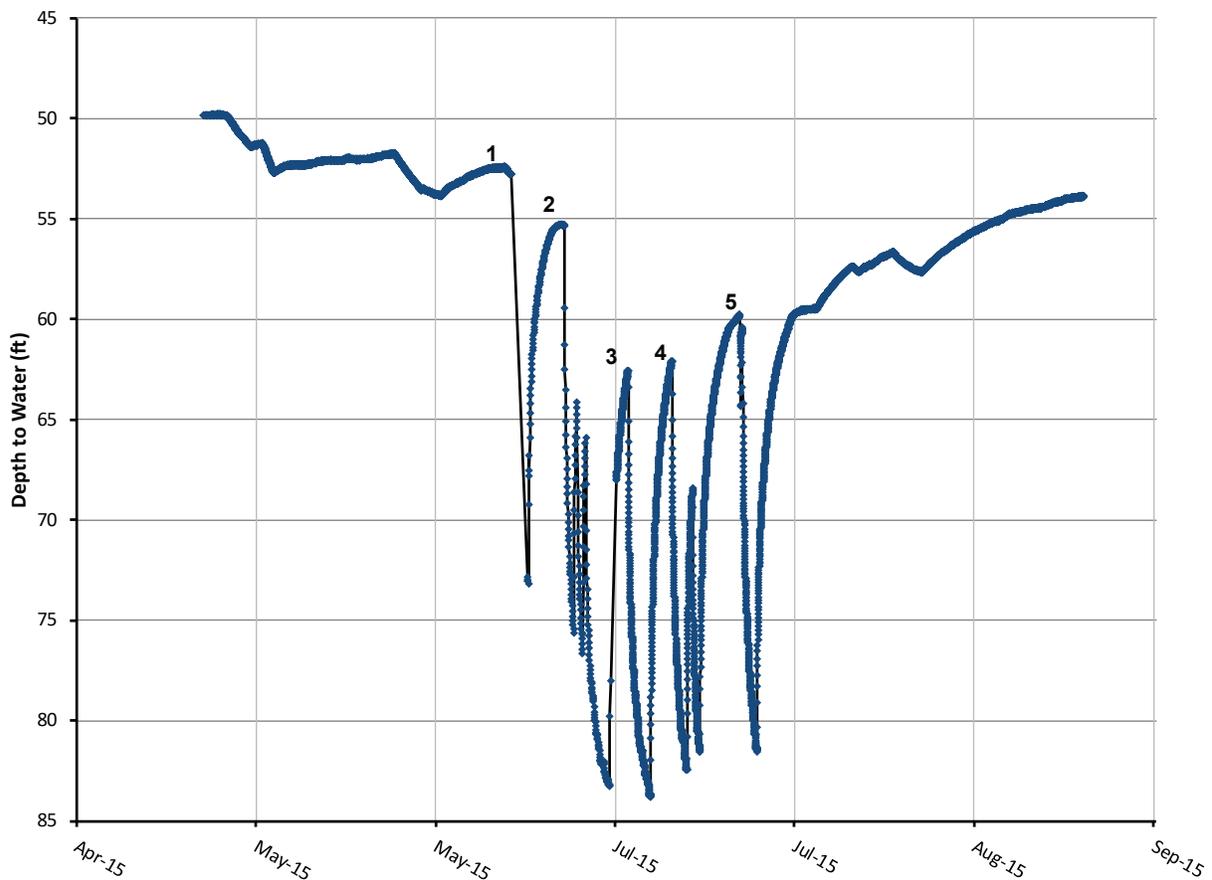


Figure C2-1. Timing of five aquifer tests at Nelson 2, pumping irrigation well 155930 and monitoring drawdown at well 155929 located 680 ft away from the irrigation well.

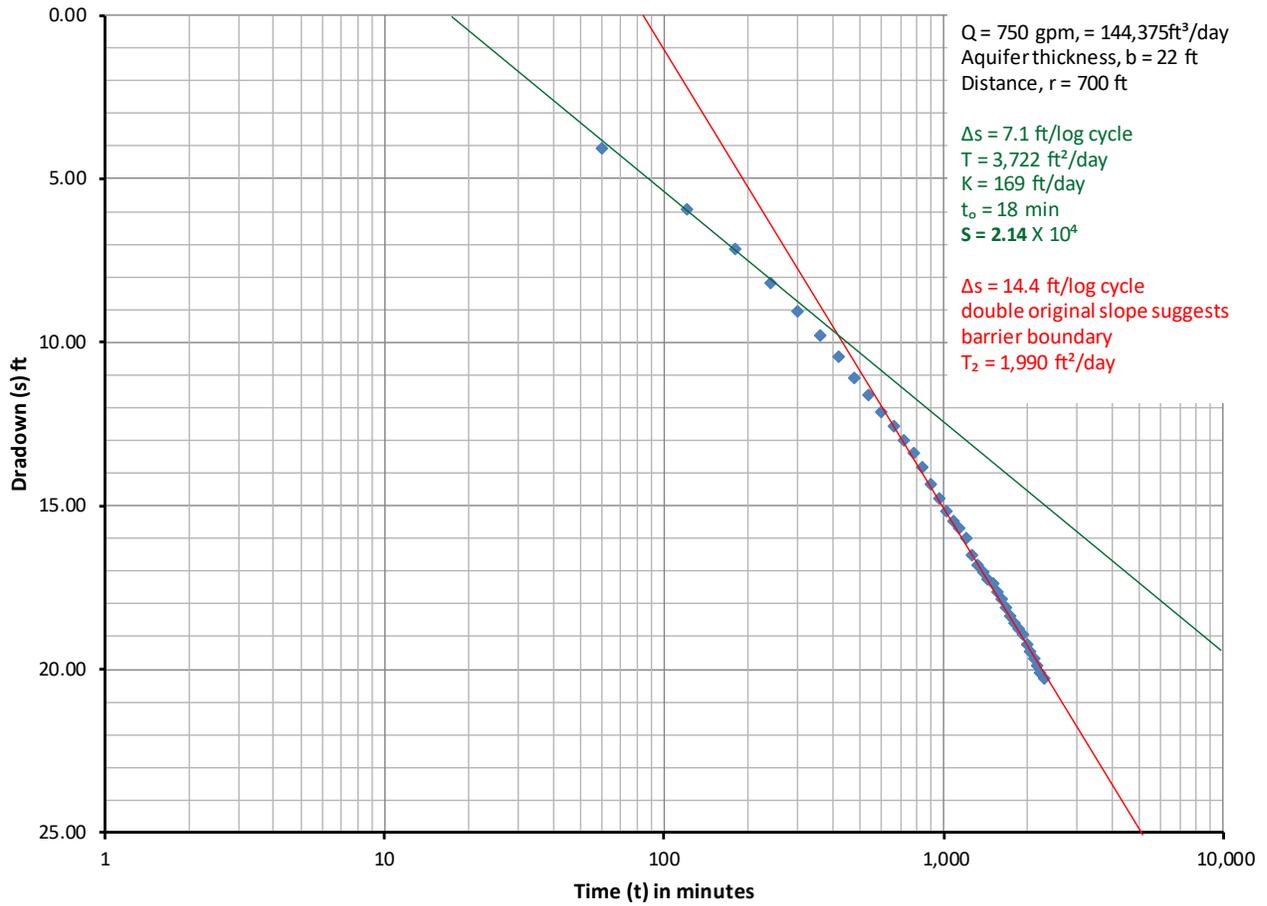


Figure C2-2. Results of Nelson 2 aquifer test #1.

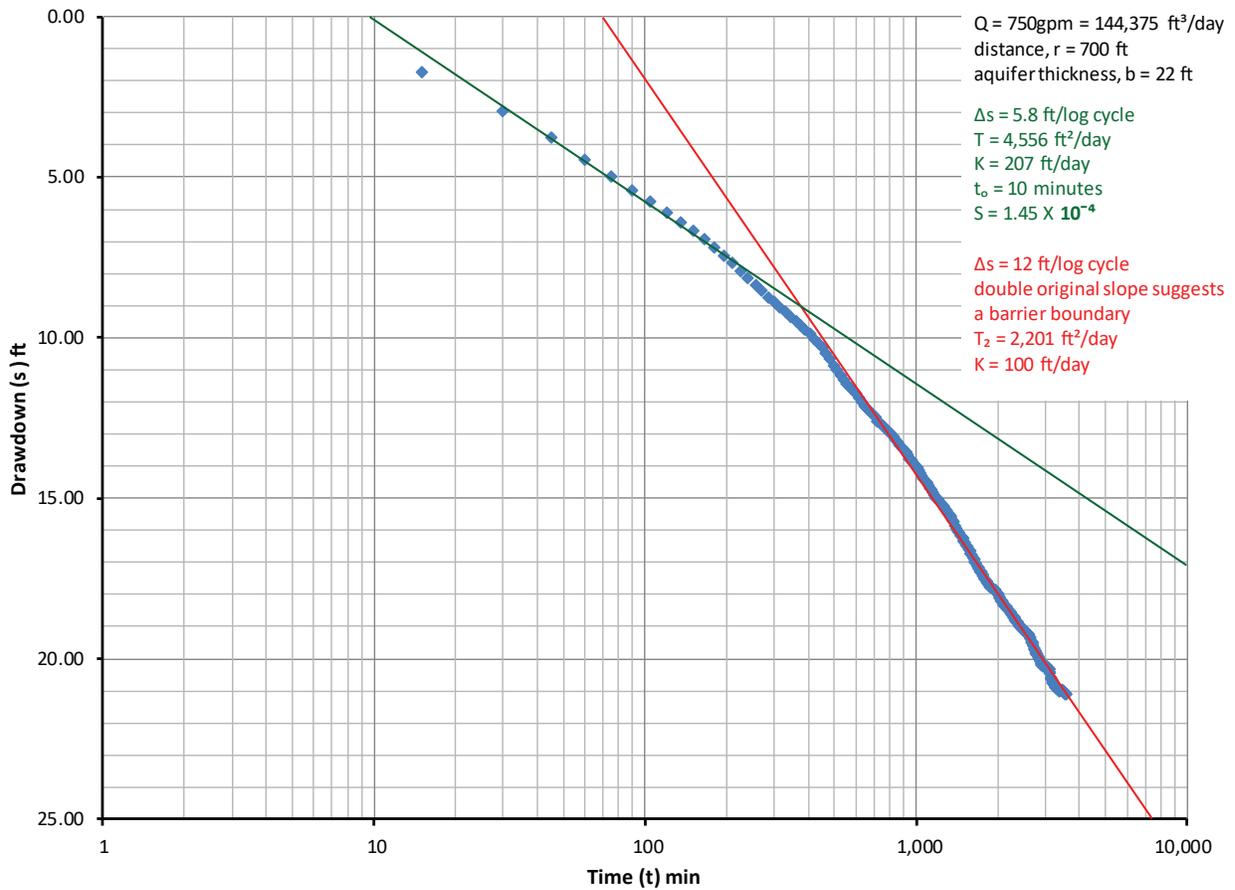


Figure C2-3. Results of Nelson 2 aquifer test #2.

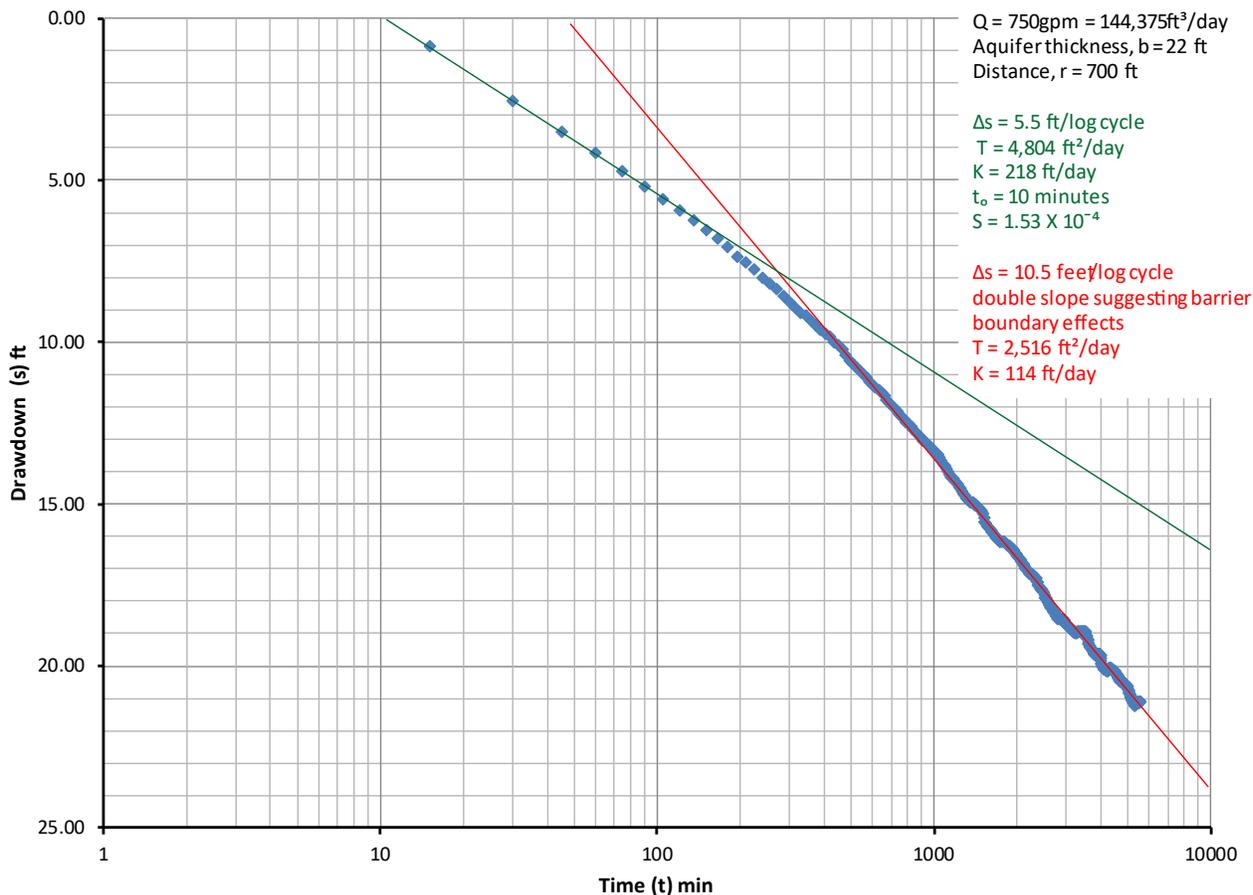


Figure C2-4. Results of Nelson 2 aquifer test #3.

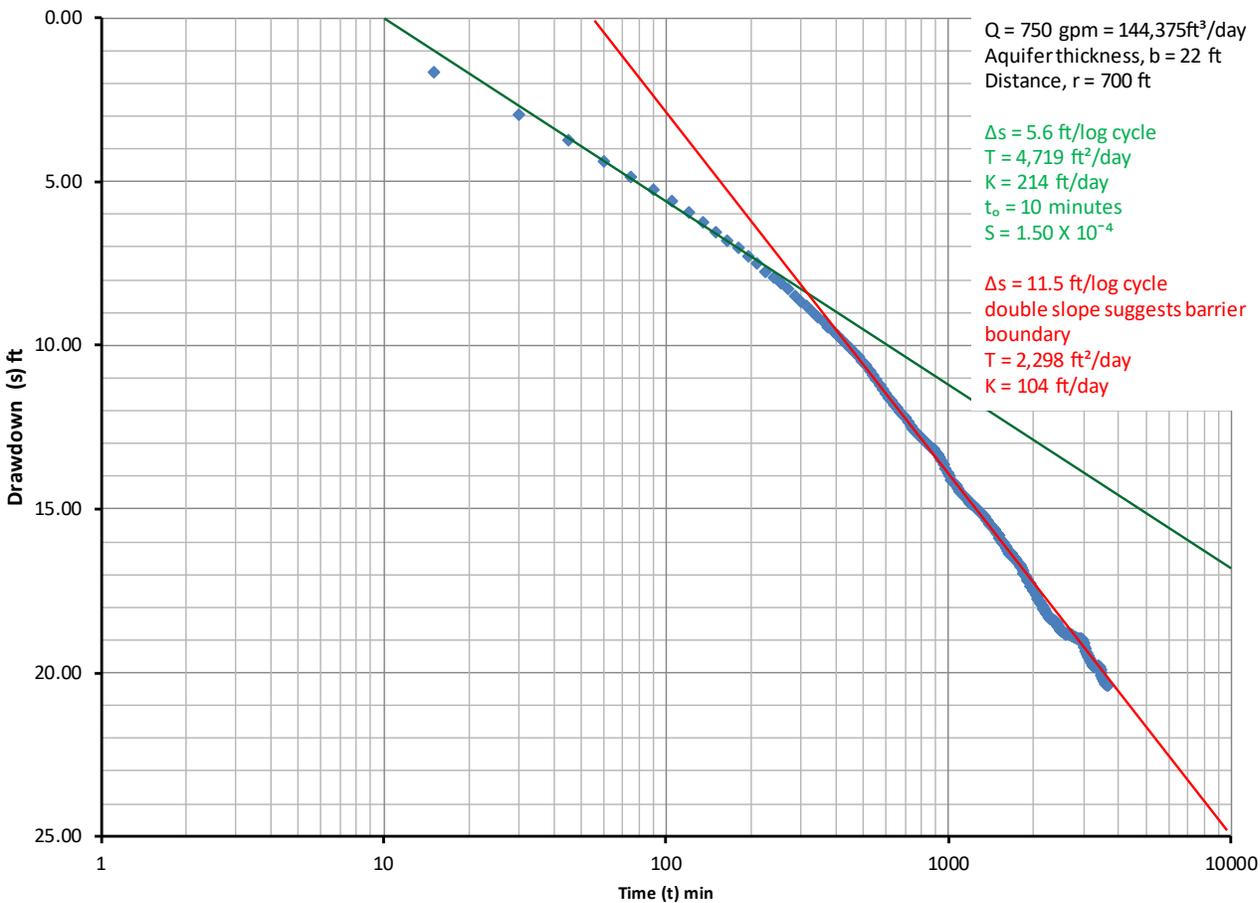


Figure C2-5. Results of Nelson 2 aquifer test #4.

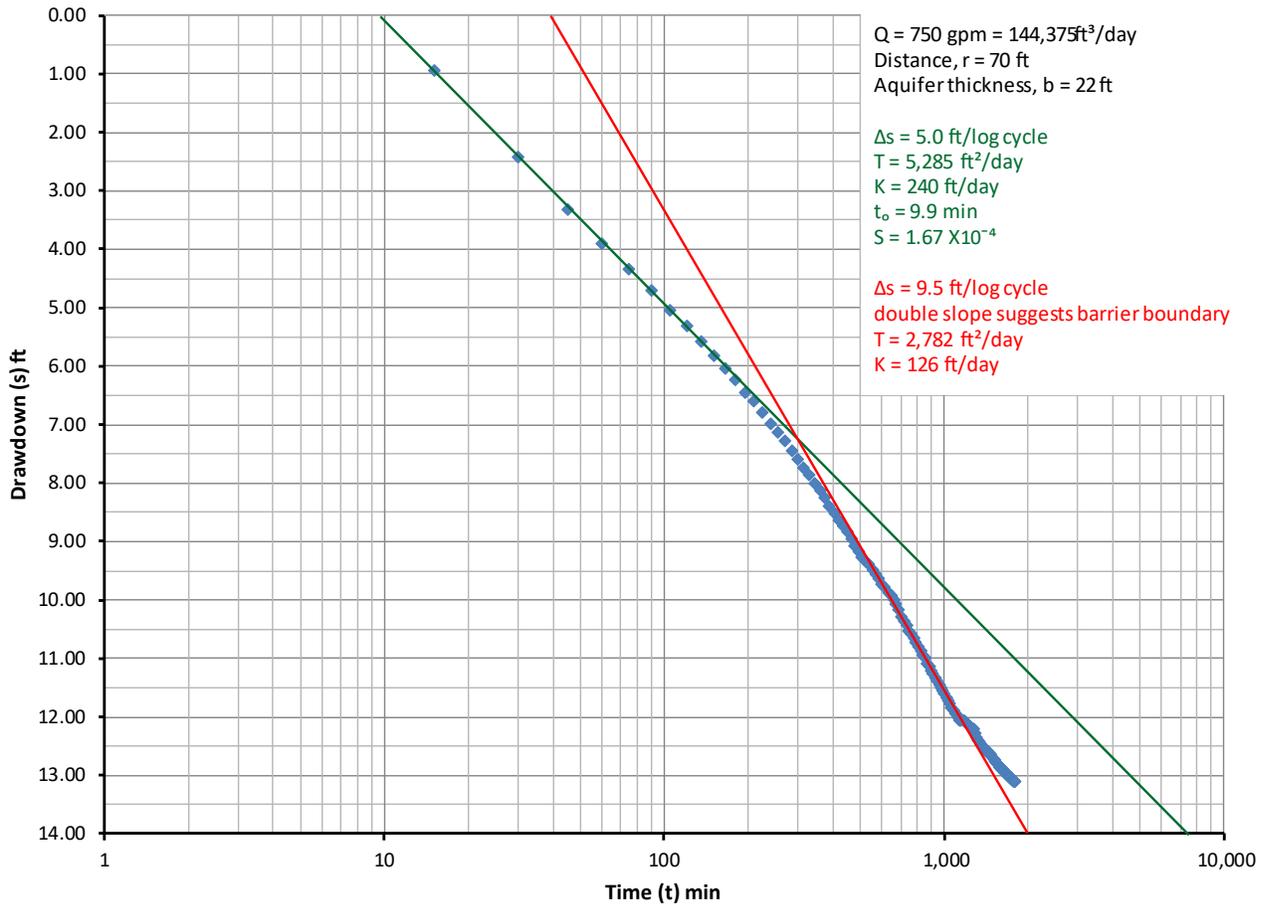


Figure C2-6. Results of Nelson 2 aquifer test #5.

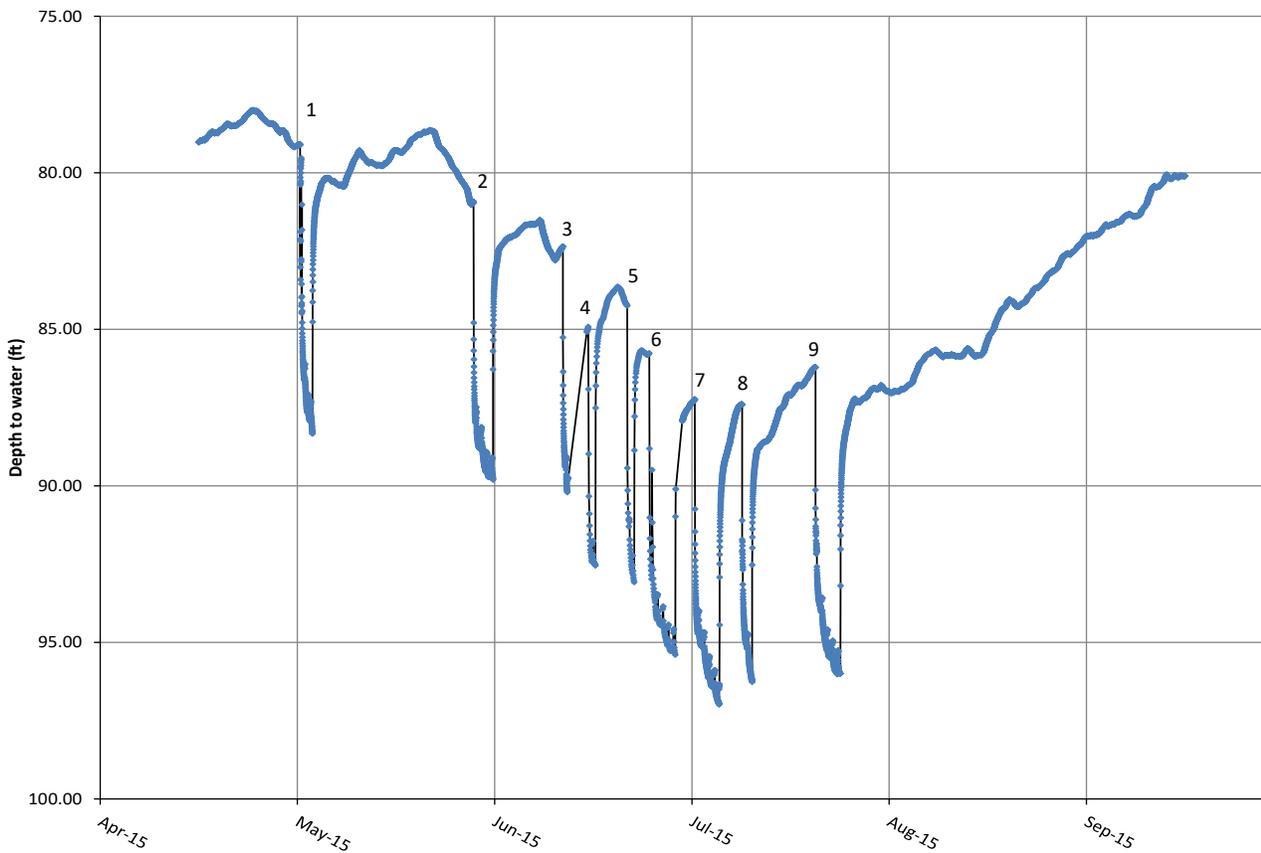


Figure C3-1. Timing of nine aquifer tests at Nelson 3, pumping irrigation well 212128 and monitoring draw-down at well 169217 located 230 ft away from the irrigation well.

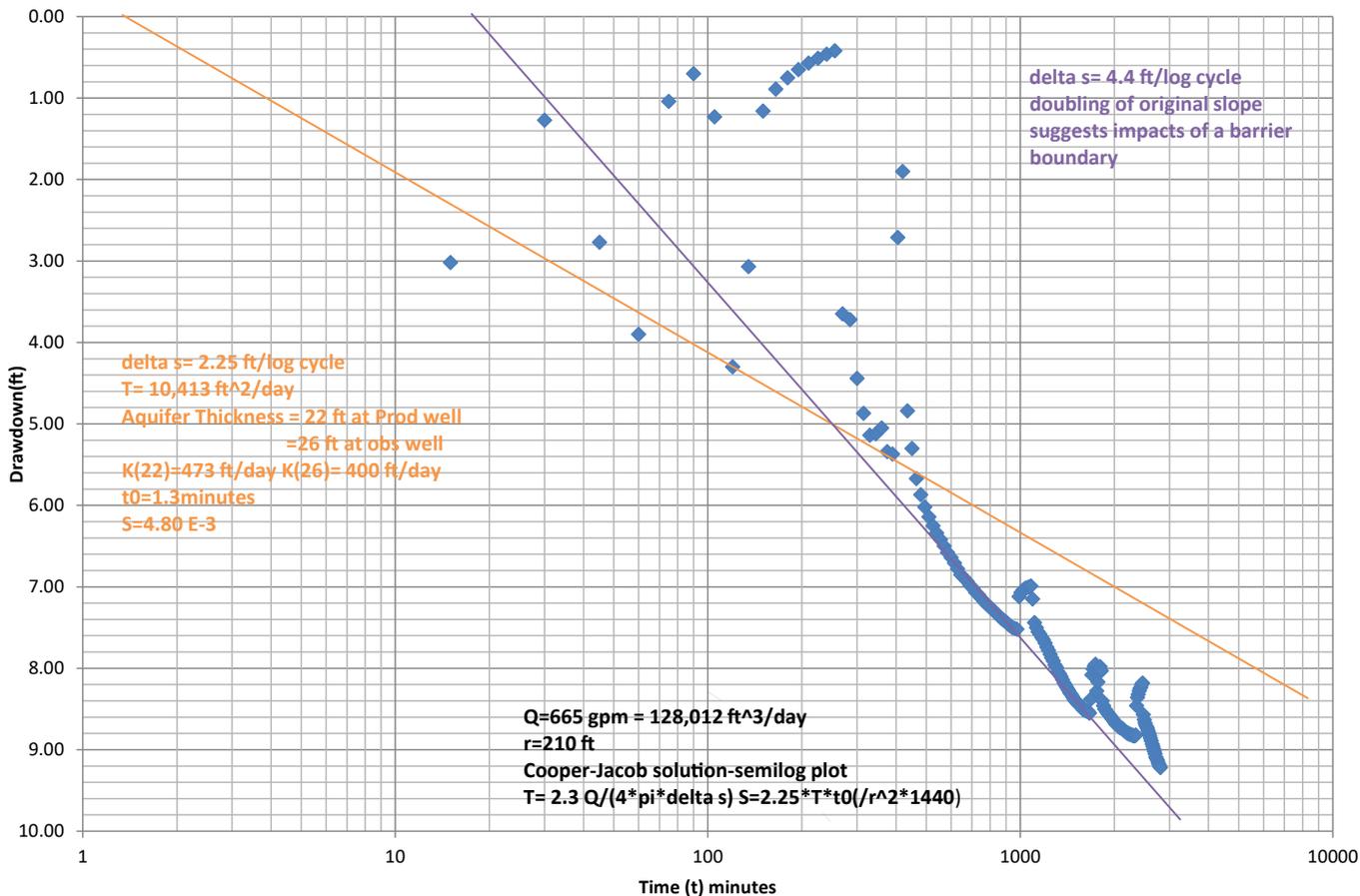


Figure C3-2. Results of Nelson 3 aquifer test #1.

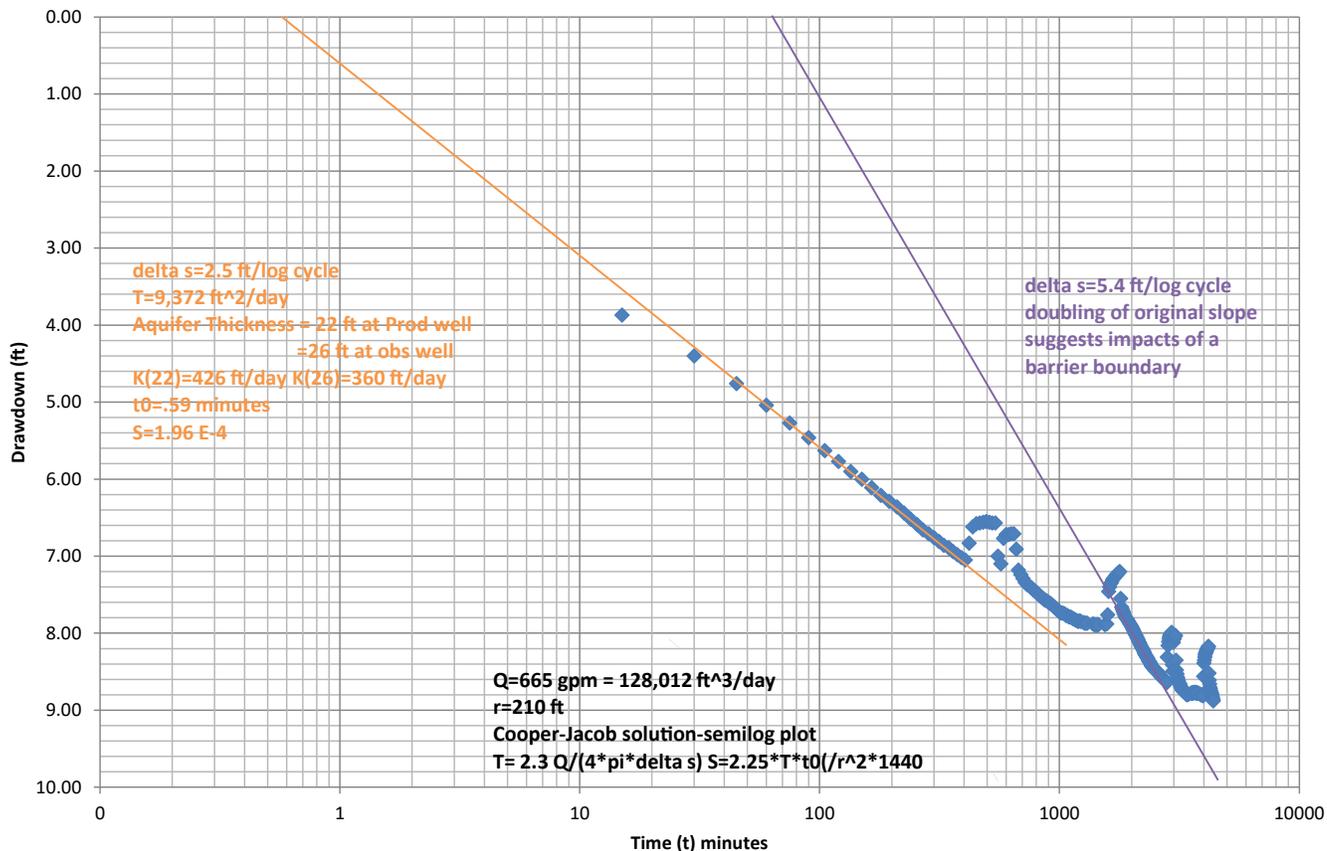


Figure C3-3. Results of Nelson 3 aquifer test #2.





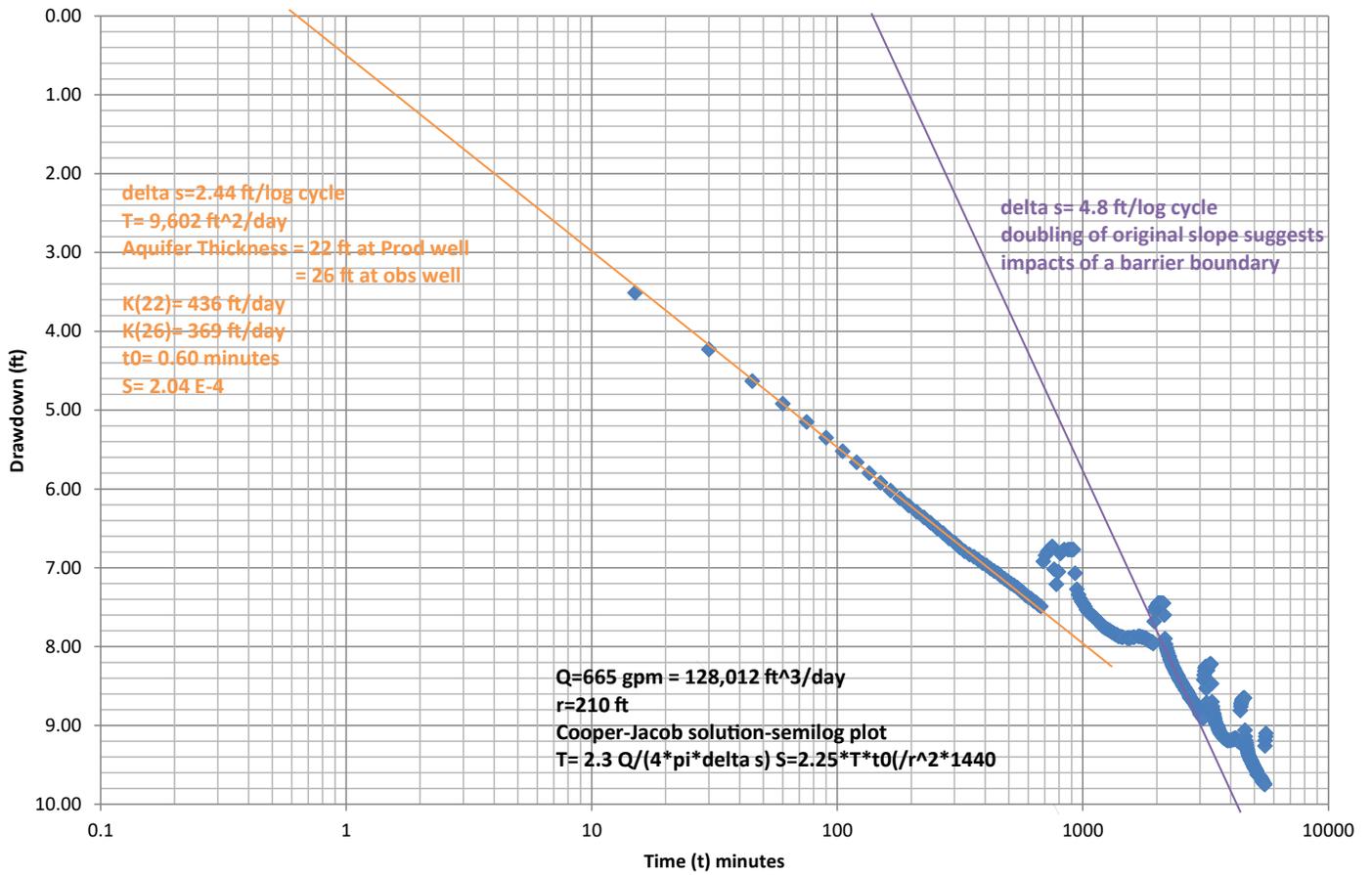


Figure C3-8. Results of Nelson 3 aquifer test #7.

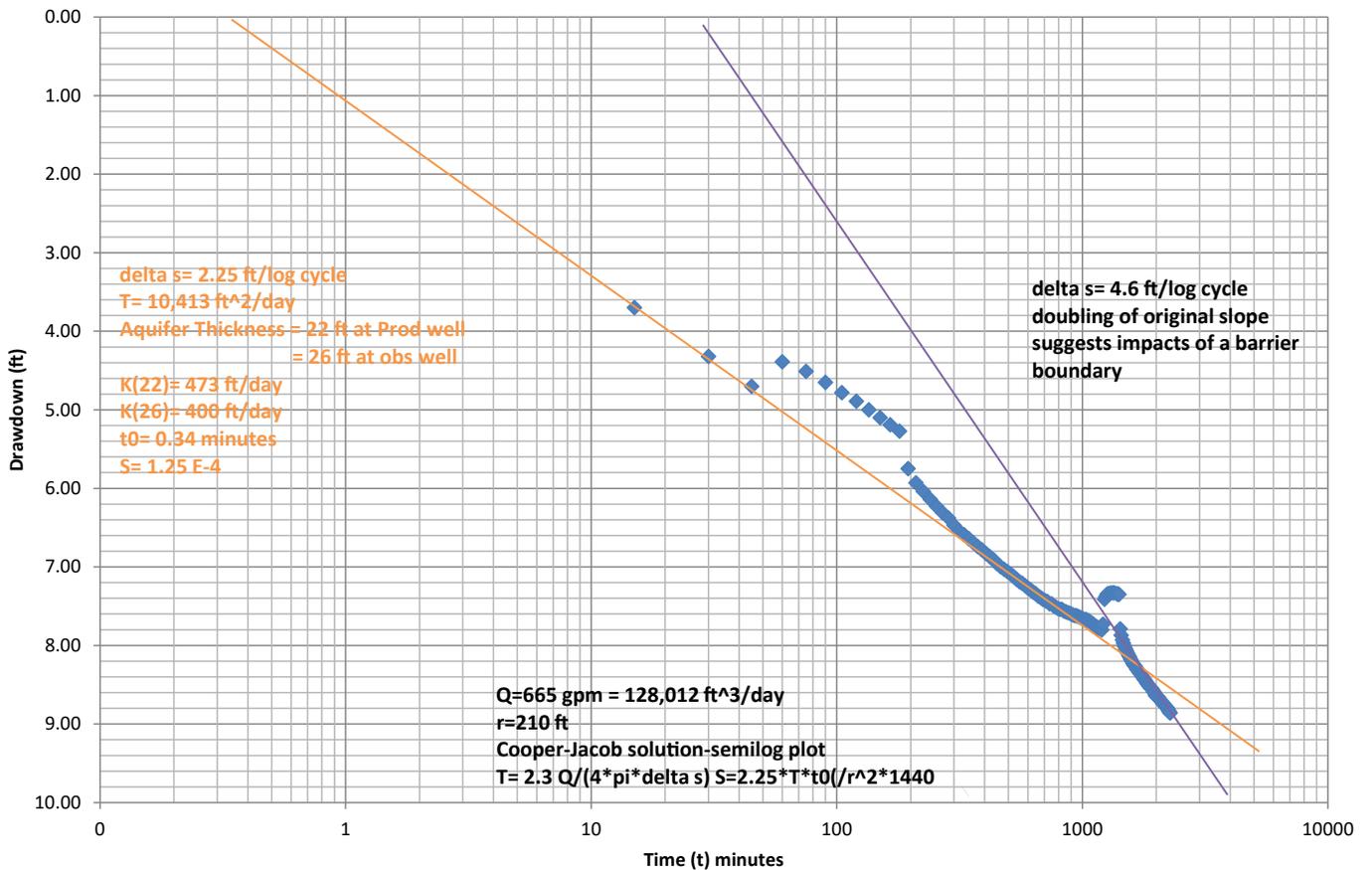


Figure C3-9. Results of Nelson 3 aquifer test #8.



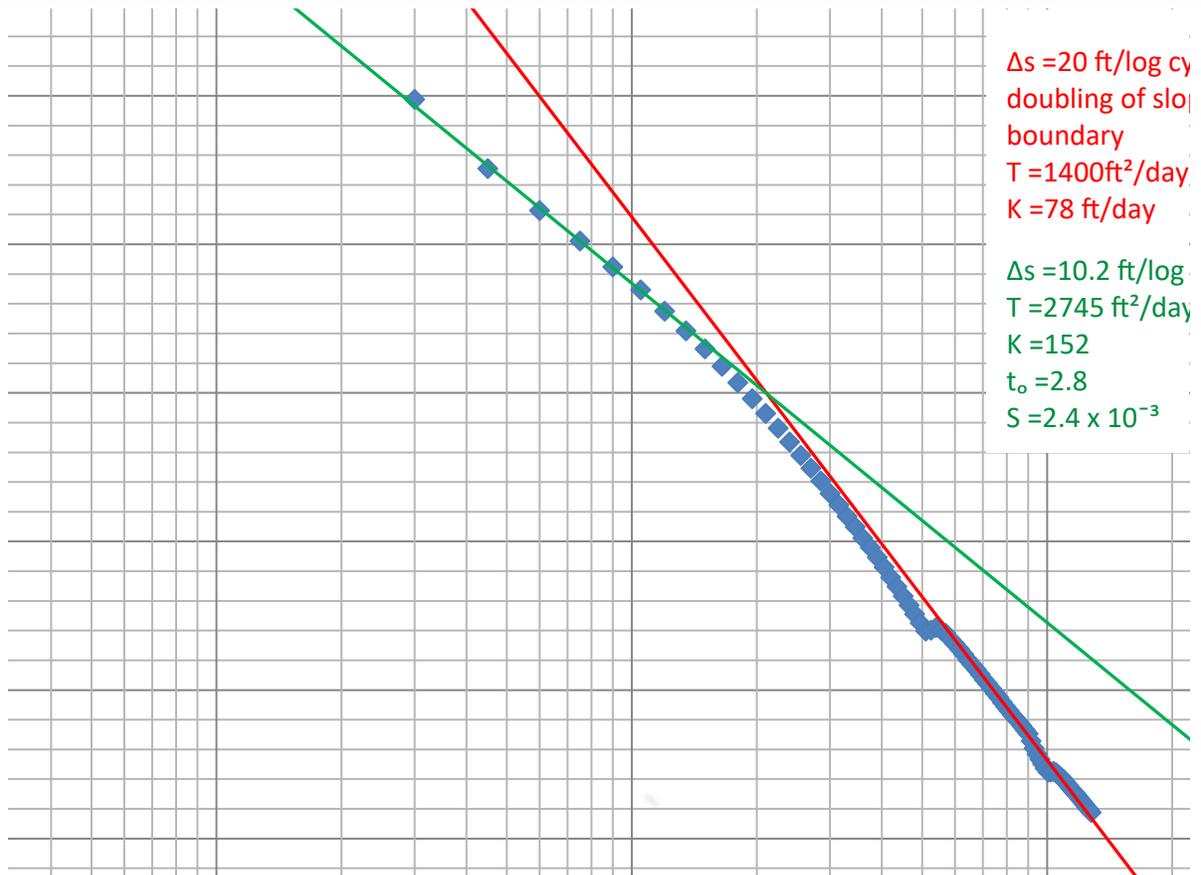


Figure C4-2. Results of Bolstad 1 aquifer test #1.

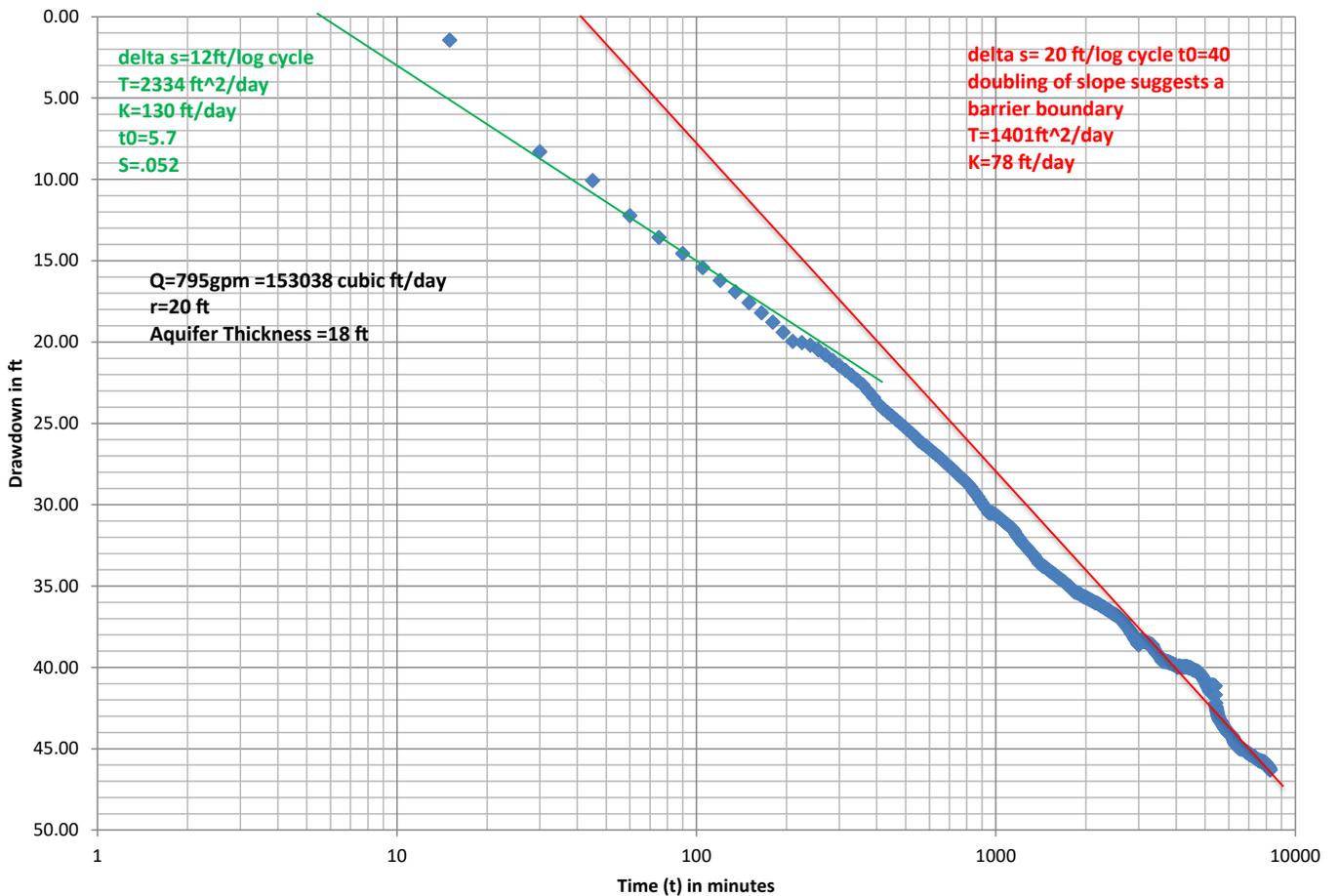


Figure C4-3. Results of Bolstad 1 aquifer test #2.

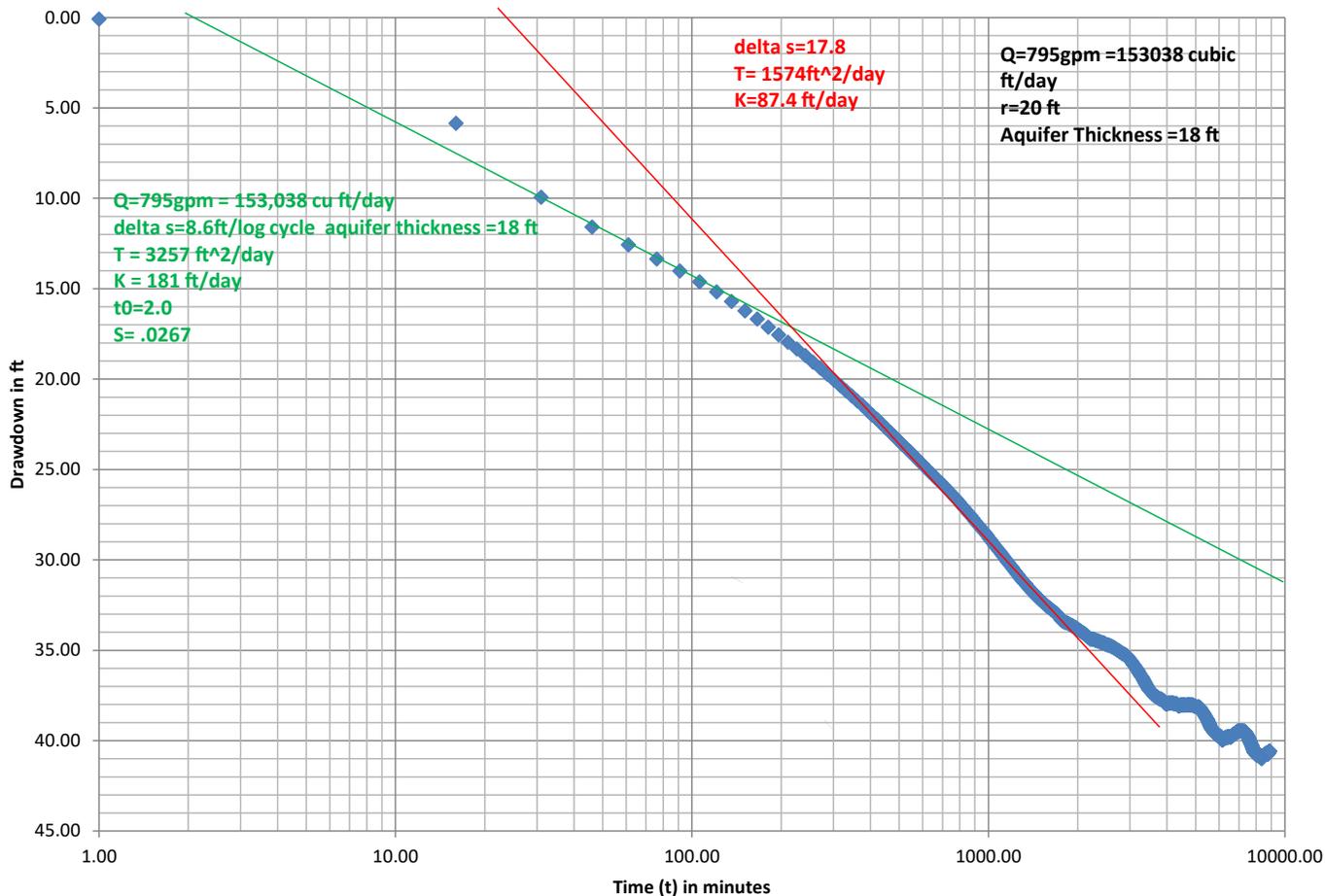


Figure C4-4. Results of Bolstad 1 aquifer test #3.

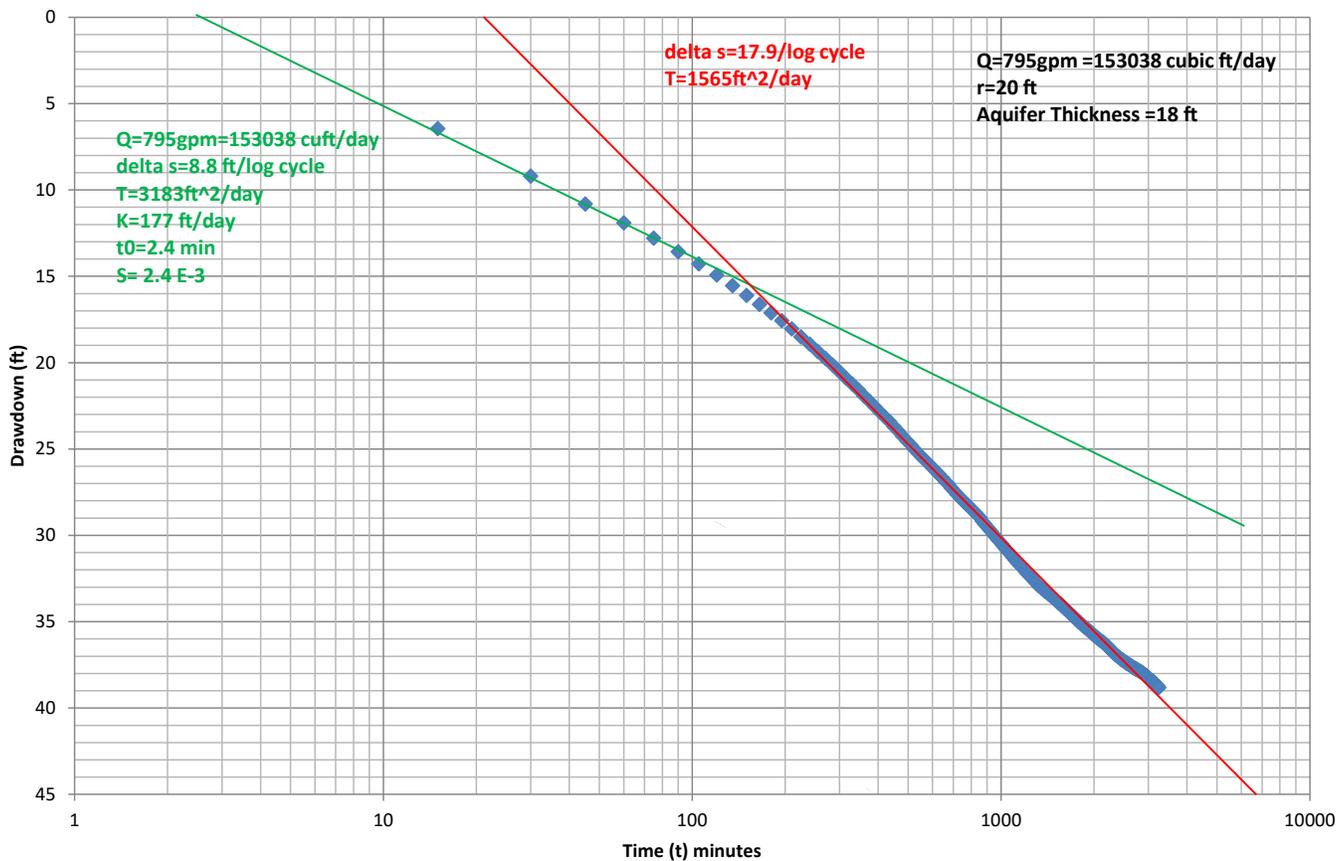


Figure C4-5. Results of Bolstad 1 aquifer test #4.

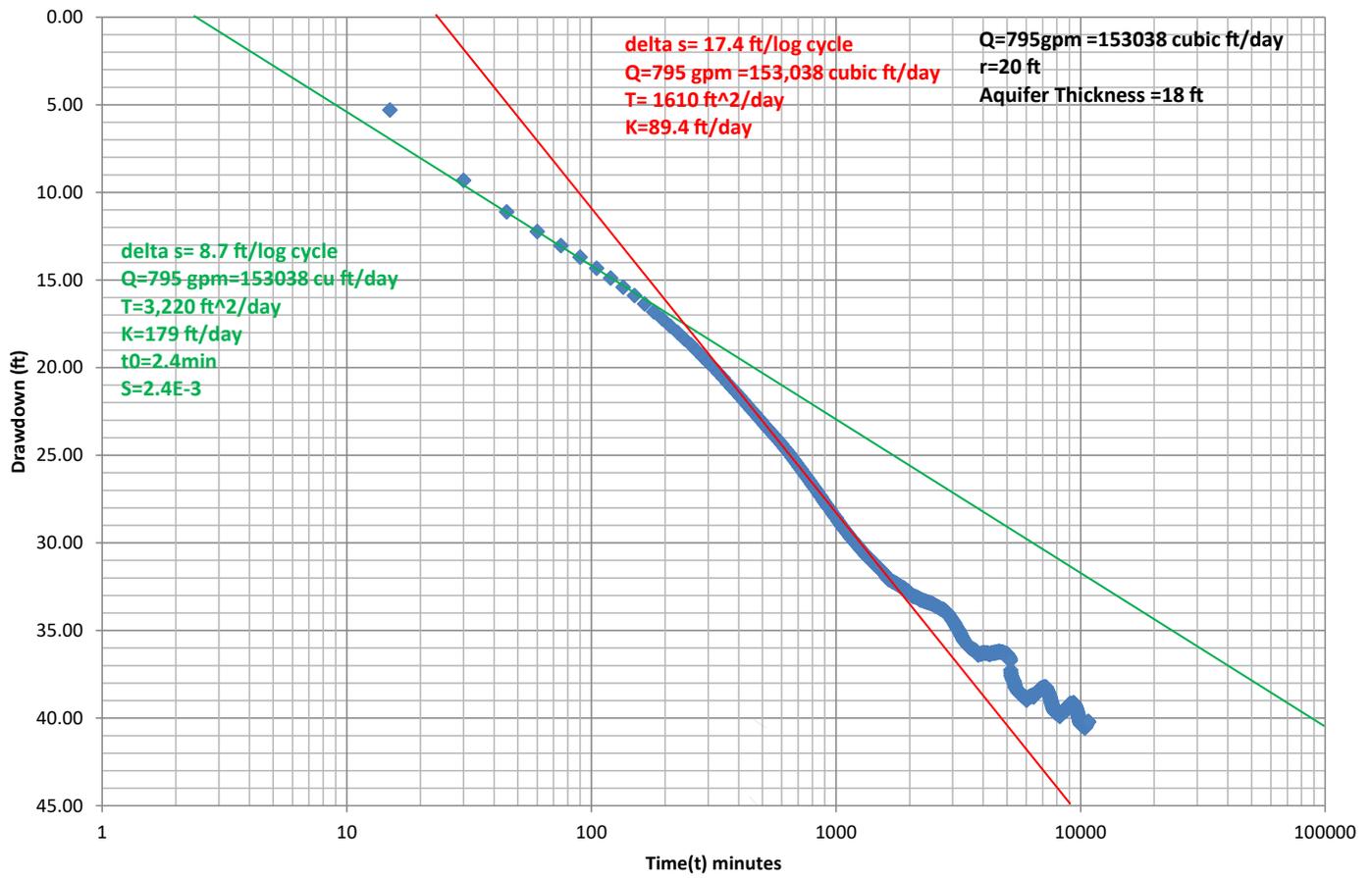


Figure C4-6. Results of Bolstad 1 aquifer test #5.